

Prospective Environmental Impact Assessment (*premise*): A streamlined approach to producing databases for prospective life cycle assessment using integrated assessment models

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ARTICLE INFO

Keywords:

Prospective scenario
Integrated assessment model
Emerging technology
Life cycle assessment

ABSTRACT

Prospective Life Cycle Assessment (pLCA) is useful to evaluate the environmental performance of current and emerging technologies in the future. Yet, as energy systems and industries are rapidly shifting towards cleaner means of production, pLCA requires an inventory database that encapsulates the expected changes in technologies and the environment at a given point in time, following specific socio-techno-economic pathways. To this end, this study introduces *premise*, a tool to streamline the generation of prospective inventory databases for pLCA by integrating scenarios generated by Integrated Assessment Models (IAM). More precisely, *premise* applies a number of transformations on energy-intensive activities found in the inventory database ecoinvent according to projections provided by the IAM. Unsurprisingly, the study shows that, within a given socio-economic narrative, the climate change mitigation target chosen affects the performance of nearly all activities in the database. This is illustrated by focusing on the effects observed on a few activities, such as systems for direct air capture of CO₂, lithium-ion batteries, electricity and clinker production as well as freight transport by road, in relation to the applied sector-based transformation and the chosen climate change mitigation target. This work also discusses the limitations and challenges faced when coupling IAM and LCA databases and what improvements are to be brought in to further facilitate the development of pLCA.

1. Introduction

The globalization and digitalization of the economy, as well as the electrification of industry and different means of transport, imply that the environmental footprint of products and services consumed is increasingly dependent on the performance of global supply chains and the energy systems that support them. As energy systems and industrial processes are rapidly changing in the attempt to reduce GHG emissions, understanding the expected changes in energy supply becomes as important as correctly modeling the product itself for performing LCA and quantifying environmental burdens in a comprehensive way. Furthermore, decision support in environmental and climate policy, for

example, usually requires insights into the performance of future technologies. Traditional LCA and its underlying static database is poorly equipped to this end. This gave way to pLCA, where projections in time are introduced in LCI [1].

The body of pLCA literature is broad [2–7]. However, advanced pLCA, in which LCA is informed by prospective energy systems or integrated assessment models, is relatively rare and has only recently gained attention [1]. First exercises linking prospective energy system models and LCA were limited to power generation, residential heating and passenger vehicles. They were used to either quantify the environmental burdens of single future technologies, or environmental impacts on a system level for different transformation pathways until the mid-century. Gibon et al. [8] used the ecoinvent database v.2.2 [9]

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<https://doi.org/10.1016/j.rser.2022.112311>

Received 30 April 2021; Received in revised form 1 February 2022; Accepted 21 February 2022

Available online 3 March 2022

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Abbreviations

BECCS	BioEnergy with Carbon Capture and Storage
BEV	Battery Electric Vehicle
CCS	Carbon Capture and Storage
CHP	Combined heat-power plant
CoP	Coefficient of Performance
DACCS	Direct Air Carbon Capture and Storage
FCEV	Fuel Cell Electric Vehicle
GHG	Greenhouse Gas(es)
IAM	Integrated Assessment Model
ICE	Internal Combustion Engine
IMAGE	Integrated Model to Assess the Global Environment
IPCC	Intergovernmental Panel on Climate Change
LCA	Life Cycle Assessment
LCI	Life Cycle Inventory
Li-NMC	Lithium Nickel Manganese Cobalt oxide
pLCA	prospective Life Cycle Assessment
pLCI	prospective Life Cycle Inventory
REMIND	REgional Model of Investment and Development
RCP	Representative Concentration Pathway
SSP	Shared Socio-economic Pathway

together with energy scenarios from the International Energy Agency and prospective industry-related inventories from the NEEDS database to generate the “THEMIS” modeling framework – an integrated, prospective hybrid LCA model that covers nine world regions with a time frame of up to 2050. Future performance of power generation technologies and selected industrial activities were integrated in the background LCI database. The THEMIS framework was further developed by Arvesen et al. [10] who presented life cycle coefficients for a wide range of future power generation technologies up to 2050. Pehl et al. [11] built on THEMIS to quantify life cycle-based energy use as well as direct and indirect GHG emissions coefficients for power generation technologies and the global electricity sector up to 2050 according to different scenarios of the IAM REMIND [12]. Finally, Luderer et al. [13] built on THEMIS to combine IAM scenarios with pLCA to explore how alternative technology choices in the power sector compare in terms of non-climate environmental impacts at the system level. Another approach more directly integrating IAM into LCA has been proposed in Refs. [14–16]. All three studies used projections from the IAM IMAGE [17] to generate prospective pLCI databases and used these to conduct the pLCA of passenger vehicles. The same approach, but with the IAM E3ME-FTT-GENIE [18], was used by Knobloch et al. [19]. They addressed the climate change impact potential of future passenger vehicles and residential heating systems. Using a similar approach, Rauner et al. [20] quantified the life cycle-based co-benefits of a global coal-exit on human health and ecosystems, this time based on prospective scenarios from REMIND; and Harpprecht et al. [21] performed a prospective LCA of key metals’ supply integrating scenarios from IMAGE. More recently, Dirnaichner et al. [22] coupled the transport model EDGE-T [23] with REMIND to calculate mid- and endpoint LCA indicators for the European passenger car fleet under different policies. The key element of these studies was a modification of the background LCI database that resulted in pLCI databases reflecting expected developments within the power generation sector.

These previous efforts were valuable as they introduced the idea of enhancing pLCA thanks to prospective scenarios of IAM and demonstrated its feasibility. However, these works were conducted with the assessment of specific systems in mind, without adjusting entire clusters of industrial activities other than the sector of interest in the pLCI database. For example, projecting efficiency gains and market developments within the electricity supply sector certainly encapsulates a

large share of the benefits to be expected when the focus is on battery electric cars and heat pumps. But other important sources of environmental damage, such as the production of metals that enter the composition of the chassis, or the cement used to build the road infrastructures, have so far not been addressed. Additionally, the technical integration of IAM scenarios in LCI databases in these studies was not designed with a large-scale applicability that would allow the use of different IAM or LCI databases.

Building on the work of Beltran and Cox [14,15], this paper presents a tool that follows a streamlined approach to integrating IAM prospective scenarios into the LCI database ecoinvent [24] to allow for pLCA. More specifically, the tool allows:

- the integration of expected transformations within five major energy-intensive sectors, namely power generation, cement and steel production, freight and passenger road transportation, and supply of conventional and alternative fuels
- applicability across different IAM
- the export of pLCI databases to different LCA software

With these functionalities, pLCI databases can be generated in a consistent manner across prospective scenarios based on socio-economic pathways and climate change mitigation targets, produced by one or multiple IAM. It also allows for the comparison of pLCI databases based on scenarios that follow a similar combination of socio-economic pathway and climate change mitigation target, but whose targets are reached based on different technology choices or constraints. Hence, *premise* contributes to the improvement of the quality of pLCA, as lack of transparency (i.e., modelling choices) and consistency (i.e., between the LCA model and the scenario) are aspects that currently undermine this emerging field, according to Ref. [4]. Finally, the tool allows for consistent and reproducible databases giving similar results regardless of the LCA software used. This last aspect echoes the work of Joyce et al. [25] which demonstrates the interesting concept of *recipes*, used to reproduce (modified) prospective LCI databases. With such aspects being handled by *premise*, practitioners can focus on the foreground modeling of the product system studied.

The next section describes the approach used to produce pLCI databases. Its benefits for prospective LCA are illustrated in Section 3 with the example of road construction, battery production, capture of CO₂ from the atmosphere and a few other cases. These examples rely on technologies that will play an important role in deep de-carbonization pathways (i.e., cement production, metals extraction and recycling, power generation). They are however energy- and material-intensive, and are expected to undergo rapid development in the next decades.

2. Method

The open-source Python library *premise* builds on the work of Beltran and Cox [14–16] and increases the extent of prospective scenario integration in LCA across multiple models (REMIND and IMAGE are used as case studies in this paper, but the method can be extended to others), different versions of the ecoinvent database (from 3.5 to 3.8) and multiple industry sectors such as power generation, cement, steel and fuel production. *premise* is currently able to work “out-of-the-box” with IMAGE and REMIND, although extending its ability to work with other IAM is straightforward. Mapping files available as part of the tool’s documentation [26] associate the IAM, *premise* and ecoinvent terminologies and minimize the effort when extending the tool’s compatibility to other IAM models. It is worth noting however that the extent to which the integration of a given sector is performed often depends on the information the IAM model can provide. *premise* limits the use of data from external sources to maximize the consistency between the resulting pLCI database and the IAM scenario it is built from.

Fig. 1 depicts the general workflow to produce a pLCI database. In Step 1, IAM prospective scenarios are used as inputs together with the

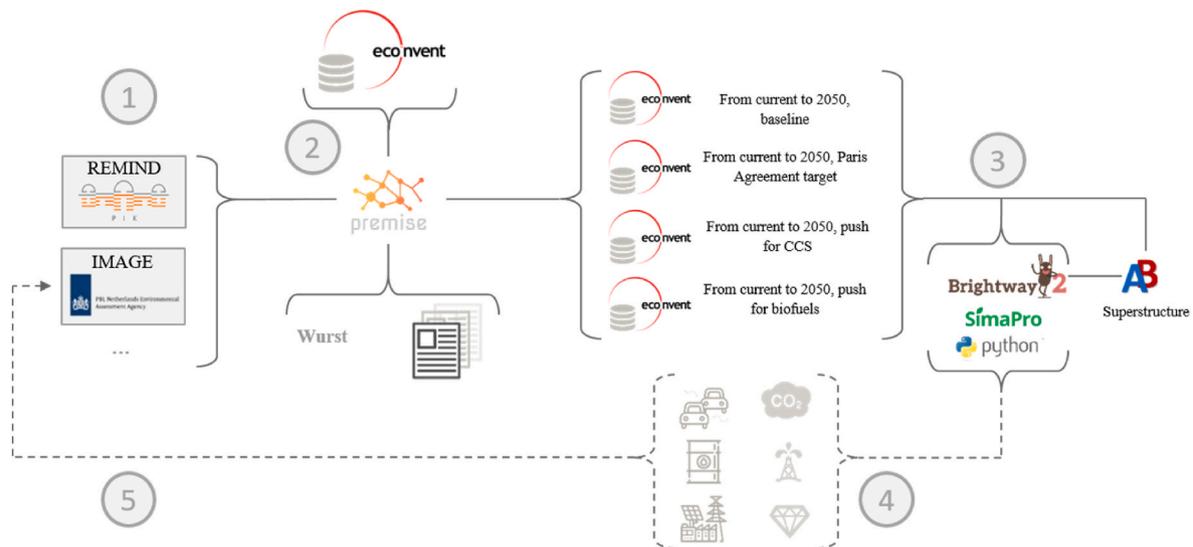


Fig. 1. General IAM-LCA coupling workflow.

LCI database (in this case, ecoinvent). Section 2.1 describes the nature and content of IAM scenarios. In Step 2, using the library *wurst* [27], *premise* operates a number of transformations on the LCI database. This step requires the use of additional inventories to represent emerging and future technologies not originally available in the LCI database. This is done by collecting inventories from the literature (e.g., hydrogen and synthetic fuel production, direct air capture, heavy-duty trucks, etc.). The second step results in a modified LCI database for a given year, transformed according to the prospective scenario chosen. Section 2.2 describes the approach used to operate such transformations. Step 3 consists of exporting the database into a format that common LCA software (i.e., Brightway2, Simapro) accept or as a set of sparse matrix representations that numerical libraries can handle. A third option consists of producing a “scenario difference file” to support a “superstructure” database that can be read by Activity Browser [28] – this option allows to write only one database to disk while being able to explore multiple scenarios – as described in Ref. [29]. Finally, Steps 4 and 5 consist of producing LCA resource and environmental indicators that feed back to the IAM. These two last steps are discussed briefly in Section 4.2.

2.1. IAM prospective scenarios

Process-detailed IAM describe transformation pathways of the interlinked energy-economy-land-climate systems. They are distinct from cost-benefit IAM in that they represent the energy system and other sources of GHG emissions as well as mitigation technologies with detailed energy stocks, flows, and conversion technologies. The reader can refer to Ref. [30] for a definition of process-based IAM. Cost-benefit IAM such as DICE [31] and Fund [32], by contrast, only have a stylized representation of GHG abatement potential as a function of carbon prices, without representing underlying system changes and their interactions.

In the present study, we use REMIND v2.1 [33] and IMAGE v3.2 [17] as illustrative examples of process-based IAM. Both IMAGE and REMIND model the energy system. On the supply side, they represent a large variety of energy conversion technologies supplying electricity, liquid fuels, hydrogen and other energy carriers. On the demand side, they represent energy services and demands from the transport (refer to Ref. [23] for REMIND or [34] for IMAGE), buildings (refer to Ref. [35] for REMIND or [36] for IMAGE) and industry sectors (refer to Ref. [37] for IMAGE). Cross-linkages to land use via bioenergy and other land-based mitigation options such as afforestation or abatement of CH₄

and N₂O emissions from land use are part of the IMAGE model [38,39], while REMIND can be coupled to the MAGPIE land use model [40], as demonstrated in Ref. [41]. To derive climate change mitigation pathways, constraints on GHG emissions are imposed (e.g., in terms of cumulative emissions until the end of the century), and the CO₂ price is adjusted iteratively to meet the target GHG emissions level. In response to the price signal, the models derive de-carbonization strategies, for instance via efficiency improvements, bioenergy [38–41], or renewables-based electrification [42]. Other environmental constraints can be considered, such as the area of land available for bioenergy and crop production. REMIND and IMAGE also represent air pollutant emissions [20] and water demands [43,44] by type of power source. A crucial difference between IMAGE and REMIND are assumptions on how the decision-making process is formed. The inter-temporal optimization used in REMIND generally implies perfect foresight by agents taking investment decisions. IMAGE, by contrast, uses recursive-dynamic modeling (i.e., system configurations in each time step are determined sequentially based on the state of the system in the previous time step). In both models, the output includes time series in five or ten-year steps of primary, secondary, final, and useful energy, for each geographical region and by fuel type, technology, or application. The number of regions differs across IAM (e.g., 12–21 for REMIND depending on configuration, 26 for IMAGE).

The IAM community has developed SSP as a means of structuring uncertainty about future socio-economic developments, such as national GDP, education and demographics [45]. In parallel, RCP describe several potential trajectories for atmospheric radiative forcing by 2100, ranging from 1.9 to 8.5 W/m². Combining both frameworks, IAM make long-term energy and land-use projections that comply with atmospheric radiative forcing targets (given by the RCP) across a set of societal and economic conditions (given by the SSP). The reader may refer to Ref. [46] for further details on how RCP and SSP relate. Bisinella et al. [4] qualify these prospective scenarios as *normative*, as they describe potential pathways from the present to reach a climate-based target in the future.

Both REMIND and IMAGE are among the five IAM that were used for deriving marker scenarios of the Shared Socio-Economic Pathways [47], and they also contributed with a substantial share of the scenarios assessed in past IPCC reports [48,49]. This study displays the integration of IAM scenarios in the LCI database ecoinvent using the SSP2 “Middle-of-the-Road” socio-economic pathway. This pathway describes developments in line with what has been historically observed in the past century. The reader can refer to Ref. [50] for additional detail on the

SSP2 pathway. Based on this, prospective scenarios that comply with the climate change mitigation targets RCP 6, 2.6 and 1.9 are presented – corresponding to a global atmospheric temperature increase by 2100 with respect to pre-industrial levels of 3.5° C., below 2° C. and 1.5° C., respectively. *premise* also works with prospective scenarios that consider other socio-economic pathways and climate change mitigation targets.

2.2. Transformations on the LCI database

A LCI database usually presents itself as a pair of matrices populated with product and emission exchanges between man-made systems

(hereafter referred to as the “technosphere”) and parts of the natural world (hereafter referred as the “biosphere”). A simplified representation of an LCI database is shown in Fig. 2.a – where both technosphere and biosphere flows are represented in the same matrix. In this example, Product A (first column) is supplied via the global market for Product A (first row). This market receives inputs from Czech Republic, Norway and German-based production activities (first column, second to fourth row). These production activities (second to fourth column) respectively require some fuel from the global fuel market and emit some CO₂ as well as NO_x (last two rows). The global fuel market requires some input from a fuel production activity, which itself leads to some emissions. It is of

	Product A	Product A	Product A	Product A	...	Fuel	Fuel	CO ₂	NO _x
Global market for A	1						-0.5		
Prod. of A from CZ	-0.5	1							
Prod. of A from NO	-0.3		1						
Prod. of A from DE	-0.2			1					
...					...				
Global market for conv. fuel		-0.1	-0.1	-0.1		1			
Prod. of conv. fuel from Global						-1	1		
CO ₂		0.3	0.3	0.3			1.2	1	
NO _x		0.02	0.02	0.02			0.05		1

a) Simplified representation of a LCI database (positive values represent outputs, negative values represent inputs).

	Product A	Product A	Product A	...	Product A	Product A	Product A	Product A	...	Fuel	Fuel	Fuel	Fuel	Fuel	CO ₂	NO _x
Global market for A	1															
European market for A	-0.3	1											-0.5			
Indian market for A	-0.2		1											-0.5		
...				...												
Production of A from CZ	-0.6				1											
Production of A from NO	-0.1					1										
Production of A from DE	-0.3						1									
Production of A from IN			-1					1								
...									...							
European market for fuel						-0.15	-0.08	-0.11		1						
Indian market for fuel											-0.12					
Conv. fuel prod. from Europe												-1	1			
Conv. fuel prod. from India														-0.2	1	
Synth. fuel prod. from India																-0.8
CO ₂						0.47	0.25	0.35	0.38				1.1	1.2	0.1	1
NO _x						0.01	0.02	0.01	0.03				0.05	0.05	0.4	1

Markets transformations
Regional production
Efficiency
Emerging technologies
Non-CO₂ emissions

b) Simplified representation of transformations operated on the LCI database (positive values represent outputs, negative values represent inputs).

Fig. 2. Schematic representation of transformations operated by *premise*.

course possible that the fuel production activity itself requires some inputs from the global market for Product A (seventh column, first row). *premise* performs a number of transformations on these matrices to reflect the data granularity and expected changes dictated by the IAM scenario, as shown in Fig. 2.b. Markets based on IAM regions are created for certain products (green shaded cells), which receive inputs from production activities located in their respective geographical scope. Production activities based on IAM regions are also created (blue shaded cells), for which a region-specific energy efficiency is applied (yellow shaded cells) – which also affects CO₂ emissions, if applicable –, as well as a region-specific correction factor for non-CO₂ emissions (grey shaded cells). Additionally, activities representing emerging technologies are added to align with the IAM scenario (orange shaded cells). Finally, inputs-consuming activities relink to the newly created market activities located in their geographical area.

premise follow the above-described principle to integrate IAM scenario data that relate to the following sectors:

- electricity generation: regional markets for biomass and electricity, efficiency update of power plants (including photovoltaic panels)
- steel and cement production: regional markets for primary and secondary steel, as well as cement, and efficiency update of primary and secondary steel and clinker production
- liquid and gaseous fuels production: regional markets for fuels (including biomass-based and synthetic fuels), CO₂ emissions update of fuel-consuming activities
- road transport: regional fleet average vehicles, notably medium- and heavy-duty trucks

For a detailed description of the approach used for each of these sectors, the reader should refer to the documentation of the tool [26].

As the following results section shows, the sector-wide transformations described above may lead to remarkable changes in the database, more even so as a few key activities, such as the provision of electricity or freight transport, provide inputs to a large number of activities.

3. Results

Section 3.1 starts with presenting the effect of the different climate change mitigation targets relative to the original LCI database, as

transformations are incrementally performed. In Sections 3.2 and 3.3, the focus is set on the effect of such transformations on a few specific activities in the database over time and across climate change mitigation targets and locations. Finally, Section 3.4 compares the GHG emissions of a specific activity for one specific climate change mitigation target and year, but across different IAM scenarios. A last case study is presented in the Supplementary Information (SI) and shows the GHG emissions for the production of 1 kg of Li-NMC-622 battery cell. This case study is selected to demonstrate the importance of integrating metal recycling as well as the integration of other sub-sectors such as heat supply at a later stage of software development of *premise*.

3.1. The influence of climate change mitigation targets on energy- and material-intensive product systems

Using the IAM REMIND, Fig. 3 illustrates the normalized effect of transformations applied to the LCI database considering the SSP2 pathway across three climate change mitigation targets for the year 2050: RCP 6.0, RCP 2.6 and RCP 1.9. Four cases are plotted to distinguish the effect of the transformation applied: (1) the electricity sector only, (2) the electricity and fuel sectors (3), the electricity, fuel and cement sectors, and (4) all sectors, which adds, among others, medium and heavy-duty trucks. An LCA has been performed on the database activities to obtain their unitary GHG emissions using the impact assessment method IPCC 2013 GWP100a, including biogenic carbon dioxide emissions, as provided in Ref. [51]. Market and treatment activities are excluded to avoid double counting in the cumulative sum. The horizontal axis shows the number of activities included. Impacts are normalized by the cumulative GHG emissions of the reference database ecoinvent v.3.8, for which the sum is denoted by '1' (or 100%). This allows comparing the carbon intensity of the database across sector transformations. Updating the *Electricity sector (1)* with variables given by the SSP2-RCP 1.9 scenario in 2050 results in a sum of cumulative GHG emissions 61% lower than that of the reference database. Note that steep increases along those curves are caused by a few activities that have a large carbon footprint, such as the construction of port facilities, hydropower plants as well as airports, respectively.

Specific sector integrations can have a significant impact for individual activities or some sector-related product systems. Interestingly, the choice in terms of climate change mitigation target also has a large influence on the results. While the scenario using the climate change

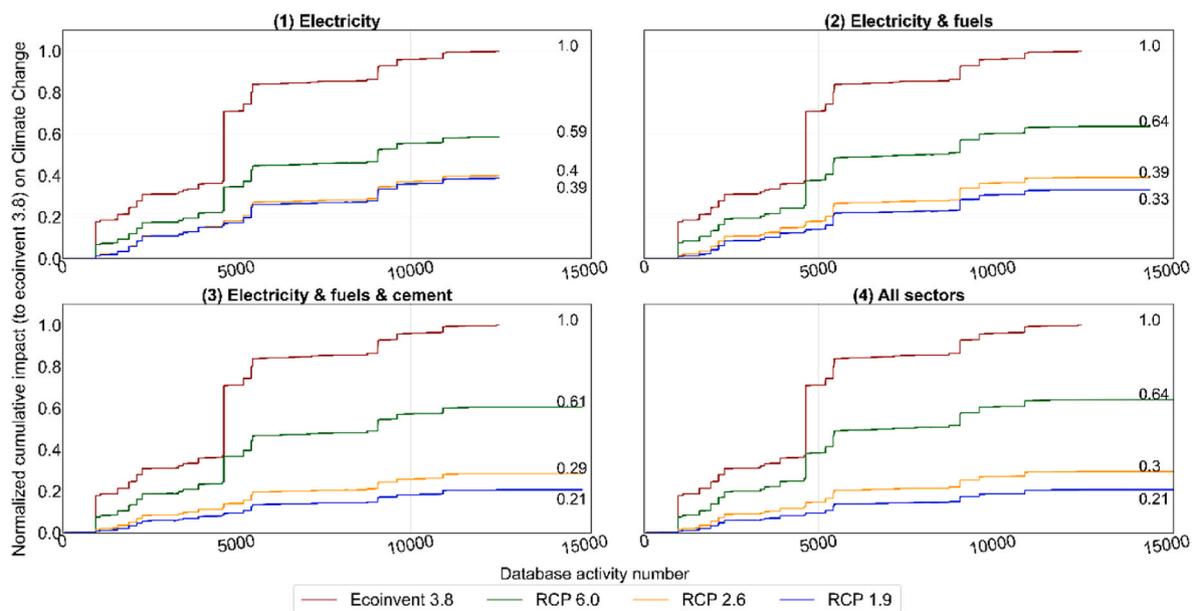


Fig. 3. Cumulative sum of GHG emissions across activities in ecoinvent, for several climate change mitigation targets, in 2050, using REMIND SSP2-based scenarios.

mitigation target RCP 6.0 reduces the sum of GHG emissions of the database by a third in 2050 considering all sector transformations, the more stringent scenario using RCP 1.9 leads to a reduction of about 80% that same year. It is worth noting that the relative difference in results between scenarios using RCP 2.6 and 1.9 is negligible by 2050 in panel 1 of Fig. 3, but it becomes more pronounced when regional fuel markets are introduced, as panel 2 shows. In the REMIND scenario using RCP 1.9, the use of woody biomass-based and synthetic fuels (from gasified biomass) represent more than 50% of the energy consumed as liquid fuel globally, leading to significant reductions in CO₂ emissions, while that share reaches only 16% in the scenario using the RCP 2.6 climate change mitigation target.

3.2. Road construction

Using again REMIND scenarios, this section analyzes the effect of two parameters on the GHG emissions of road infrastructure over time: an increasingly stringent climate change target in the prospective scenario, as well as the incremental application of sector-wide transformations in the pLCA database. Fig. 4 illustrates the GHG emissions associated with the construction of 1 m of a road, full width, normalized by its lifetime (name of the dataset: road construction, unit: meter-year, region: Rest of the World) for four different years – 2020, 2030, 2040 and 2050 – across the three different climate change mitigation targets. Again, the four subplots present different transformations of specific sectors as explained in the previous paragraphs. The bar plots also show the contribution of different components in the total GHG emissions of a meter-year of road.

The reference (static) GHG emissions for the construction of 1 m-year of road is 12.7 kg CO₂-eq. in 2020. The de-carbonization of the power system between 2020 and 2050 in the scenario complying with the RCP 6 climate mitigation target leads to a 20% reduction in GHG emissions,

and changing from RCP 6 to RCP 1.9 for that same year leads to another reduction of more than 10% in GHG emissions. However, results also indicate that aligning the fuel markets and composition in the pLCA database with the IAM scenario has the largest influence on the GHG emissions of road construction by 2050. Indeed, an additional 40% GHG emission reduction in 2050 in the scenario complying with the RCP 1.9 target is observed, as the share of biomass-based fuels in the global liquid fuel mix goes from 1% in 2020 to 54% in 2050. This underlines the importance of not limiting pLCA to the integration of changes in the sector of power generation, as previous studies have done. The transformations applied on the cement sector lead to an additional reduction in GHG emissions of about 10% using the RCP 1.9 climate mitigation target, compared to the integration of the electricity and fuel sectors. This is mainly due to three mechanisms: the utilization of cement in concrete with a lower clinker content, an improved clinker kiln efficiency as well as the capture of both process and fuel CO₂ emissions. Overall, between 2020 and 2050 in the scenario complying with the RCP 1.9 target, the clinker-to-cement ratio drops for all regions by 13% on average (they start at different levels across regions), the fuel efficiency of the kiln increases by 30 to 50% depending on the region, and the rate of carbon capture ranges between 65 and 75% by then. The integration of all sectors in 2050 using RCP 1.9 leads to reducing GHG emissions by a fourth (compared to integrating transformations associated to the electricity sector only). Interestingly, not all sector transformations lead to a reduction in GHG emissions. For all scenarios, the transformations of the electricity, fuel, steel and cement sectors reduce the overall GHG emissions as expected, while the transformation of the road transport sector increases them. More specifically in this case, the transformation of the transport sector increases GHG emissions from the “Gravel” supply. This is explained by the gravel being transported by a less performant fleet average heavy-duty vehicle than initially modeled in the reference database. This is due to the combined effects of: (1) *premise*

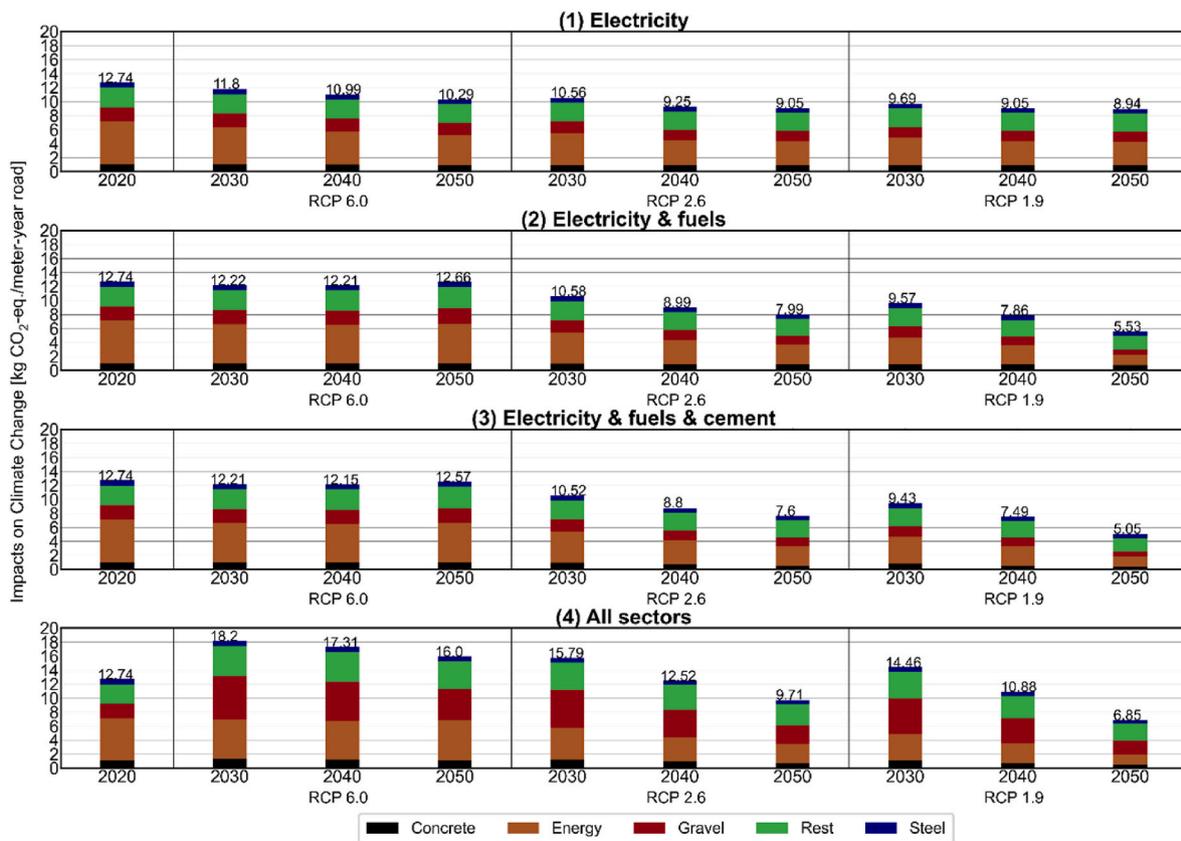


Fig. 4. GHG emission for the construction of 1 m-year of road sector transformations, for different years and climate change mitigation targets, using REMIND SSP2-based scenarios.

introducing heavy-duty trucks with a lower load factor than those originally present in ecoinvent (i.e., 9.1 tons of cargo transported against 16 tons for >32t trucks in ecoinvent), (2) a regional, but fuel-intensive, driving cycle applied for shorter distances, as well as (3) the electrification of the fleet supplied with electricity that is not de-carbonized enough (in the case of the scenario complying with the RCP 6 target).

3.3. Direct air carbon capture and storage

Using the REMIND model, the performance of atmospheric removal of carbon is analyzed across different climate change mitigation targets as well as geographical contexts. Fig. 5 illustrates the life cycle GHG emissions for the capture of 1 ton CO₂ from ambient air with its subsequent storage, referred to as DACCS. It uses the regional grid electricity as an energy source to operate the process, including a heat pump to provide the necessary heat with a CoP of 2.9, as described in Ref. [52]. Two subplots and database transformations are considered; (1) transformations of the electricity and fuel sectors, and (2) transformations of all sectors, as this specific product system is relatively energy-intensive. A negative climate change impact on the vertical axis indicates the net permanent removal of CO₂ from the atmosphere – i.e., the amount of CO₂ sequestered from which various GHG emissions that result from the life cycle of the DACCS system are subtracted. The secondary vertical axis shows the corresponding CO₂ removal effectiveness. Six geographical regions are included – Japan, Latin America, Europe, the United States, China and India – to show the region-specific climate change impacts of DACCS deployment.

The comparison between the upper and lower panels of Fig. 5 indicates very small differences between the transformations of (1) the electricity and fuel sectors and of (2) all sectors, on the climate change impacts of DACCS deployment. The life cycle climate change impacts of DACCS are largely driven by the GHG-intensity of energy sources needed for CO₂ capture; a substantial amount of grid electricity is for example required for grid-coupled DACCS systems [52], especially as the case presented produces the heat needed to re-generate the solid sorbent via a heat pump. DACCS deployment in geographical regions with GHG-intensive electricity supply and the integration of the electricity sector of a specific IAM scenario have an important influence on the total

climate change impacts (or carbon removal effectiveness). The most stringent climate change mitigation target RCP 1.9 has, for example, a minimum carbon removal effectiveness of 85–90% in 2050 – mainly due to the de-carbonization of the electricity sector – against only 45–50% for the RCP 6 target, leaving room for variation across geographical regions. Regions with a GHG-intensive electricity supply (e.g., India and China) exhibit a very low carbon removal effectiveness compared to regions with a cleaner electricity supply, such as Latin America and Europe. In fact, under the wrong conditions, the operation can exhibit a net positive sum (e.g., see India in 2020 and 2030, under the RCP 6 climate change mitigation target). First, this implies that grid-coupled DACCS systems are only suitable in geographical regions with clean electricity supply. Second, more ambitious climate policies will increase the carbon removal efficiency of grid-coupled DACCS. Both findings are in line with the work of Terlouw et al. [52].

3.4. Convergence and divergence of results between IAM

Fig. 6 illustrates the relative change in climate change impacts - normalized to the reference database ecoinvent v.3.8 – with respect to four activities; (1) clinker production (in the United States), (2) medium voltage electricity supply (global average), (3) low-alloyed steel production (global average) and (4) transportation with a heavy duty vehicle (European fleet average). The analysis combines SSP2 with the three climate change mitigation targets as used previously, to compare results from IMAGE (green lines) to those of REMIND (yellow lines) from 2020 to 2050.

Regarding emissions levels in 2050 using the RCP 6 climate change mitigation target, REMIND and IMAGE roughly agree in respect to clinker production, steel production and heavy-duty transport, but much less so about electricity production. The divergence regarding the production of global average electricity, which is a market that consists of a production volume-weighted electricity mix from the different IAM regions, is explained by the extent to which renewable sources of energy are used in RCP 6: they represent 60% of the medium voltage production mix that year in REMIND, against 21% only in IMAGE. It is worth noting that neither REMIND nor IMAGE consider the use of CCS in any sectors for that climate change mitigation target. Looking at the RCP 1.9 climate

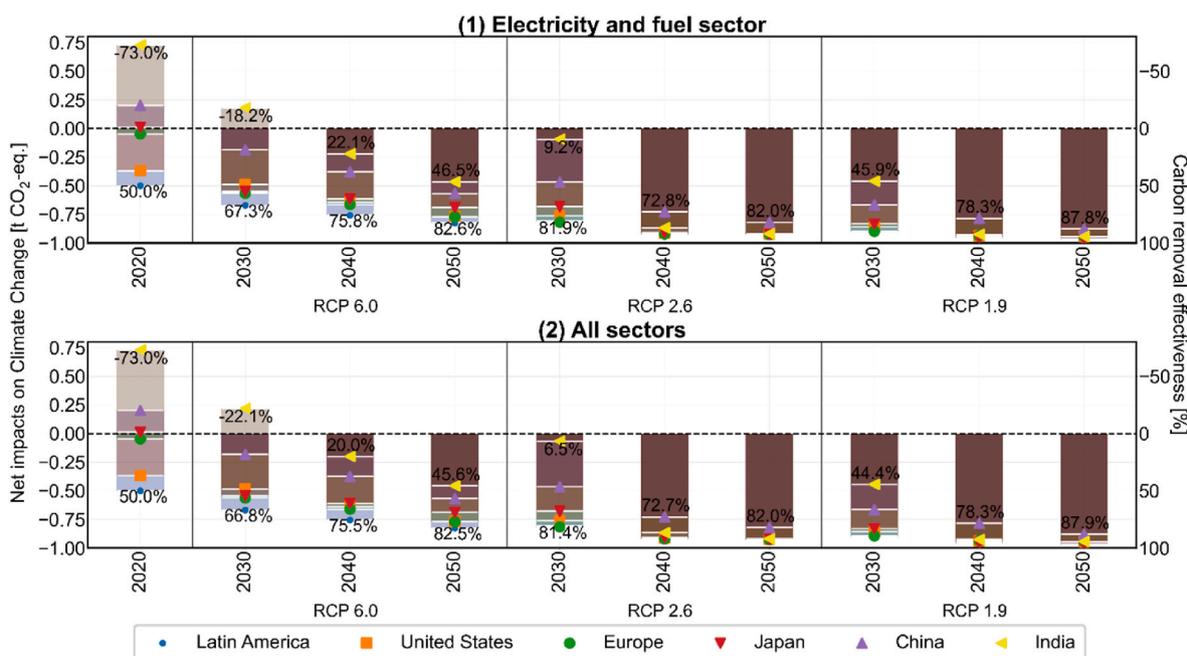


Fig. 5. Net GHG emissions for the capture and storage of 1 ton of CO₂ from the atmosphere performed by DACCS, for different IAM regions, years and climate change mitigation targets, using REMIND SSP2-based scenarios.

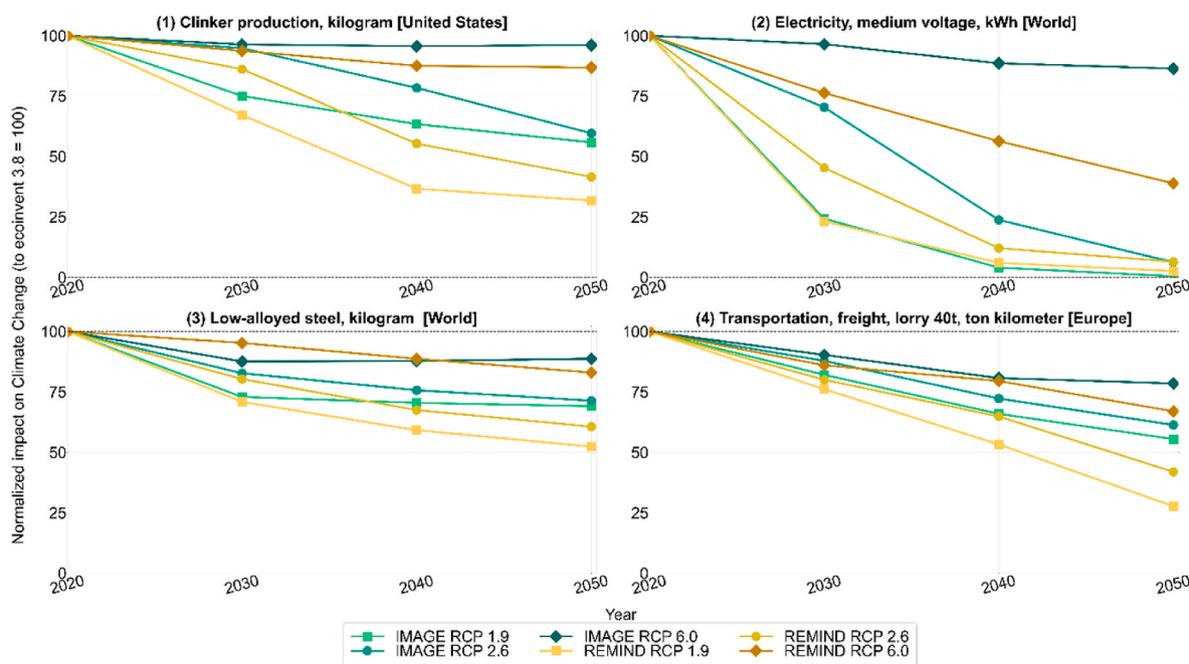


Fig. 6. Relative change in GHG emissions for different products and services compared to 2020, across years and climate change mitigation targets, using REMIND and IMAGE SPP2-based scenarios, all sectorial transformations considered.

change mitigation target, all activities calculated with the REMIND scenario scores consistently lower, except for electricity, where both models indicate an almost complete de-carbonization of the electricity production. The respective global production mixes chosen by the IAM models differ for 2050: while REMIND relies extensively on renewable sources of energy (90%, when summing hydropower, photovoltaic and wind power), IMAGE relies comparatively more on combustion-based technologies (56%), as well as renewables and nuclear power representing 31% and 11%, respectively. CCS is applied to 64% of the electricity production involving a combustion process, 20% of which is applied on biomass-based power generation (i.e., BECCS), leading to net negative GHG emissions. Regarding heavy-duty transport, the numbers presented here change over time as a result of the combination of several dynamics. First, the drivetrain efficiency and onboard energy storage for the different powertrains constituting the trucks fleet change over time. Second, the fuel and electricity regional markets that sustain the operational phase of the trucks also change over time. Lastly, the change in the fleet composition, in terms of size class and powertrain types is also considered. The share of kilometers driven by European battery and fuel cell electric trucks reaches 51% in 2050 in the REMIND scenarios (i.e., those shares do not differ across RCP). Hence, the difference in the carbon-intensity of the regional electricity mix used to sustain the operation of these trucks explains in part the difference in results between RCP 6 and RCP 1.9. The other part is explained by the share of alternative fuels in the regional diesel market between the two scenarios: 26% in the scenario complying with the RCP 6 target, against 58% for the RCP 1.9-compliant scenario. While battery electric freight vehicles are not modeled in IMAGE scenarios, fuel cell electric and plugin hybrid electric trucks constitute 94% of the kilometers driven in 2050. Here, the difference in use of alternative fuels between the two scenarios is less pronounced: 6% in the scenario complying with the RCP 6 target, against 15% in the RCP 1.9-compliant scenario. This underlines the influence the scenario-specific regional electricity mix has on the performance of the transport of heavy goods by road, together with the penetration of alternative liquid fuels.

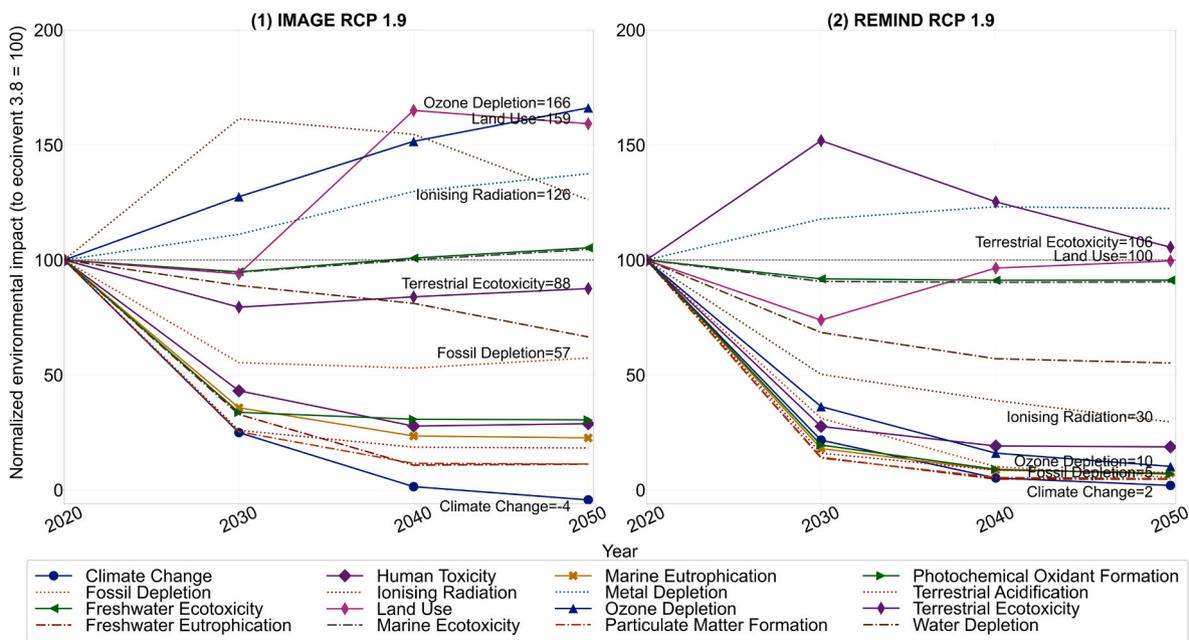
While both electricity mixes are part of a solution that is bound to reach the same climate change mitigation target, the effect on indicators other than climate change can differ. Using midpoint indicators from the

impact assessment method ReCiPe 2008 v.1.13, Fig. 7 displays the evolution over time of the impact of low voltage global electricity supply in the SSP2-RCP 1.9 scenario. The x-axis presents the year of assessment, while the y-axis indicates the characterized impact normalized by the impact of the same activity in 2020.

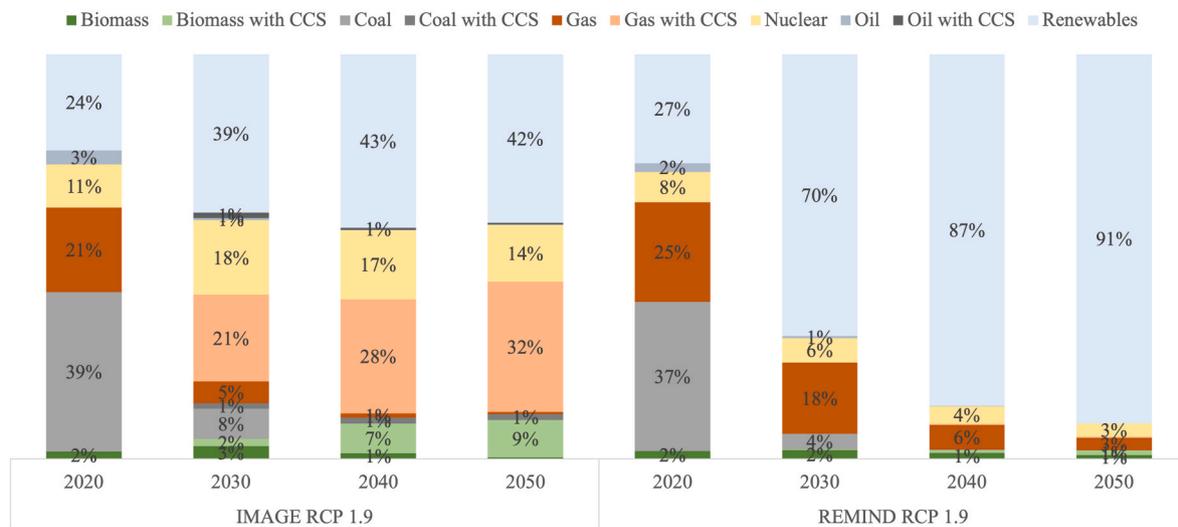
As expected, both model scenarios result in much lower impacts on climate change: -104% and -98% for IMAGE and REMIND in 2050, respectively, with respect to 2020. The negative change in climate change impacts for the IMAGE scenario (i.e., -104%) is explained by the use of BECCS which represents almost 10% of gross global production mix in 2050 (see Fig. 7b), resulting in a net permanent carbon removal.

However, the associated environmental trade-offs are different between the two scenarios. In both models, the reduction of GHG emissions comes with an increase in other types of impacts, such as agricultural and urban land occupation as well as terrestrial eco-toxicity and emissions of stratospheric ozone-depleting gases. More specifically, the IMAGE scenario relies on high capacities of conventional generators (mainly natural gas) and biomass generators, both with CCS (see Fig. 7b), while the REMIND scenario relies on high shares of renewable electricity generation, such as photovoltaics in built environments and wind-based electricity supply. Both scenarios affect requirements in land use, whether it be agricultural in the case of the IMAGE scenario (i.e., provision of “purpose grown” biomass, despite 61% of all the biomass used for power generation being supplied by agricultural and forestry residues), or a mix of agricultural and urban land in the case of the REMIND scenario (i.e., open-ground photovoltaics installations).

Also, while both scenarios reduce impacts related to fossil fuel use, the IMAGE scenario still relies to some extent on fossil energies in 2050, leading to the emission of ozone-depleting gases (see Ozone depletion indicator), since a high capacity of conventional generators (mainly natural gas-fired combined-cycle and CHP power plants) is maintained in combination with CCS. This occurs because of emissions of Halon gases along the supply chain of natural gas to the power plants. The REMIND scenario, however, exhibits higher impacts in terms of terrestrial eco-toxicity, mainly stemming from the production of commercial photovoltaic panels and wind turbines. Two reasons are identified. First, the requirement in terms of road transport for the production and maintenance of photovoltaic panels and wind turbines is high. This leads



a) Change relative to 2020 for several midpoint indicators for the provision of low voltage electricity, global average, across years, for the SSP 2-RCP 1.9 scenario, for IMAGE (left) and REMIND (right).



b) Gross global electricity production mix, across years, for the SSP 2-RCP 1.9 scenario, for IMAGE (left) and REMIND (right). Technologies are aggregated by type of energy carrier.

Fig. 7. Relative changes in midpoint indicators relative to 2020 (a) and corresponding gross production mix (b).

to toxic metal emissions coming from brake and tire abrasion (e.g., cadmium, copper). Combined to the important share of biodiesel from energy crops in the regional diesel market, it also leads to a significant use of insecticide (e.g., cypermethrin) to grow the necessary crops – it can however be argued that the extent and toxicity of insecticide use in the future may be reduced, which is something neither *premise* nor IAM consider. Second, the emissions of fine metal particles (e.g., lead, tin and silver) during the manufacture of photovoltaic cells eventually deposit on the surrounding land. It is also worth noting that both scenarios increase the strain on metal reserves, notably rare earths, although the

differences between the two scenarios are small. Future dynamics in terms of metals recycling are not considered by *premise* at the moment – hence these numbers are probably overestimating the need for metals extraction in the future. Finally, nuclear-based electricity generation is maintained to a large extent in the IMAGE scenario (between 14% and 18% of the gross production mix in the future, against 11% today), explaining the increased emissions of ionizing radiation.

These results imply that each scenario requires a thorough assessment of environmental trade-offs and co-benefits associated to reaching a de-carbonized economy in the future. These additional indicators

could prove useful in designing climate change mitigation scenarios.

4. Discussion

Some aspects that relate to the challenges faced when coupling and finding a common terminology between IAM scenarios and LCA, but also the purpose and future use cases of such exercise, are discussed in this section.

4.1. Limitations

The largest challenge in coupling IAM and LCA is the potential mismatch between the modeled technologies in IAM and the corresponding life cycle inventories. In order to reduce the computational complexity, IAM group similar technologies and assign generic properties to them based on historical information. They also cluster regions spatially based on their location and socio-economic properties. Life cycle inventories have no computational constraints and strive to be as detailed and differentiated between technologies and geographic scopes as possible. The discrepancy in data granularity might lead to semantic ambiguity, where an IAM process could have one or multiple corresponding activities with unspecified shares in the LCI database. A typical example would be the lack of a distinction based on grades for steel products in the IAM, while such distinction has a significant importance in terms of material and energy inventory in the LCI database. It could also lead to semantic mismatch, for example if the region “Europe” has a different geographical definition for the IAM and for the LCI database. While these discrepancies cannot be alleviated completely without altering the resolution of the IAM and/or the LCI database, it is possible to minimize their impact through a proper understanding and a correct interpretation of the data on both sides. The use of transparent mapping files, described in the tool’s documentation [26], that link the variables between models can help, if not to solve such ambiguities, to better trace them back.

Also, for some transformations, *premise* relies at the moment on external data sources, such as the GAINS model for the projections on the reduction of non-CO₂ emissions, but also on inventories from the LCA literature for various emerging technologies. This can potentially introduce modelling inconsistencies as indicated also in the *premise* documentation [26].

Furthermore, there is also a temporal constraint: while IAM provide projections up to 2100, it seems difficult to extend reliably the coupling between IAM and LCA beyond 2050–2060. While the LCI database can accommodate incremental shifts in efficiency, which is what *premise* does, it cannot anticipate potentially disruptive shifts in technologies (e. g., nuclear fusion).

Finally, results given for toxicity-related indicators (e.g., human toxicity, terrestrial toxicity, ecosystems toxicity) should be regarded as highly uncertain, especially when those increase or become important when normalized. Such case appears with the use of pesticide to produce biodiesel described in the previous section. IAM do not carry any information regarding the use, fate or toxicity of chemicals in the future, and scaling the use of current chemicals up or down as *premise* currently does can be misleading.

4.2. Integrating IAM projections into LCA: a means to a larger goal?

The results have shown the effects of integrating IAM projections into the LCI database, by contrasting results calculated with the static ecoinvent database v.3.8 to those derived from the coupling. Unlike a scenario-based pLCA relying on independent assumptions, the use of the IAM provides a coherent narrative, balancing the global perspective with the regional singularities. The benefits for IAM are equally unequivocal: through the LCA coupling, it is possible to quantify impacts that are not directly modeled in the IAM, such as impacts on human health and ecosystems, land use or metal depletion. Hence, without

affecting the computational complexity of the IAM or straying away from the objective of system de-carbonization, it is possible to quantify the environmental side effects of different scenarios. Ultimately, it may be possible to feed the LCA impacts back into the IAM, for instance by monetizing them and recalculating the cost-optimal solution in an iterative process, or by introducing additional constraints. This has the potential to provide a holistic approach to the system transformation, anticipate resource bottlenecks and environmental criticalities, and identify transformation strategies accounting for multiple environmental goals.

4.3. Next steps

The open-source library *premise* is continuously improved, with new features expected in the short and mid-term. The addition of new sectors – heat supply, extraction, refining and recycling of metals, and negative CO₂ emission technologies – will expand its functionality. Also, the foreseeable dominance of renewable energy sources in the projected electricity mixes highlights the necessity to improve inventories of renewable power plants, such as photovoltaic panels and wind turbines, in particular their efficiency and load factor, which are only rudimentarily represented in ecoinvent. Currently, *premise* only adjusts the efficiency of photovoltaic panels by modifying the panel surface needed per kilowatt of peak power capacity installed. But ultimately, parametrized models such as those developed in Refs. [53,54] for wind turbines and [55] for photovoltaic panels will be needed to create region and year-specific inventories.

Furthermore, new collaborations with other IAM are sought after in order to develop an interface to additional models. In addition, substantial efforts are channeled into improved reporting. Ultimately, a fully detailed report should be generated with each pLCI database produced, to indicate all the changes made as well as the boundary conditions behind the scenario narrative and climate change mitigation target used by the IAM.

5. Conclusion

As commitments to curb emissions of greenhouse gases accelerate, rapid transformations are expected in energy systems and industries. This makes prospective LCA useful to assess the environmental performance of quickly developing, but also emerging or yet-to-be-developed technologies. This study shows that it is possible to streamline the production of comprehensive pLCI databases in order to facilitate the development and increase the quality of pLCA studies. It also shows that the scenario narrative chosen as well as the selected IAM and its specific way of achieving climate change mitigation goals can have significant effects on the LCI database (and thereby any foreground model that relies on it). There is therefore some critical uncertainty in any LCA study using such databases, as the de-carbonization pathway and the actual technological breakthroughs are unknown. However, *premise* allows to have a broad idea on the effect such uncertainty can have on the LCI database as comparing a multitude of scenario-specific databases is made easy.

Funding sources

This work has partially received financial support from the Copernicus Project Ariadne (FKZ 03SFK5A), funded by the German Federal Ministry of Education and Research.

Data availability

The tool presented in this study as well as its source code are available in the following repository: <https://github.com/polca/premise>.

Credit author statement

Romain Sacchi: Methodology, Software, Writing – original draft, **Alois Dirnaichner:** Methodology, Software, **Christopher L. Mutel:** Methodology, Software, **Christian Bauer:** Supervision, Writing – review & editing, **Brian L. Cox:** Conceptualization, Writing – review & editing, **Vassilis Daioglou:** Writing – review & editing, **Tom Terlouw:** Software, Visualization, **Kais Siala:** Writing – review & editing, **Gunnar Luderer:** Supervision, Writing – review & editing.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgement

The authors would like to thank Bernhard Steubing, Angélica Mendoza Beltrán and Laurent Vandepaer for their early contribution and discussions around the topic of IAM-LCA integration. The authors are also grateful to Tapajyoti Ghosh and Patrick Lamers from the National Renewable Energy Laboratory (NREL) for the extensive testing of the tool.

Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.rser.2022.112311>.

References

- van der Giesen C, Cucurachi S, Guinée J, Kramer GJ, Tukker A. A critical view on the current application of LCA for new technologies and recommendations for improved practice. *J Clean Prod* 2020;259:120904. <https://doi.org/10.1016/j.jclepro.2020.120904>.
- van der Giesen C, Cucurachi S, Guinée J, Kramer GJ, Tukker A. A critical view on the current application of LCA for new technologies and recommendations for improved practice. *J Clean Prod* 2020;259:120904. <https://doi.org/10.1016/J.JCLEPRO.2020.120904>.
- Thonemann N, Schulte A, Maga D. How to conduct prospective life cycle assessment for emerging technologies? A systematic review and methodological guidance. *2020;12 Sustain* 2020;12:1192. <https://doi.org/10.3390/SU12031192.1192>.
- Bisinella V, Christensen TH, Astrup TF. Future scenarios and life cycle assessment: systematic review and recommendations. *Int J Life Cycle Assess* 2021;26:2143–70. <https://doi.org/10.1007/S11367-021-01954-6>. 2611 2021.
- Thomassen G, Van Passel S, Dewulf J. A review on learning effects in prospective technology assessment. *Renew Sustain Energy Rev* 2020;130:109937. <https://doi.org/10.1016/J.RSER.2020.109937>.
- Arvidsson R, Tillman AM, Sandén BA, Janssen M, Nordelöf A, Kushnir D, et al. Environmental assessment of emerging technologies: recommendations for prospective LCA. *J Ind Ecol* 2018;22:1286–94. <https://doi.org/10.1111/JIEC.12690>.
- Moni SM, Mahmud R, High K, Carbajales-Dale M. Life cycle assessment of emerging technologies: a review. *J Ind Ecol* 2020;24:52–63. <https://doi.org/10.1111/JIEC.12965>.
- Gibon T, Wood R, Arvesen A, Bergesen JD, Suh S, Hertwich EG. A methodology for integrated, multi-regional life cycle assessment scenarios under large-scale technological change. *Environ Sci Technol* 2015;49:11218–26. <https://doi.org/10.1021/acs.est.5b01558>.
- Frischknecht R, Jungbluth N, Althaus HJ, Doka G, Dones R, Heck T, et al. The ecoinvent database: overview and methodological framework. *Int J Life Cycle Assess* 2005;10:3–9. <https://doi.org/10.1065/lca2004.10.181.1>.
- Arvesen A, Luderer G, Pehl M, Bodirsky BL, Hertwich EG. Deriving life cycle assessment coefficients for application in integrated assessment modelling. *Environ Model Software* 2018;99:111–25. <https://doi.org/10.1016/j.envsoft.2017.09.010>.
- Pehl M, Arvesen A, Humpenöder F, Popp A, Hertwich EG, Luderer G. Understanding future emissions from low-carbon power systems by integration of life-cycle assessment and integrated energy modelling. *Nat Energy* 2017;2:939–45. <https://doi.org/10.1038/s41560-017-0032-9>.
- Aboumahboub T, Auer C, Bauer N, Baumstark L, Bertram C, Bi S, et al. Model documentation | version 2.1.0 | REMIND - REgional model of INvestments and development. 2020. <https://doi.org/10.5281/ZENODO.3730919>.
- Luderer G, Pehl M, Arvesen A, Gibon T, Bodirsky BL, de Boer HS, et al. Environmental co-benefits and adverse side-effects of alternative power sector decarbonization strategies. *Nat Commun* 2019;10:1–13. <https://doi.org/10.1038/s41467-019-13067-8>.
- Mendoza Beltran A, Cox B, Mutel C, van Vuuren DP, Font Vivanco D, Deetman S, et al. When the background matters: using scenarios from integrated assessment models in prospective life cycle assessment. *J Ind Ecol* 2018;24:64–79. <https://doi.org/10.1111/jiec.12825>.
- Cox B, Mutel CL, Bauer C, Mendoza Beltran A, Van Vuuren DP. Uncertain environmental footprint of current and future battery electric vehicles. *Environ Sci Technol* 2018;52:4989–95. <https://doi.org/10.1021/acs.est.8b00261>.
- Cox B, Bauer C, Mendoza Beltran A, van Vuuren DP, Mutel CL. Life cycle environmental and cost comparison of current and future passenger cars under different energy scenarios. *Appl Energy* 2020;269:115021. <https://doi.org/10.1016/j.apenergy.2020.115021>.
- Stehfest E, Vuuren D van, Kram T, Bouwman L, Alkemade R, Bakkenes M, et al. Integrated assessment of global environmental change with IMAGE 3.0. Model description and policy applications. 2014.
- Mercure JF, Pollitt H, Edwards NR, Holden PB, Chewprecha U, Salas P, et al. Environmental impact assessment for climate change policy with the simulation-based integrated assessment model E3ME-FTT-GENIE. *Energy Strateg Rev* 2018;20:195–208. <https://doi.org/10.1016/j.esr.2018.03.003>.
- Knobloch F, Hanssen SV, Lam A, Pollitt H, Salas P, Chewprecha U, et al. Net emission reductions from electric cars and heat pumps in 59 world regions over time. *Nat Sustain* 2020;3:437–47. <https://doi.org/10.1038/s41893-020-0488-7>.
- Rauner S, Bauer N, Dirnaichner A, Dingenen R Van, Mutel C, Luderer G. Coal-exit health and environmental damage reductions outweigh economic impacts. *Nat Clim Change* 2020;10:308–12. <https://doi.org/10.1038/s41558-020-0728-x>.
- Harpprecht C, van Oers L, Northey SA, Yang Y, Steubing B. Environmental impacts of key metals' supply and low-carbon technologies are likely to decrease in the future. *J Ind Ecol* 2021;25:1543–59. <https://doi.org/10.1111/JIEC.13181>.
- Dirnaichner A, Rottoli M, Sacchi R, Rauner S, Cox B, Mutel CL, et al. Life-cycle impacts from different decarbonization pathways for the European car fleet. *Environ Res Lett* 2022. <https://doi.org/10.1088/1748-9326/ac4fdb>.
- Rottoli M, Dirnaichner A, Kyle P, Baumstark L, Pietzcker R, Luderer G. Coupling a detailed transport model to the integrated assessment model REMIND. *Environ Model Assess* 2021;1:3. <https://doi.org/10.1007/s10666-021-09760-y>.
- Wernet G, Bauer C, Steubing B, Reinhard J, Moreno-Ruiz E, Weidema B. The ecoinvent database version 3 (part I): overview and methodology. *Int J Life Cycle Assess* 2016;21:1218–30. <https://doi.org/10.1007/s11367-016-1087-8>.
- Joyce PJ, Björklund A. Futura: a new tool for transparent and shareable scenario analysis in prospective life cycle assessment. *J Ind Ecol* 2021. <https://doi.org/10.1111/JIEC.13115>.
- Sacchi R. User guide — premise 1.0.0 documentation 2022. accessed January 24, <https://premise.readthedocs.io/en/latest/>; 2022.
- Mutel CL. Introduction — wurst 0.1 documentation 2017. accessed April 25, <https://wurst.readthedocs.io/>; 2021.
- Steubing B, de Koning D, Haas A, Mutel CL. The Activity Browser — an open source LCA software building on top of the brightway framework. *Softw Impacts* 2020;3:100012. <https://doi.org/10.1016/j.simpa.2019.100012>.
- Steubing B, de Koning D. Making the use of scenarios in LCA easier: the superstructure approach. *Int J Life Cycle Assess* 2021;26:2248–62. <https://doi.org/10.1007/s11367-021-01974-2/FIGURES/7>.
- Weyant J. Some contributions of integrated assessment models of global climate change. *Rev Environ Econ Pol* 2017;11:115–37. <https://doi.org/10.1093/reep/rew018>.
- Nordhaus W. Estimates of the social cost of carbon: concepts and results from the DICE-2013R model and alternative approaches. *J Assoc Environ Resour Econ* 2014;1:273–312. <https://doi.org/10.1086/676035>.
- Anthoff D, Tol RSJ. The uncertainty about the social cost of carbon: a decomposition analysis using fund. *Clim Change* 2013;117:515–30. <https://doi.org/10.1007/s10584-013-0706-7>.
- Baumstark L, Bauer N, Benke F. REMIND2.1: transformation and innovation dynamics of the energy-economic system within climate and sustainability limits. *Geosci Model Dev Discuss* 2021. <https://doi.org/10.5194/gmd-2021-85>.
- Girod B, van Vuuren DP, de Vries B. Influence of travel behavior on global CO2 emissions. *Transport Res Part A Policy Pract* 2013;50:183–97. <https://doi.org/10.1016/j.tra.2013.01.046>.
- Levesque A, Pietzcker RC, Baumstark L, Luderer G. Deep decarbonisation of buildings energy services through demand and supply transformations in a 1.5°C scenario. *Environ Res Lett* 2021. <https://doi.org/10.1088/1748-9326/abd0f7>.
- Daioglou V, van Ruijven BJ, van Vuuren DP. Model projections for household energy use in developing countries. *Energy* 2012;37:601–15. <https://doi.org/10.1016/j.energy.2011.10.044>.
- van Sluisveld M, de Boer HS, Daioglou V, Hof A, van Vuuren D. A race to zero - assessing the position of heavy industry in a global net-zero CO2 emissions context. *Energy Clim Chang* 2021.
- Daioglou V, Doelman JC, Wicke B, Faaij A, van Vuuren DP. Integrated assessment of biomass supply and demand in climate change mitigation scenarios. *Global Environ Change* 2019;54:88–101. <https://doi.org/10.1016/j.gloenvcha.2018.11.012>.
- Doelman JC, Stehfest E, Vuuren DP, Tabeau A, Hof AF, Braakhekke MC, et al. Afforestation for climate change mitigation: potentials, risks and trade-offs. *Global Change Biol* 2020;26:1576–91. <https://doi.org/10.1111/gcb.14887>.
- Dietrich JP, Bodirsky BL, Humpenöder F, Weindl I, Stevanović M, Karstens K, et al. MAgPIE 4-a modular open-source framework for modeling global land systems. *Geosci Model Dev* 2019;12:1299–317. <https://doi.org/10.5194/gmd-12-1299-2019>.

- [41] Bauer N, Klein D, Humpenöder F, Kriegler E, Luderer G, Popp A, et al. Bio-energy and CO₂ emission reductions: an integrated land-use and energy sector perspective. *Clim Change* 2020;163:1675–93. <https://doi.org/10.1007/s10584-020-02895-z>.
- [42] Luderer G, Madeddu S, Merfort L, Ueckerdt F, Pehl M, Pietzcker R, et al. Impact of declining renewable energy costs on electrification in low-emission scenarios. *Nat Energy* 2021;1:1–11. <https://doi.org/10.1038/s41560-021-00937-z>.
- [43] Mouratiadou I, Bevione M, Bijl DL, Drouet L, Hejazi M, Mima S, et al. Water demand for electricity in deep decarbonisation scenarios: a multi-model assessment. *Clim Change* 2018;147:91–106. <https://doi.org/10.1007/s10584-017-2117-7>.
- [44] Bijl DL, Biemans H, Bogaart PW, Dekker SC, Doelman JC, Stehfest E, et al. A global analysis of future water deficit based on different allocation mechanisms. *Water Resour Res* 2018;54:5803–24. <https://doi.org/10.1029/2017WR021688>.
- [45] O'Neill BC, Kriegler E, Riahi K, Ebi KL, Hallegatte S, Carter TR, et al. A new scenario framework for climate change research: the concept of shared socioeconomic pathways. *Clim Change* 2014;122:387–400. <https://doi.org/10.1007/s10584-013-0905-2>.
- [46] Hausfather Z. Explainer: how “shared socioeconomic pathways” explore future climate change. *Carbon* 2018. accessed April 25, 2021, <https://www.carbonbrief.org/explainer-how-shared-socioeconomic-pathways-explore-future-climate-change>.
- [47] Riahi K, van Vuuren DP, Kriegler E, Edmonds J, O'Neill BC, Fujimori S, et al. The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: an overview. *Global Environ Change* 2017;42:153–68. <https://doi.org/10.1016/j.gloenvcha.2016.05.009>.
- [48] Clarke L, Jiang K, Akimoto K, Babiker M, Blanford G, Fisher-Vanden K, et al. Assessing transformation pathways. In: *Climate change 2014: mitigation of climate change. Contribution of working group III to the fifth assessment report of the intergovernmental panel on climate change*. Cambridge, United Kingdom New York, NY, USA: Cambridge Univ. Press; 2014. p. 413–510.
- [49] Rogelj J, Popp A, Calvin KV, Luderer G, Emmerling J, Gernaat D, et al. Scenarios towards limiting global mean temperature increase below 1.5 °C. *Nat Clim Change* 2018;8:325–32. <https://doi.org/10.1038/s41558-018-0091-3>.
- [50] Fricko O, Havlik P, Rogelj J, Klimont Z, Gusti M, Johnson N, et al. The marker quantification of the Shared Socioeconomic Pathway 2: a middle-of-the-road scenario for the 21st century. *Global Environ Change* 2017;42:251–67. <https://doi.org/10.1016/j.gloenvcha.2016.06.004>.
- [51] Sacchi R. *premise.gwp 2022*. <https://github.com/polca/premise.gwp>. [Accessed 27 January 2022]. accessed.
- [52] Terlouw T, Treyer K, Bauer C, Mazzotti M. Life cycle assessment of direct air carbon capture and storage with low-carbon energy sources. *Environ Sci Technol* 2021;55:11397–411. <https://doi.org/10.1021/acs.est.1c03263>.
- [53] Sacchi R, Besseau R, Pérez-López P, Blanc I. Exploring technologically, temporally and geographically-sensitive life cycle inventories for wind turbines: a parameterized model for Denmark. *Renew Energy* 2019;132:1238–50. <https://doi.org/10.1016/j.renene.2018.09.020>.
- [54] Besseau R, Sacchi R, Blanc I, Pérez-López P. Past, present and future environmental footprint of the Danish wind turbine fleet with LCA_WIND_DK, an online interactive platform. *Renew Sustain Energy Rev* 2019;108:274–88. <https://doi.org/10.1016/j.rser.2019.03.030>.
- [55] Tannous S, Besseau R, Prieur-Vernat A, Clavreul J, Payeur M, Blanc I, et al. A parameterized model for the estimation of life cycle environmental impacts of crystalline PV systems. 2019.