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# 1 Livestock and human use of land: productivity trends and 2 dietary choices as drivers of future land and carbon dynamics

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23  
24 **Abstract.** Land use change has been the primary driving force of human alteration of  
25 terrestrial ecosystems. With 80% of agricultural land dedicated to livestock  
26 production, the sector is an important lever to attenuate land requirements for food  
27 production and carbon emissions from land use change. In this study, we quantify  
28 impacts of changing human diets and livestock productivity on land dynamics and  
29 depletion of carbon stored in vegetation, litter and soils. Across all investigated  
30 productivity pathways, lower consumption of livestock products can substantially  
31 reduce deforestation (47-55%) and cumulative carbon losses (34-57%). On the supply  
32 side, already minor productivity growth in extensive livestock production systems  
33 leads to substantial CO<sub>2</sub> emission abatement, but the emission saving potential of  
34 productivity gains in intensive systems is limited, mainly due to trade-offs with soil  
35 carbon stocks. If also accounting for uncertainties related to future trade restrictions,  
36 crop yields and pasture productivity, the range of projected carbon savings from  
37 changing diets increases to 23-78%. Highest abatement of carbon emissions (63-  
38 78%) can be achieved if reduced consumption of animal-based products is combined  
39 with sustained investments into productivity increases in plant production. Our  
40 analysis emphasizes the importance to integrate demand- and supply-side oriented  
41 mitigation strategies and to combine efforts in the crop and livestock sector to enable  
42 synergies for climate protection.

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44  
45 **Keywords:** livestock productivity; diets; land use; deforestation; carbon emissions; greenhouse gas  
46 mitigation

50 **1. Introduction**

51 Land transformation has been the primary driving force of human alteration of terrestrial  
52 ecosystems, strongly interacting with most other aspects of global environmental change  
53 (Lambin et al., 2001; Steffen et al., 2015; Vitousek et al., 1997). Carbon emissions from land  
54 use and land-cover change contribute 12.5% to anthropogenic carbon emissions (Houghton et  
55 al., 2012), thus representing the second-largest source after fossil fuel combustion (van der  
56 Werf et al., 2009). In view of the serious danger that climate change poses to ecosystems and  
57 human welfare (Smith et al., 2009), the capacity of land to sequester carbon is one of its  
58 crucial functions. Besides the protection and restoration of forests, recent efforts to foster  
59 climate action like the “4 per 1000 initiative” under the framework of the Lima-Paris Action  
60 Agenda emphasise the importance of soil carbon which is also stored in agricultural  
61 ecosystems.

62 The livestock sector is a key element of land related human interference with the Earth  
63 system, consuming 58% of the economically used plant biomass (12.1 Pg/yr) in contrast to  
64 12% directly serving as food (Krausmann et al., 2008). Resulting overall land use of livestock  
65 production accounts for 80% of agricultural land (Steinfeld et al., 2006). Deforestation is the  
66 most critical aspect of land use change, with livestock playing a pivotal role through the  
67 establishment of new pastures or expansion of arable land to produce crops like soybeans in  
68 the wake of intensifying livestock feeding practices around the world (Herrero et al., 2009;  
69 Naylor et al., 2005). While cattle ranching is considered as the major proximate cause of  
70 forest clearing in the Legal Amazon, soy cultivation often expanded into areas previously  
71 used as pastures, thereby indirectly triggering forest-to-pasture conversion elsewhere (Barona  
72 et al., 2010). Moreover, soy production may have contributed to deforestation by other  
73 indirect pathways, such as boosting land prices and infrastructure development (Barona et al.,  
74 2010; Fearnside, 2001, 2005; Nepstad et al., 2006).

75 Accordingly, restraining land requirements is increasingly regarded as a decisive measure to  
76 alleviate detrimental impacts of livestock production on the environment (Smith et al., 2013;  
77 Steinfeld and Gerber, 2010; Wirsenius et al., 2010), either on the supply side by changes in  
78 livestock production systems or on the demand side by lower consumption of land-intensive  
79 livestock commodities. On the supply side, substantial differences in feed conversion  
80 efficiencies across regions and levels of intensification indicate a large potential to transform  
81 biomass flows within the global food system and attenuate pressures on natural resources  
82 (Bouwman et al., 2013; Havlík et al., 2014; Herrero et al., 2013b, 2015; Weindl et al., 2015;  
83 Wirsenius et al., 2010). Intensification of livestock production systems does not only  
84 considerably alter feed and overall resource use per animal product, but it also affects the  
85 composition of feed baskets, shifting the focus from residues, food waste and grazed biomass  
86 to higher quality and nutrient-rich feed. However, resulting increase in the importance of

87 cropland at the expense of pastures could impede carbon sequestration, since grasslands have  
88 a high root turnover and build up substantial soil organic carbon stocks (Conant et al., 2001;  
89 Don et al., 2011).

90 In consequence, understanding the link between livestock, land and carbon requires a detailed  
91 representation of feeding regimes and a comprehensive coverage of different land use types  
92 and related carbon pools. While several studies highlight the importance of feeding  
93 efficiencies and shifts in livestock production systems to attenuate pressures on land and to  
94 reduce greenhouse gas (GHG) emissions (Cohn et al., 2014; Havlík et al., 2014; Herrero et  
95 al., 2013b; Valin et al., 2013), they consider aggregated carbon dioxide (CO<sub>2</sub>) emissions  
96 without separating carbon pools and channels of land conversion or limit the scope to nitrous  
97 oxide (N<sub>2</sub>O) and methane (CH<sub>4</sub>) emissions. However, a dedicated coverage of soil carbon and  
98 non-forest land is essential for designing efficient climate protection schemes, since exclusion  
99 of non-forest carbon stocks from mitigation policies entails significant carbon leakage (Popp  
100 et al., 2014) and carbon stored in soils represents more than twice the amount found in the  
101 atmosphere (Smith, 2008).

102 This study aims at specifically addressing the impacts of future livestock production on the  
103 interplay between different managed and unmanaged land types and related trade-offs in  
104 terms of carbon losses from vegetation, litter and soils. Special attention is hereby given to  
105 sector-specific options to mitigate pressures on terrestrial ecosystems like changes in human  
106 diets and different livestock productivity pathways, either representing a catch-up of low  
107 productive systems to higher productivity levels, a stagnation of productivity in extensive  
108 systems or a moderate productivity reduction in intensive systems. For this aim, we apply a  
109 spatially explicit, global agro-economic model, where links between livestock, land and crop  
110 production are established through product-specific feed baskets that evolve with the  
111 productivity level, through manure provision, investments into research and development and  
112 trade flows.

## 113 **2. Methods and data**

### 114 *2.1. Modelling framework*

115 The Model of Agricultural Production and its Impact on the Environment (MAGPIE) is a  
116 global partial equilibrium agro-economic model that combines spatially explicit biophysical  
117 constraints with regional socioeconomic information for ten world regions (Table 1) to derive  
118 optimal resource allocation and agricultural production patterns (Bodirsky et al., 2014; Lotze-  
119 Campen et al., 2008; Popp et al., 2014, 2017; Stevanović et al., 2016). Possible future  
120 developments of the agricultural and land-use sectors are simulated in a recursive dynamic

121 mode by minimizing a nonlinear global objective function defining global agricultural  
122 production costs.

123

124

125 **Table 1.** Socio-economic regions in MAgPIE.

Acronyms	MAgPIE regions
AFR	Sub-Sahara Africa
CPA	Centrally Planned Asia (incl. China)
EUR	Europe (incl. Turkey)
FSU	Former Soviet Union
LAM	Latin America
MEA	Middle East and North Africa
NAM	North America
PAO	Pacific OECD (Australia, Japan and New Zealand)
PAS	Pacific Asia
SAS	South Asia (incl. India)

126

127

128 Pasture productivity, crop yields under both rainfed and irrigated conditions, related irrigation  
129 water demand per crop, water availability for irrigation and carbon densities are simulated by  
130 the process-based, dynamic global vegetation and water balance model LPJmL (Lund-  
131 Potsdam-Jena model with managed Land) (Bondeau et al., 2007; Müller and Robertson,  
132 2014) on 0.5 degree resolution and aggregated to 1000 clusters for this study (Dietrich et al.,  
133 2013). LPJmL simulates growth, production and phenology of 9 plant functional types (Sitch  
134 et al., 2003) and of 11 crop functional types as well as managed grassland (Bondeau et al.,  
135 2007). Water and carbon fluxes are directly connected to vegetation patterns and dynamics  
136 through the linkage of transpiration, photosynthesis and plant water stress.

137 Food demand projections are exogenously calculated based on an econometric regression  
138 model for national caloric intake per capita (Bodirsky et al., 2012, 2015; Valin et al., 2014).

139 Regional feed demand depends on livestock production quantities and regional system-  
140 specific feed baskets that evolve with livestock productivity trajectories. Global demand for  
141 agricultural commodities is allocated to the supply regions via trade dynamics based on an  
142 exogenous rate of trade liberalization, defining the proportion of agricultural goods that are,  
143 on top of historical trade patterns, traded according to comparative advantages (Schmitz et al.,  
144 2012). Through investments in research and development, the model can endogenously  
145 increase crop yields and pasture productivity, with the costs of technological change  
146 depending on the current technology level (Dietrich et al., 2014). More information on the  
147 model version underlying this study can be found in the SI appendix.

148

149 *2.2. Land use change*

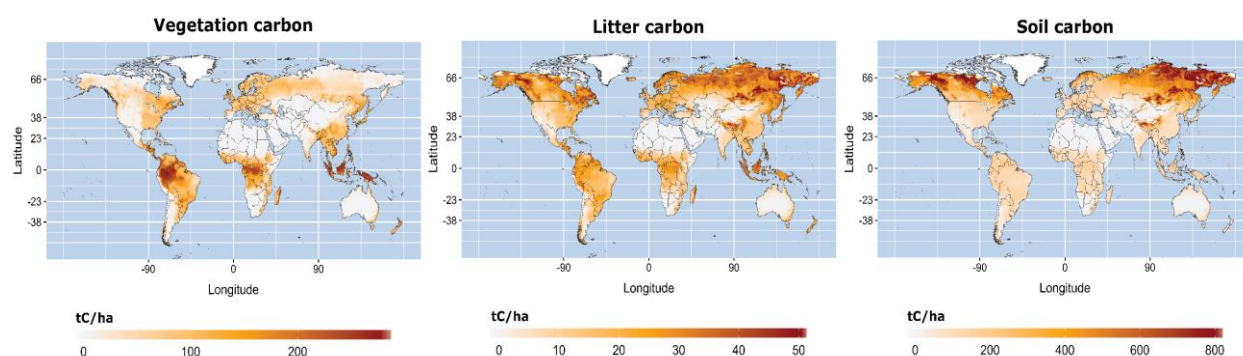
150 Competition for land is explicitly addressed for cropland, pasture, forest (including forestry),  
151 and other land (other natural vegetation such as savannahs and abandoned agricultural land).  
152 Forest areas designated for wood production (about 30% of the initial forest area) and pristine  
153 forests in protected areas (12.5% of global forests (FAO, 2010)) are excluded from  
154 conversion into agricultural land. The suitability of land for crop cultivation further constrains  
155 the conversion of natural vegetation or pastures to cropland and is primarily determined using  
156 crop yields from LPJmL. Additionally, cropping can only occur on land that is at least  
157 marginally suitable for rainfed crop production with regard to climate, topography and soil  
158 type according to the Global Agro-Ecological Assessment (GAEZ) methodology on land  
159 suitability (Fischer et al., 2002; Krause et al., 2013; van Velthuis et al., 2007). In response  
160 to production costs (SI appendix A.1.5) and biophysical constraints, MAgPIE optimizes  
161 spatial distribution of crops and pasture within current agricultural land as well as the balance  
162 between land expansion, trade, and improvements in land productivity.

163

164 *2.3. Carbon emissions*

165 Carbon emissions in MAgPIE are computed as the change in terrestrial carbon stocks from  
166 land conversion processes between simulated land types. Spatially explicit carbon stocks for  
167 all land types and carbon pools (vegetation, litter and soils) are calculated by multiplying  
168 pool- and land-specific carbon densities with land area. Negative carbon emissions occur  
169 when cropland is set-aside from agricultural production and subsequent ecological succession  
170 restores natural vegetation carbon stocks (Humpeöder et al., 2014), thus turning land into a  
171 sink for atmospheric carbon. In case of regrowth, vegetation carbon density increases over  
172 time along sigmoid growth curves which are based on a Chapman-Richards volume growth  
173 model (Murray and von Gadow, 1993; von Gadow and Hui, 2001) and parameterized using  
174 vegetation carbon density of natural vegetation. Carbon densities for vegetation, litter and soil  
175 carbon pools of natural vegetation (Fig. 1) are provided by LPJmL.

176



**Fig. 1.** Potential carbon densities for vegetation, litter and soil carbon pools in tC/ha calculated by LPJmL assuming that all terrestrial grid cells are covered with natural vegetation.

177

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#### 181 *2.4. Livestock sector*

182 Livestock products are supplied by five animal food systems (beef cattle, dairy cattle, pigs,  
 183 broilers and laying hens). Feed conversion (total feed per product in dry matter) and feed  
 184 baskets (demand for different feed types per product in dry matter) are derived by compiling  
 185 system-specific feed energy balances (Wirsenius, 2000; Wirsenius et al., 2010), using feed  
 186 energy requirements of all animals within the respective animal food system, i.e. reproducers,  
 187 producers and replacement animals as estimated by Wirsenius (2000). These estimates are  
 188 based on standardized bio-energetic equations and include energy requirements for  
 189 maintenance, growth, lactation, reproduction and other basic biological functions of the  
 190 animals. Non-linear regression models for feed conversion and feed composition (share of  
 191 different feed groups in feed baskets) with livestock productivity (annual production per  
 192 animal [ton/animal/year]) as predictor permit the construction of productivity dependent feed  
 193 baskets. Incorporation of spatial heterogeneity and climatic conditions into weighted non-  
 194 linear regression models for feed composition is facilitated by a proxy based on Koeppen-  
 195 Geiger climate zones (Portland State University, 2015). More information on the livestock  
 196 sector implementation can be found in the SI appendix A.2.

197

#### 198 *2.5. Scenario description*

199 Socio-economic drivers are parametrized in line with the “Middle of the Road” scenario of  
 200 the Shared Socioeconomic Pathways (SSPs) for climate change research (Kriegler et al.,  
 201 2017; O’Neill et al., 2014; Popp et al., 2017). In this scenario (SSP2), gross domestic product  
 202 and population trajectories reach global values of 230 trillion US Dollars (at 2005 prices and  
 203 adjusted for purchasing power parity) and 9.1 billion people in 2050 (IIASA, 2013). Global  
 204 trade barriers are relaxed by 5% per decade.

205

206 **Table 2.** Overview of scenario setting.

Scenario		Description
Dietary choices	SSP2	Food demand trajectories according to the SSP2 narrative with an average per capita food demand of 3051 kcal per day and 24% animal-based products in dietary calories in 2050
	DEMI	Gradual change towards a demitarian Western diet with a share of animal-based products in dietary calories of no more than 15% in 2050
Livestock productivity	BASELINE	Livestock productivity trajectories according to the SSP2 narrative with medium pace in productivity increases and a slight catch-up of low productive systems
	DIVERGENCE	Continuation of historically observed very divergent productivity trends with little improvements in low productive systems
	CATCH-UP	SSP2 + further closure of the productivity gap by 45% for ruminant systems and by 60% for monogastric systems until 2050
	MODERATION	SSP2 + productivity reductions in highly productive systems to the level of 75% relative to the productivity frontier

207

208

209 We investigate eight scenarios as combinations of two variants of future *dietary choices* and  
 210 four variants of *livestock productivity* trends (Table 2). Supplementing the baseline diet  
 211 scenario (SSP2), we define an alternative development of dietary patterns (SI appendix, Fig.  
 212 S7), representing a gradual change of SSP2 diet projections to lower shares of animal-based  
 213 calories in diets, with 15% as upper limit in 2050 for calories from livestock and fish (DEMI).  
 214 With the share of animal-based calories being approximately half the currently observed level  
 215 in OECD countries, the DEMI scenario builds upon the concept of a “demitarian” Western  
 216 diet (Bodirsky et al., 2014; Stevanović et al., 2017; Sutton and Ayyappan, 2013).

217 The diet scenarios are combined with four alternative livestock productivity pathways (SI  
 218 appendix, Fig. S8). Besides exploring impacts of productivity gains, we also explore how de-  
 219 intensification strategies could affect land and carbon dynamics. The BASELINE scenario,  
 220 following the SSP2 narrative, is characterized by a medium pace in productivity  
 221 improvements, where low-productive regions catch up to a certain extent (Popp et al., 2017).  
 222 With little improvements in some regions’ low productive systems, the DIVERGENCE  
 223 scenario represents the continuation of historically observed very divergent productivity  
 224 developments and is constructed by following the extrapolation of historical trends between  
 225 1970 and 2010, if they are lower than SSP2 projections. The ambitious CATCH-UP scenario  
 226 assumes a further closure of the productivity gap, defined by top-performing countries in  
 227 2010, by 45% for ruminant systems and by 60% for monogastric systems until 2050. The  
 228 MODERATION scenario explores a variation of SSP2 livestock productivity trends at the  
 229 opposite end of the range, the highly intensive systems. Until 2050, these systems are  
 230 assumed to experience a reduction in livestock productivity to the level of 75% relative to the  
 231 productivity frontier defined by top-performing countries in 2010.



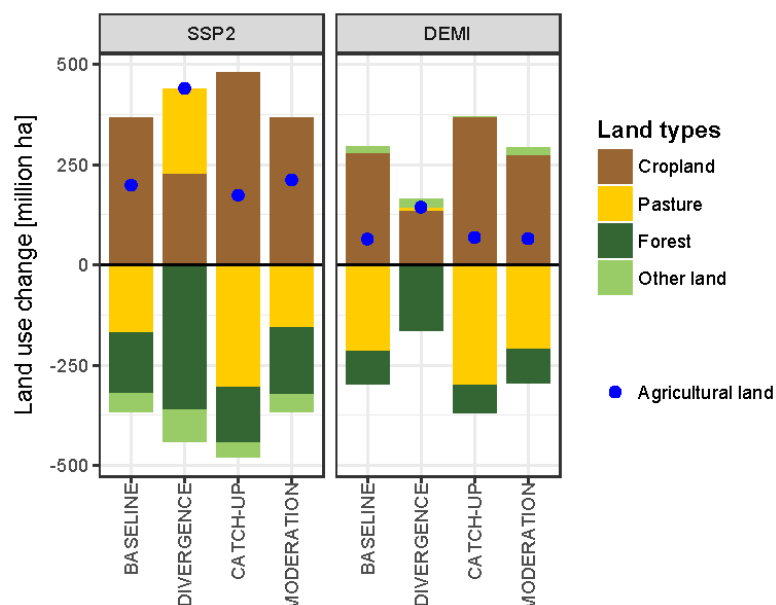
232 To understand the role of trade and land productivity for land use change and related  
233 emissions, we conduct a sensitivity analysis applying three additional scenario settings: a)  
234 ***Restricted trade*** (relative to the default SSP2 setting) where we assume that interregional  
235 trade patterns, in terms of self-sufficiency ratios and relative shares of regional trade flows,  
236 are constant over time; b) ***Liberalized trade*** where global trade barriers are relaxed by 10%  
237 per decade (instead of 5% as in the SSP2 default setting), which is close to observed  
238 liberalization trends of the last decade; and c) ***Exogenous yield*** where all standard  
239 productivity and diet scenarios are calculated with exogenous trajectories of crop yields and  
240 pasture productivity, based on the endogenously calculated crop and pasture productivity  
241 trends from the SSP2 BASELINE simulation in the default model setting.

### 242 3. Results

#### 243 3.1. Land dynamics

244 The potential of the livestock sector to substantially alter land use dynamics is clearly visible  
245 on the global scale (Fig. 2). The interaction between cropland and pasture dynamics plays an  
246 important role for deforestation and is strongly influenced by livestock productivity  
247 trajectories, but also subject to demand-side preferences. In the SSP2 BASELINE scenario,  
248 total agricultural land increases from 4630 Mha in 2010 to 4830 Mha in 2050 as a result of  
249 substantial cropland expansion (+370 Mha, +26%) that is partly compensated by a reduction  
250 in pasture area (-170 Mha, -5%). By 2050, forest losses amount to 150 Mha, while conversion  
251 of other natural vegetation represents a minor contribution to land use change (50 Mha).  
252 Across all diet and productivity scenarios, projected deforestation ranges between 70 and 360  
253 Mha. Dietary changes towards less livestock products reduce pressures on land, translating  
254 into lower cropland expansion (23-39% less than under SSP2 diets) and avoided deforestation  
255 (47-55%).

256



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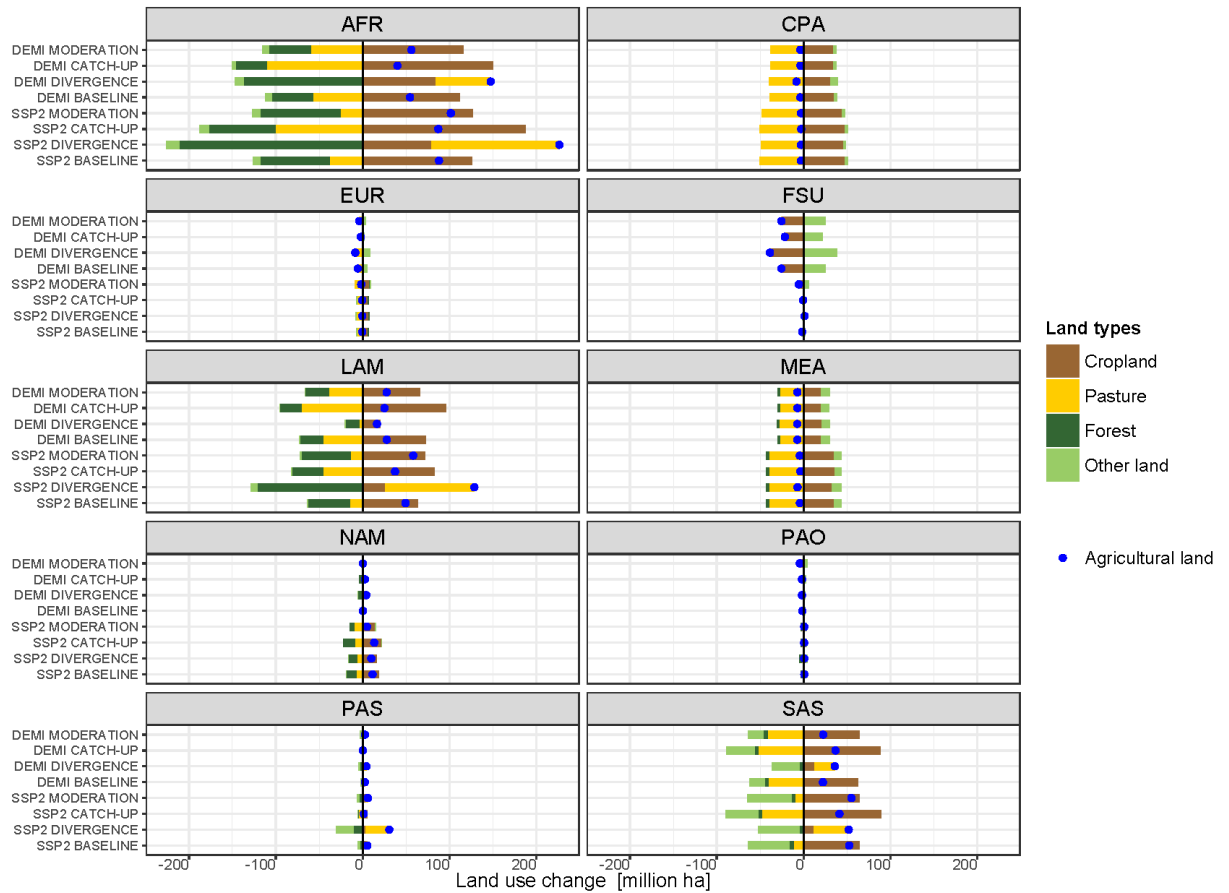
258 **Fig. 2.** Changes in global cropland, pasture, forest and other natural vegetation between 2010 and 2050 in Mha.  
 259 Blue points indicate the net change in global agricultural land.

260

261 All scenarios involve expansion of cropland (10-35%) which increases with higher livestock  
 262 productivity and decreases with lower consumption of livestock products. Implications for  
 263 deforestation depend on the potential of pasture-to-cropland conversion to counterbalance  
 264 increased land demand to grow crops. Reductions in pasture area in the wake of higher  
 265 livestock productivity outpace related increases in cropland, thus entailing a land sparing  
 266 effect. Only under stagnating low livestock productivity in some regions together with a  
 267 growing demand for livestock products (SSP2 DIVERGENCE), we observe an increase in  
 268 pasture area (+210 Mha) and consequently the highest estimate for deforestation. The  
 269 MODERATION productivity scenarios entail very similar dynamics as the BASELINE  
 270 scenarios, with slightly higher deforestation for SSP2 diets.

271 Global patterns of land use change are a congeries of diverse regional developments (Fig. 3).  
 272 In Latin America, Sub-Saharan Africa and South Asia, land conversion processes across  
 273 scenarios are strongly influenced by livestock productivity trends, resulting in a large regional  
 274 spread of deforestation and loss of other natural ecosystems. In Centrally Planned Asia,  
 275 Former Soviet Union, North America, Middle East and North Africa, land dynamics primarily  
 276 react to dietary changes, ending forest conversion in North America and resulting in land  
 277 abandonment and regrowth of natural vegetation in the Former Soviet Union. In the Middle  
 278 East and North Africa, expansion of agricultural activities is heavily constrained by the  
 279 scarcity of natural resources, with pasture being the only land resource available for cropland  
 280 expansion. Establishment of new pastures, discernibly linked to loss of forests or other natural

281 vegetation, is only simulated under the DIVERGENCE pathway with prevailing low  
 282 productivity of regional livestock production. Regional patterns highlight the important role  
 283 of developments in Sub-Saharan Africa and Latin America for further alteration of terrestrial  
 284 ecosystems.  
 285



286  
 287 **Fig. 3.** Changes in regional cropland, pasture, forest and other natural vegetation between 2010 and 2050 in Mha.  
 288 Blue points indicate changes in regional agricultural land defined as the sum of cropland and pasture. Results are  
 289 depicted for the ten socio-economic regions in MAgPIE: AFR (Sub-Saharan Africa), CPA (Centrally Planned  
 290 Asia), FSU (Former Soviet Union), EUR (Europe, including Turkey), LAM (Latin America), MEA (Middle East–  
 291 North Africa), NAM (North America), PAO (Pacific OECD), PAS (Pacific Asia), and SAS (South Asia).

292

### 293 3.2. Carbon dynamics

294 Agricultural expansion and losses of natural ecosystems across all scenarios drive further  
 295 depletion of terrestrial carbon stocks, but by different orders of magnitude (Fig. 4). Until  
 296 2050, cumulative carbon releases amount to 20-80 Gt C, which is equivalent to 74-295 Gt  
 297 CO<sub>2</sub> emitted to the atmosphere (Table 3). As in the case of deforestation, the predominant role  
 298 of Sub-Saharan Africa and Latin America is clearly visible in our results, contributing 74-  
 299 93% to global carbon losses. If low historical productivity improvements are assumed to

300 continue, both regions together are projected to double (DEMI) or triple (SSP2) their carbon  
 301 emissions from land use change compared to BASELINE trends. Thus, already intermediate  
 302 livestock productivity improvements, as assumed under the BASELINE pathways for these  
 303 regions, lead to substantial abatement of carbon emissions. The role of different land types  
 304 within overall land dynamics affects the extent at which the different above and belowground  
 305 carbon pools contribute to net carbon losses, both at the regional and global scale. In the SSP2  
 306 BASELINE scenario, changes in vegetation carbon account for 51%, depletion of soil carbon  
 307 for 39% and losses of carbon in litter for 10% of total releases (124 Gt C).

308

309

310 **Table 3.** Cumulative CO<sub>2</sub> emissions between 2010 and 2050 for all scenarios in Gt CO<sub>2</sub>.

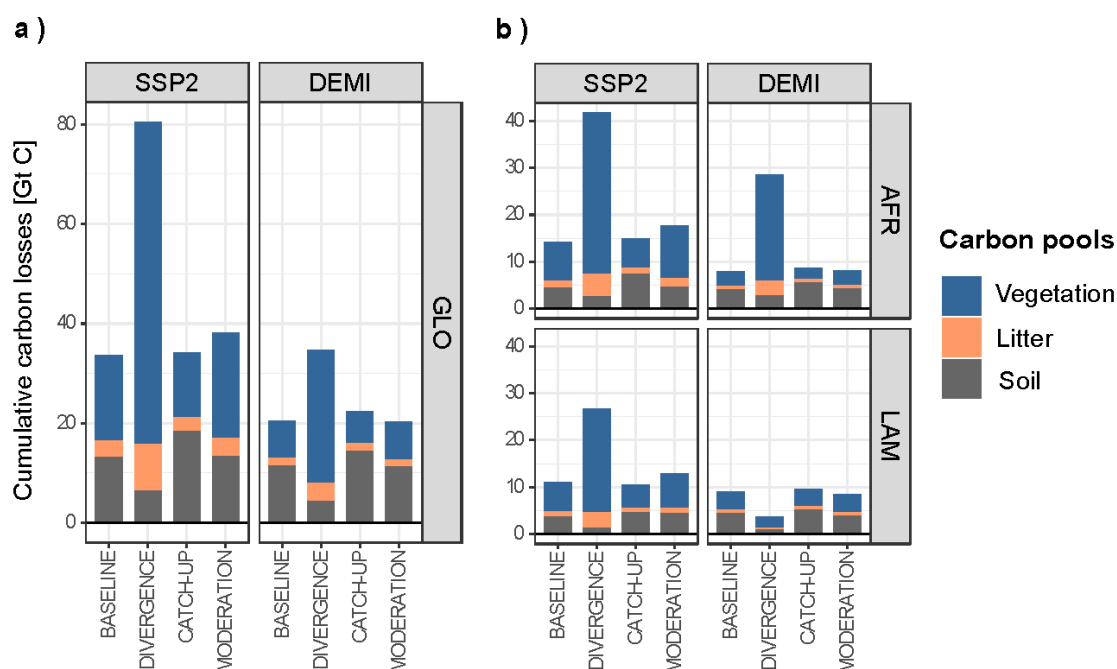
Diets	Productivity	Vegetation	Litter	Soil	All pools
SSP2	BASELINE	63	12	49	124
	DIVERGENCE	236	34	24	295
	CATCH-UP	47	10	68	125
	MODERATION	76	14	49	140
DEMI	BASELINE	27	5	43	75
	DIVERGENCE	97	13	17	127
	CATCH-UP	23	5	54	82
	MODERATION	27	5	42	74

311

312

313 CATCH-UP pathways entail very similar cumulative carbon losses compared to BASELINE  
 314 productivity trends, with a higher contribution of soil carbon and a lower share of vegetation  
 315 carbon. Although deforestation is slightly lower, considerable pasture-to-cropland conversion  
 316 depletes carbon stored in soils and counteracts minor potential carbon savings from avoided  
 317 deforestation. High deforestation, as triggered by the DIVERGENCE pathway in combination  
 318 with SSP2 diets, results in high carbon emissions. However, total carbon releases and  
 319 especially soil carbon losses would be even larger if we only considered loss of forest carbon  
 320 stocks, as expanding pastures can also sequester significant amounts of carbon in soils. While  
 321 in the SSP2 MODERATION scenario, deforestation and resulting carbon emissions are  
 322 higher than in the BASELINE, no difference can be observed for reduced consumption of  
 323 livestock products. In the DEMI scenarios, expansion of cropland is in general less linked to  
 324 deforestation and relies stronger on conversion of pastures, resulting in a higher contribution  
 325 of soil carbon to total carbon releases. Across all productivity pathways, dietary changes  
 326 towards less livestock products can substantially reduce cumulative carbon losses (34-57%).

327



328  
 329 **Fig. 4.** Cumulative carbon losses between 2010 and 2050 in Gt C from vegetation, litter and soil carbon pools. The  
 330 left panel (a) illustrates global values and the right panel (b) shows values for Sub-Saharan Africa (AFR) and Latin  
 331 America (LAM).

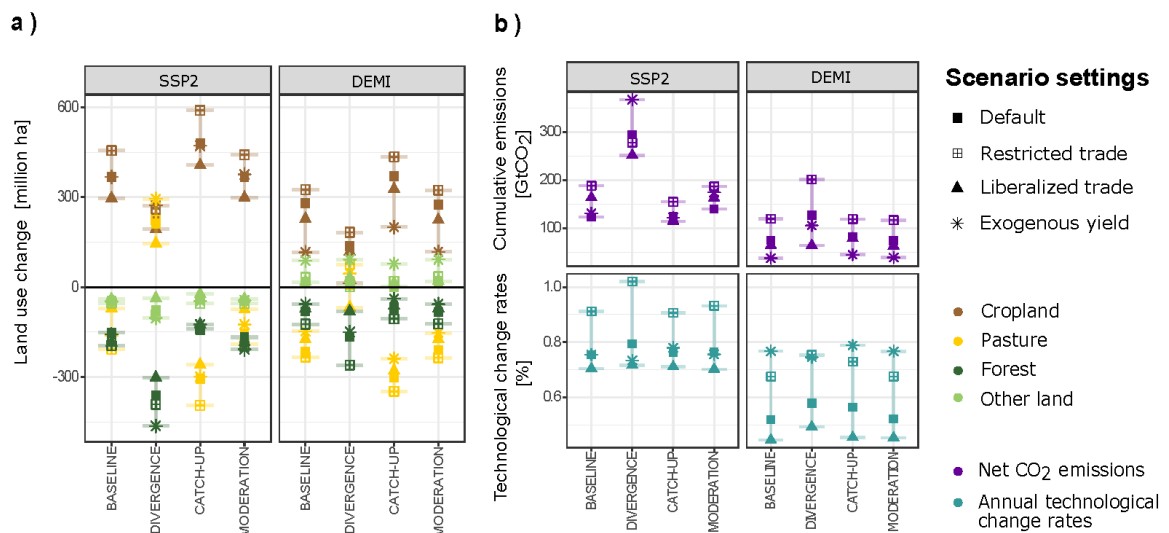
332

### 333 3.3. Uncertainties in projected land and carbon dynamics

334 How demand- and supply-side scenarios alter land and carbon dynamics also depends on the  
 335 role of intermediate processes such as reallocation of production through international trade  
 336 and efforts to invest into yield improvements and pasture management (Fig. 5, Table 4). A  
 337 *restricted trade* regime with self-sufficiency ratios and relative export flows fixed to 1995  
 338 levels constrains the possibility to balance heterogeneous demand trajectories and differences  
 339 in land availability and productivity across regions through interregional reallocation of  
 340 production. As a result, we observe more cropland expansion, deforestation and CO<sub>2</sub>  
 341 emissions, although limited options to conciliate increasing food demand and available  
 342 resources in some regions simultaneously lead to higher investments into yield increasing  
 343 technological change. Due to the low flexibility in the system, the potential of dietary changes  
 344 to attenuate land use change and related emissions (23-37% reduction in emitted CO<sub>2</sub>) is low  
 345 compared to other sensitivity settings.

346 In a *liberalized trade* setting, trade patterns endogenously respond to asymmetric regional  
 347 developments and can compensate regional inefficiencies and imbalances between food  
 348 demand and availability of natural resources. Production is allocated according to  
 349 comparative advantages between regions, which could also favour locations where land is  
 350 abundant and lead to lower incentives to invest into yield increases. Thus, impacts of trade

351 liberalization on land and carbon dynamics are ambiguous and depend on overall  
 352 development pathways of agriculture. In the case of the SSP2 BASELINE and  
 353 MODERATION scenarios, trade liberalization entails higher forest losses and CO<sub>2</sub> emissions,  
 354 while production costs and investments into research and development are lower. In the case  
 355 of diverging livestock productivity trends, however, a reallocation of trade flows and  
 356 production can exploit the large heterogeneity of regional livestock productivities and feed  
 357 efficiencies, resulting in avoided deforestation and mitigation of CO<sub>2</sub> emissions.  
 358 Comparison of scenarios assuming *exogenous yield trajectories* with default simulations  
 359 highlights the buffering effect of yield increasing innovation and management. Efforts to  
 360 invest into land productivity depend on land scarcity and are driven by demand- and supply-  
 361 side pressures on the agricultural system. Scenarios with exogenous yield trajectories exclude  
 362 this dampening effect, thus leading to stronger signals of changes in productivity pathways  
 363 and dietary choices. Assuming persistent efforts to increase land productivity independent  
 364 from demand trajectories, the land sparing effect of a reduced consumption of livestock  
 365 products is more pronounced, with a decline in deforestation by 64-72% and emissions  
 366 abatement by 63-78%.  
 367



368  
 369 **Fig. 5.** Sensitivity analysis exploring the influence of international trade and yield trajectories on land use change  
 370 and related emissions between 2010 and 2050. Panel a) illustrates changes in regional cropland, pasture, forest and  
 371 other natural vegetation in Mha. Panel b) shows cumulative CO<sub>2</sub> emissions from changes in vegetation, litter and  
 372 soil carbon stocks in Gt CO<sub>2</sub> and average annual technological change rates (%) between 2010 and 2050.

373

374 **Table 4.** Impacts of dietary changes on deforestation and cumulative CO<sub>2</sub> emissions between 2010 and 2050 for  
 375 all productivity scenarios in the default and additional model settings of the sensitivity analysis (changes in CO<sub>2</sub>  
 376 emissions (%)) for DEMI diet scenarios relative to SSP2 diet scenarios).

		BASELINE	DIVERGENCE	CATCH-UP	MODERATION
Deforestation	<i>Default</i>	-47%	-55%	-49%	-50%
	<i>Restricted trade</i>	-37%	-34%	-25%	-38%
	<i>Liberalized trade</i>	-61%	-73%	-50%	-62%
	<i>Exogenous yield</i>	-64%	-68%	-69%	-72%
CO <sub>2</sub> emissions	<i>Default</i>	-39%	-57%	-34%	-47%
	<i>Restricted trade</i>	-36%	-28%	-23%	-37%
	<i>Liberalized trade</i>	-61%	-74%	-31%	-62%
	<i>Exogenous yield</i>	-71%	-71%	-63%	-78%

377

378

#### 379 **4. Discussion**

380 Recent work suggests that emission and land saving potentials of livestock system  
 381 intensification by far outpace contributions from the crop sector (Cohn et al., 2014; Havlík et  
 382 al., 2013, 2014; Valin et al., 2013). Moreover, there is evidence that shifts in dietary patterns  
 383 have a similar potential to abate GHG emissions as an agricultural GHG tax policy, but  
 384 without possible negative effects on food prices (Stevanović et al., 2017). Building upon these  
 385 insights, we further disentangle impacts of livestock productivity growth and dietary changes  
 386 on land and carbon dynamics, focusing on the interplay between different land types and  
 387 related trade-offs in terms of carbon losses from vegetation, litter and soils.

388

##### 389 *4.1. Data uncertainty and assumptions in livestock modelling*

390 Estimation of feed baskets for different animal food systems is based on feed energy balances,  
 391 following the approach of Wirsenius et al. (2010), where we use the net energy (NE) system  
 392 for cattle and the metabolizable energy (ME) system for monogastrics. Hereby, the assumed  
 393 homogenous energy densities within the different feed commodities are a potential source of  
 394 inaccuracy, especially in the case of grasses and forage crops that are characterized by large  
 395 variations in nutrient content. Digestible energy (DE) of harvested grass does not only depend  
 396 on species composition and climate, but also on management factors such as fertilizer input  
 397 and stage of maturity (Stergiadis et al., 2015). Moreover, the non-linear empirical equations  
 398 used to calculate NE contents from DE contents of different feed commodities are based on  
 399 experimental diets with relatively high digestibility (NRC, 1996; Wirsenius, 2000), which  
 400 introduces an additional potential source of inaccuracy if applied to extensive cattle systems  
 401 relying on native pastures and crop residues.

402 Establishment of feed energy balances also necessitates information on feed energy supply.  
403 While for food crops and food industry byproducts, feed use of the whole livestock sector at  
404 country-scale is derived from FAOSTAT (2013), we use estimates from the literature to  
405 determine feed use of crop residues and food waste (Bodirsky et al., 2012; Eggleston et al.,  
406 2006; Lal, 2005; Wirsenius, 2000). Grazed biomass is calculated as the residuum required for  
407 balancing ruminant feed demand with supply and therefore the most uncertain component  
408 within feed rations. In addition, quantification of pasture area is prone to uncertainty, as  
409 conceptual assumptions of classifying pastures and rangelands are diverse (Erb et al., 2007;  
410 Ramankutty et al., 2008; Sayre et al., 2017). Multiplying pasture area with pasture  
411 productivity from LPJmL at high spatial resolution, we obtain potential supply of grass  
412 biomass which in most regions exceeds demand.

413 At initialization of the simulation period, potential supply of grass biomass is calibrated to  
414 regional grass demand, where the calibration factors reflect regional grazing intensities. While  
415 we assume that pasture management proportionally improves with productivity gains in the  
416 crop sector leading to higher grazing intensities, we do not account for related changes in  
417 nutritional quality of grasses and for the complex responses of net primary productivity and  
418 soil carbon sequestration to management and grazing at different stocking rates (Asner et al.,  
419 2004; Lambin et al., 2001; Rolinski et al., 2017). In South Asia, the high calibration factor  
420 points to a mismatch between feed energy requirements and availability of considered feed  
421 categories, similar to observations by Singh (1997) and Wirsenius (2000). In this region, a  
422 heterogeneous array of not considered feed source such as browse, biomass from weeding and  
423 grazing in forests and other non-agricultural land is likely to substantially contribute to feed  
424 availability (Wirsenius, 2000).

425 For projecting future feed demand under different productivity pathways, we derive  
426 relationships between feed baskets and livestock productivity. Results indicate that  
427 productivity gains lead to better feed efficiencies together with a shift from low-cost and low-  
428 energy feed, sourced from pastures or available as by-products from the agricultural supply  
429 chain, to cropland feed with higher nutrient densities, similar to findings obtained by Herrero  
430 et al. (2013b) and Wirsenius (2000). However, larger dispersion in the model for feed  
431 composition compared to feed conversion highlights the need for a more in-depth analysis of  
432 other potentially relevant factors such as availability of pasture compared to cropland feed,  
433 agro-ecological and climatic conditions that favour selected feed items, innovation and  
434 management that improve the quality of non-cropland or by-product feed components, and  
435 socio-economic as well as cultural determinants.

436



437 *4.2. Implications of productivity improvements*

438 Our model simulations indicate that livestock productivity gains drive cropland expansion,  
439 whose consequences regarding deforestation depend on the relative reduction in pasture and  
440 the suitability of these areas for cropping. Assuming no rebound effect of productivity  
441 increases on food demand, already minor productivity gains in extensive livestock systems  
442 are identified as an effective lever to avoid deforestation (50-58% reduction in BASELINE  
443 scenarios compared to DIVERGENCE pathways) and abate carbon emissions (41-58%  
444 reduction), since decreases in pasture area occur faster than expansion of cropland. Trade-offs  
445 with soil carbon losses equivalent of 25 Gt CO<sub>2</sub> are more than compensated by substantially  
446 lower emissions from vegetation carbon stored in native forests.

447 However, if further proceeding to high productivity levels, trade-offs with ecosystem services  
448 on managed land are more pronounced since large-scale pasture-to-cropland conversion  
449 impair carbon sequestration in agricultural soils and biodiversity (Alkemade et al., 2013). Our  
450 simulations reveal that strong increases in livestock productivity involve substantial depletion  
451 of soil carbon stocks, which can lead to a net increase of carbon emissions, although feed  
452 demand and deforestation are slightly lower under CATCH-UP pathways compared to  
453 BASELINE scenarios. Thus, a metric assessing the sustainability of livestock production that  
454 is solely oriented on feed or resource use efficiency may reach its limits in the case of  
455 significant conversion of pastures to cropland triggered by high livestock productivity gains.

456

457 *4.3. Context, barriers and potential limits of livestock system intensification*

458 A transition towards more intensive livestock systems is already occurring in many parts of  
459 the world in the wake of socio-economic changes such as increased population densities and  
460 rangeland fragmentation (Herrero et al., 2009; Hobbs et al., 2008). Yet, there is concern that  
461 especially poor and vulnerable farmers and pastoralists might not be able to keep pace with  
462 these rapid change processes and the combined challenge of satisfying the growing demand  
463 for food without using more resources and impeding climate mitigation in the agricultural  
464 sector (Garnett et al., 2013; Herrero et al., 2009). While in developing countries large farms  
465 might achieve innovation and technological change relying on the private sector, institutional  
466 mechanisms and public investments need to be designed to especially reach smallholder  
467 farmers and herders, providing access to formal and informal markets, extension services and  
468 innovation platforms as well as fostering efficient value chain and infrastructure development  
469 (Herrero et al., 2013a; McDermott et al., 2010).

470 Moreover, livestock production does not only take place under changing socio-economic  
471 conditions, but also in the context of a changing climate that is likely to negatively affect  
472 yields of major feed crops such as maize as well as rangeland productivity (Müller and

473 Robertson, 2014; Nelson et al., 2010; Weindl et al., 2015). Shifts in livestock production  
474 systems are an important lever to alter biomass flows and counterbalance detrimental impacts  
475 of climate change on the natural resource base. A transition towards more efficient mixed  
476 crop-livestock system can simultaneously reduce adaptation costs and deforestation, thereby  
477 responding to both mitigation and adaptation imperatives (Weindl et al., 2015).

478 While increasing productivity of extensive systems in developing regions is perceived as  
479 beneficial both with regard to environmental and social considerations (Herrero et al., 2009;  
480 Steinfeld et al., 2006), there is an increasing concern about the downsides of large intensive  
481 operations (Franzluebbers et al., 2014; Lemaire et al., 2014). Pollution of terrestrial and  
482 aquatic ecosystems through excessive nitrogen and pesticides, decreasing soil fertility,  
483 conflicts with animal welfare, breeding of antibiotic-resistant pathogens, and the exploitation  
484 of non-renewable resources question the long-term sustainability of modern livestock  
485 industries (Carvalho et al., 2010; Franzluebbers, 2007; Herrero et al., 2010; Russelle et al.,  
486 2007). Analysing land and carbon effects of a moderate de-intensification of highly  
487 productive systems, we observe only small and ambiguous impacts on the system, starting  
488 with a slight growth in total feed demand and minor reduction in cropland feed, which  
489 translate into a small increase in deforestation and carbon emissions in the case of SSP2 diets  
490 and into almost identical land and carbon outcomes (compared to BASELINE) in the case of  
491 DEMI diet trajectories. Thus, potentially beneficial effects of moderate productivity decreases  
492 in intensive livestock systems on pollution and other aspects of the broader sustainability  
493 context are not jeopardized by impacts on land use and carbon losses, especially under  
494 reduced consumption of livestock products.

#### 495 *4.4. Implications of food consumption patterns*

496 Positive effects of changing diets for climate protection are well documented (Aiking et al.,  
497 2006; Bajželj et al., 2014; Popp et al., 2010; Stehfest et al., 2009; Stevanović et al., 2017).  
498 While supply-side climate policies have repercussions on food prices in developing regions  
499 (Havlík et al., 2014; Stevanović et al., 2017), demand-side oriented strategies aim at a  
500 reduction of animal-based food in affluent societies. Besides synergies in the area of public  
501 health, a shift in consumption patterns has various co-benefits, like ecosystem recovery  
502 through abandonment of land and mitigation of nitrogen pollution (Bodirsky et al., 2014;  
503 Springmann et al., 2016; Stehfest et al., 2009). Our estimates of the annual carbon mitigation  
504 potential until 2050 are in the range of 1.1-4.2 Gt CO<sub>2</sub>/yr for our default model setting, which  
505 is lower than 5.6 Gt CO<sub>2</sub>eq/yr and 5.9 Gt CO<sub>2</sub>eq/yr suggested by Stevanović et al. (2017) and  
506 Bajželj et al. (2014). While both studies use trajectories of dietary changes comparable to our  
507 DEMI diet scenario, they additionally assume a 50% food waste reduction and also account  
508 for non-CO<sub>2</sub> emissions which are projected to represent the major contribution of agricultural

509 emissions over the 21st century. The spread of our estimates, which amounts to 0.9-6.5 Gt  
510 CO<sub>2</sub>/yr if including results of the sensitivity analysis, indicates a strong dependence of climate  
511 benefits of changing consumer preferences on future productivity pathways in the livestock  
512 and crop sector, as well as on trade regulations.

513

#### 514 *4.5. Economic processes and complex interactions in the agricultural system*

515 Our results show that the theoretical potential of flexible trade flows to exploit regional  
516 differences in feed conversion efficiency through interregional reallocation of production only  
517 unfolds in scenarios that assume prevailing large geographical disparities of livestock  
518 production. Comparative advantages of some regions characterised by high resource  
519 availability can dampen efforts to invest into land productivity, with detrimental  
520 consequences for deforestation and carbon emissions, similar to dynamics attested by Schmitz  
521 et al. (2012). However, Havlík et al. (2014) suggest that intra- and interregional relocation of  
522 livestock production could contribute 49% of total emission abatement if incentivized by a  
523 global carbon price. In case that relative trade flows are fixed to 1995 levels, the inflexibility  
524 of the system generally leads to higher carbon emissions and constrains the potential of  
525 dietary changes to attenuate CO<sub>2</sub> emissions in our scenarios.

526 In our study, highest carbon savings from changing diets (63-78%) can be achieved if relaxed  
527 pressures on land have no negative repercussions on pasture management and productivity  
528 growth in the crop sector, emphasizing the importance to combine efforts in the crop and  
529 livestock sector to enable synergies for climate protection, in line with findings obtained by  
530 Valin et al. (2013). Moreover, our two-dimensional scenario matrix reveals that the spread of  
531 cumulative carbon emissions (between 2010 and 2050) associated with the explored  
532 productivity pathways is high for SSP2 diets (125-295 Gt CO<sub>2</sub>), while dietary changes  
533 towards less livestock products smooth differences (74-127 Gt CO<sub>2</sub>). Thus, a reorientation of  
534 consumer preferences would allow for a larger option space to develop regional livestock  
535 systems, progressing from a “land and carbon-only” approach to a broader sustainability  
536 metric that also considers animal well-being, livelihoods, water resources, biodiversity and  
537 pollution through various organic and inorganic substances.

## 538 **5. Conclusion**

539 If the growing demand for livestock products in developing countries is to be met without  
540 improvements in historically observed low livestock productivities in some regions,  
541 substantial increases in feed demand would imply massive forest and carbon losses. However,  
542 already intermediate livestock productivity gains can halt the expansion of pastures into  
543 pristine ecosystems and substantially reduce net land requirements for agricultural production,

544 with significant benefits for climate change mitigation. In contrast, ambitious productivity  
545 increases that still slightly improve feed and land use efficiency involve trade-offs with  
546 carbon sequestration in agricultural soils, thereby possibly increasing net carbon emissions.  
547 At the same time, moderate de-intensification of highly intensive systems has negligible  
548 impacts on land and carbon losses, thus not jeopardizing potentially beneficial effects on  
549 pollution, animal welfare and other aspects of the broader sustainability context.  
550 On the demand side, reducing the consumption of livestock products to 15% animal-based  
551 calories in diets until 2050 can significantly abate carbon emissions from land use change by  
552 up to 78%. However, the carbon saving potential of changing diets depends not only on  
553 livestock productivity pathways, but also on productivity trends in the crop sector, pasture  
554 management and on other boundary conditions of agricultural production such as trade  
555 regimes. Thus, preference-based strategies aiming at behavioural change have to go hand in  
556 hand with supply-side oriented schemes to increase the resource efficiency of livestock  
557 production as well as with dedicated forest and climate protection policies, which counteract  
558 resource inefficiencies in global trade patterns, prevent interregional leakage and incentivize  
559 efforts to invest in the sustainable intensification of the whole agricultural and food system.

560

561

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570

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