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1 Unequal 'drivers': on the inequality of mobility
2 emissions in Germany

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11 **Abstract**

12 Transportation and mobility patterns contribute to greenhouse gas emissions. Un-
13 derstanding the drivers of these emissions, particularly for high emitters, is key to
14 designing appropriate climate and mobility policies. In this article, we study the
15 distribution of emissions from mobility in Germany and their drivers. We use a 2017
16 nation-wide mobility survey to calculate the carbon footprint of individuals associated
17 with day-to-day and long-distance travels. We use quantile regression to investigate
18 both socio-economic and attitudinal drivers of emissions across different categories
19 of emitters, and for different mobility types. We discuss our results with respect to
20 previous findings in the literature. Overall, we find that the top 10% of emitters are
21 responsible for 51% of total emissions, and for 80% of emissions from long-distance
22 travel. The statistical analysis reveals strong differences regarding the contribution of
23 socio-economic drivers such as income or location at different levels of emissions. At-
24 titudes towards different transportation modes also strongly correlate with differences
25 in mobility behaviors.

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1 Introduction

Personal mobility facilitates most of our social and economic activities. It enables us to reach distant places for purpose of work, social life or leisure, and – today more than ever – acts as a marker of social status (Ellaway et al., 2003; Mann and Abraham, 2006; Zhao and Zhao, 2020), thereby catering a multitude of needs. However, traveling often comes at great costs to society and the natural environment through externalities, such as accidents, traffic congestion, local pollution or greenhouse gas emissions (GHGs).

Transportation represents a significant share of global GHGs. Worldwide, the IEA estimates that the transportation sector amounted to 37% of CO₂ emissions from end-use sectors in 2021.¹ According to Lamb et al. (2021) global transport emissions have increased by 2% yearly between 1990 and 2018. While overall emissions have decreased in Europe, transport emissions have increased by 0.4%, suggesting that reducing emissions in this sector proves particularly difficult. This stresses the importance of better understanding the drivers of these emissions to achieve long-term climate targets.

Emissions from mobility are very unequally distributed across individuals and highly concentrated at the top of the distribution. For instance, Brand and Boardman (2008) find that in the UK, the top 10% of GHG emitters are responsible for 43% of mobility emissions, while Ko et al. (2011) show that the top 10% in the Seoul metropolis area are responsible for 63% of carbon dioxide (CO₂) emissions. Thus, policies aiming at reducing emissions in this sector should look into the determinants of the high emitters to understand the most effective ways to reduce these emissions and to unravel the potential distributive effects.

Previous literature emphasizes the role of socio-economic and demographic determinants on transport-related emissions (Büchs and Schnepf, 2013; Brand et al., 2013; Reichert et al., 2016). Such characteristics include income, age, gender or education and they explain part of the differences observed across individuals. For instance, in Germany, Aamaas et al. (2013) suggest that the contribution of the highest income group is 2.5 times greater than that of the lowest income group, while Brand and Preston (2010) estimate the ratio to be 3.5 in the UK. Geographic location is also an important driver, with households in densely urbanized areas typically emitting less for daily travels than those in less dense areas, while the contrary holds for long-distance trips (Czepkiewicz et al., 2018).

Besides, psycho-social variables are important determinants of mobility decisions (Pronello and Gaborieau, 2018). One can distinguish between different psychological factors, such as values, beliefs or attitudes. Values often relate to broader life goals and motivations (Schwartz and Bilsky, 1990), whereas beliefs are related to information about an object (Hoffmann et al., 2020), like the perceived necessity or ability to engage in a behaviour. This study integrates the impact of attitudes, which we define following Eagly and Chaiken as “a psychological tendency that is expressed by evaluating a particular entity with some degree of favor or disfavor.” (Eagly and Chaiken (1998), p. 583). While gaps between intentions and observed behaviours

¹Source: <https://www.iea.org/topics/transport>

69 are common, Hunecke et al. (2010) find that attitudes are a better predictor of travel
70 mode choice than values.

71 The contribution of this article is threefold. First, we are adding an analysis of
72 drivers of long-and short-distance mobility for Germany, the most populous European
73 country with the highest emissions for domestic transport, as well as international
74 aviation.² Compared to Reichert et al. (2016), who analyse German GHG emissions
75 from daily and long-distance travel as well, we use more recent data and focus on
76 different emitter categories. Furthermore, this study adds a focus on psychological
77 factors, in addition to spatial attributes and other socio-economic characteristics.

78 Second, we add attitudinal drivers in a large-scale (national) context. There is a
79 vast literature that integrates attitudes with mobility. However, most of these studies
80 are either performed with small samples or focus on very specific contexts, such as
81 bicycle commuters (Li et al., 2013; Heinen et al., 2011), old age groups (Haustein,
82 2012), day trippers (Anable, 2005) or certain urban areas (Pronello and Camusso,
83 2011; Wang et al., 2023). In addition, only few studies have effectively linked travel
84 emissions or environmental impact with attitudinal factors (Nilsson and Küller, 2000;
85 Hunecke et al., 2007; Ao et al., 2019; Mattioli et al., 2023). Prillwitz and Barr
86 (2011a) highlight another issue, namely that most of these studies focus on daily
87 travel behaviour. In this article, we integrate attitudinal variables with both daily
88 and occasional travel using a large nation-wide sample of the German population. We
89 are only aware of one other study that includes psychological variables for mobility
90 emissions on a national scale. Mattioli et al. (2023) study emitter types in the UK
91 focusing in particular on people with 'dissonant' travel behaviours.

92 Third, we go beyond the average effect of socio-economic, demographic and atti-
93 tudinal determinants of mobility emissions by analysing the role of drivers at different
94 levels of the emitter distribution on the national level using quantile regression. While
95 some studies have focused on different types of emitters, they apply the concept to
96 smaller geographical areas Ko et al. (2011); Bel and Rosell (2017); Leroutier and
97 Quirion (2022) or focus only on land-based passenger transport Brand et al. (2013).
98 Focusing on average effects of any determinants of emissions hinders insights into
99 how the same drivers can play different roles at distinct levels of the distribution,
100 and in particular for high emitters. This is especially relevant in the case of trans-
101 portation, because emissions are very unequally distributed. Further, oft-used OLS
102 regressions are sensitive to outliers, which are typical as mobility data generally cover
103 observations of mobility for a short period of time.

104 Here we address this issue by employing quantile regression, which allows us to
105 determine the effect of the drivers along the distribution of the outcome variable
106 (Koenker and Hallock, 2001). Quantile regression has been applied to a wide variety
107 of ecological and economic data (Koenker, 2017; Cade and Noon, 2003), and is par-
108 ticularly suited to overcome heterogeneity of variance along the distribution, which
109 we show holds in our mobility dataset. This allows us to identify which factors are
110 most important among the high emitters.

111 In the context of households' emissions, we are aware of only two applications
112 of quantile regression. Han et al. (2015) analyse households' total carbon footprint.

²Source: EEA

113 Closer to our work, Bel and Rosell (2017) apply the method to study the socio-
114 economic drivers of emissions in the context of urban mobility. Our study differs
115 from theirs in several respects. We extend the analysis to the national level, and to
116 all mobility types. Bel and Rosell (2017), like most studies, focus on daily urban
117 mobility, while we also investigate long-distance trips, which represent a significant
118 share of individual emissions. In addition to socio-economic variables, we investigate
119 the association of emissions with attitudinal variables, which we show explain a great
120 share of the heterogeneity in emissions.

121 We rely on the 2017 German mobility survey *Mobilität in Deutschland* (MiD).
122 The survey contains detailed data on household travels, which we use to quantify the
123 greenhouse gas emissions associated with different types of transportation modes and
124 purposes for each individual. We aggregate these emissions at the individual level and
125 perform both OLS and quantile regressions to identify the drivers of total emissions,
126 as well as those of emissions from long-distance travels.

127 Our study quantifies both the socio-economic and attitudinal drivers of emissions.
128 Regarding the former, our results are broadly consistent with earlier studies, and
129 confirm that income Büchs and Schnepf (2013); Brand and Preston (2010) and edu-
130 cation Brand et al. (2013); Holz-Rau et al. (2014) are strongly correlated with mobility
131 emissions. However, the analysis suggests some saturation effect of income, meaning
132 that being in the highest income group or the one below it has a similar impact on
133 emissions. We also unravel the different role these characteristics play for daily and
134 long-distance mobility.

135 We shed light on the strong correlation between individuals' attitudes and their be-
136 haviors. In particular, we find that attitudes towards different transportation modes
137 can explain some of the heterogeneity in mobility patterns. The evidence suggests
138 that a positive attitudes towards driving and biking are each strongly correlated with
139 higher overall emissions. For each of these associations, our method allows us to
140 analyse the effect at different levels of the distribution, and in particular for the
141 high-emitters.

142 The remainder of the paper is organised as follows. Section 2 discusses relevant
143 results from the existing academic literature. Section 3 describes the data and method
144 used to quantify emissions and investigate their drivers. Our results are presented in
145 Section 4, and discussed in relation to the existing literature. Section 6 concludes.

146 2 Study context

147 The empirical literature on drivers of mobility-related carbon footprints is broad. It
148 covers different geographical areas, methods, types of mobility behaviour and popu-
149 lation groups. Still, several variables and hypotheses are commonly discussed. This
150 section recaps some major insights and ongoing discussions to put our regression
151 results into perspective.

152 Regarding socio-economic characteristics higher mobility emissions are typically
153 associated with higher household income and education, as well as active labour force
154 participation. Geographical location has been identified as an important determinant
155 of mobility emissions. This is partly due to the built environment and infrastruc-

156 ture, partly due to self-selection. In particular, it is regularly found that living in
157 dense urban areas is associated with lower day-to-day travel emissions, but higher
158 long-distance travel emissions (Holden and Norland, 2005; Czepkiewicz et al., 2018;
159 Reichert et al., 2016). While density effects lower the emission intensity of moving
160 around an urban area, city-dwellers often have more dispersed social networks.

161 Other potential explanations for high emissions from long-distance trips include
162 ‘rebound effects’ from spending money that is saved by not owning a car (Ottelin
163 et al., 2014; Czepkiewicz et al., 2018), airport accessibility (Mattioli et al., 2021;
164 Bruderer Enzler, 2017; Kim and Mokhtarian, 2021), but also urban lifestyles and po-
165 tential self-selection: people may choose to live in city centers because they prefer to
166 commute less, or because they want better access to transportation infrastructures
167 such as airports to facilitate long-distance trips, rather than the other-way around
168 (Boarnet and Crane, 2001). Finally, the so-called ‘compensation hypothesis’ empha-
169 sizes the need for people living in dense urban areas to compensate for limited access
170 to green spaces (Holden and Norland, 2005).

171 With regard to demographic variables, household size can be expected to have
172 some scale effects when income is controlled for. Travelling with several people may
173 lower per capita mobility emissions. Brand and Preston (2010) find that age is only
174 relevant above a certain threshold, because of the absence of everyday commuting
175 for retired individuals, as well as a fewer long-distance trips, for instance for health
176 reasons. Previous transport research shows that generally mobility emissions are
177 higher for men than for women. This is also in line with time use studies, which
178 show that men spend a larger share of their time commuting and more leisure time
179 on out-of-home activities (Druckman et al., 2012) or that they spend more free time
180 on mobility (Smetschka et al., 2019).

181 Various studies have also taken psycho-social variables into account. Hunecke
182 et al. (2007) report that a positive attitude towards driving is associated with a
183 significant increase in the share of trips with private motorised travel modes. They
184 further find a positive, yet non-significant, relationship between pro-car attitudes and
185 GHG emissions. Several hypotheses could explain such results. A first explanation
186 relates to the modal choice: for a given trip, individuals who like to drive are more
187 likely to favor the car over other transportation modes. Another reason lies in the fact
188 that individuals who enjoy driving have a smaller disutility associated with long car
189 commutes (or even a positive utility, because they enjoy the travel itself, as suggested
190 by Ory and Mokhtarian (2005)), and thus are willing to accept longer commutes than
191 others.

192 However, the association between a positive attitude to specific means of transport
193 and mobility emissions with holiday travel seems to be weaker than with day-to-day
194 mobility (Prillwitz and Barr, 2011a; Böhler et al., 2006). This stands in contrast
195 to broader value systems, such as cosmopolitanism, which may be a more important
196 cause for long-distance travel (see e.g. Czepkiewicz et al., 2020; Kim and Mokhtarian,
197 2021) and multi-modal travel behavior Groth et al. (2021). In general, it is likely that
198 attitudes towards different modes of transportation reflect different lifestyles or even
199 deeper value systems.

200 Several studies suggest that sustainable practices at home are associated with

201 more emissions from long-distance trips. Again, several explanations lend them-
202 selves. First, such a result may be driven by groups with a specific mobility style that
203 combines sustainable daily travel with frequent flying. Such groups have for exam-
204 ple been termed ‘young travel-addicted urbanites’ (Magdolen et al., 2022), and their
205 long-distance travel behaviour explained by cosmopolitan attitudes and globalised
206 lifestyles (Czepkiewicz et al., 2019). Große et al. (2018) find that among Copenhagen
207 residents ‘the car still accounts for a considerable share (60%) of weekend trips in
208 Denmark among committed cyclists’. They also find that this group of ‘committed
209 cyclists’ undertakes holiday/weekend trips most frequently. People who like to bike
210 may take their bikes with them on long-distance trips and thus have to use a car.
211 This would be in line with a study by Aall et al. (2011), who identify more outdoor
212 recreation equipment as one of the reasons for more frequent use of cars on leisure
213 trips in Norway.

214 Second, monetary savings from less private mobility on a day-to-day basis could
215 lead to ‘rebound’ effects in long-distance travel (see e.g. Czepkiewicz et al., 2018;
216 Ottelin et al., 2014). Third, rebound effects can be fueled by mental accounting
217 techniques as well. Kaklamanou et al. (2015) find evidence of beliefs that sustainable
218 mobility practices at home can make up for unsustainable behaviors, such as flying
219 abroad (‘Compensatory Green Beliefs’).

220 Based on these previous results, we expect daily mobility emissions to be positively
221 associated with income, education, being male and in employment, living in a rural
222 area and having positive attitudes towards the car. We expect a negative correlation
223 of daily emissions with old age and a positive attitude towards biking. The impact of
224 attitudes on long-distance travel is more difficult to anticipate. If it is even relevant,
225 the effect could go in either direction.

226 3 Methods

227 3.1 Survey and data set

228 Our study draws from the 2017 German Mobility Survey, MiD, which took place
229 between May 2016 and October 2017. This survey is based on a two-stage inter-
230 view process of a representative sample of German households. It contains detailed
231 information about mobility behavior and access to transport modes, socio-economic
232 status, spatial information, as well as some attitude-related questions. The earlier
233 2008 version of the survey is used for instance in Reichert et al. (2016).

234 In a first phase, information about the household, such as household size, or
235 available transportation modes, is surveyed. In a second phase, all household members
236 over the age of ten are interviewed individually about their mobility behaviour. Given
237 that the random sampling is based on households, rather than individuals, larger
238 households are over-represented. Therefore the survey provides weights which correct
239 individuals’ observations to be representative of the general population. We use these
240 weights in all analyses that follow.

241 To avoid an overly long questionnaire with potentially low response rates, only a
242 share of the questions (‘modules’) is asked to each respondent. In particular, we are

243 interested in the following modules:

- 244 • Daily trips: reports all trips on the sample day (interviews within 14 day after
245 the sample day).
- 246 • Journeys: reports all journeys (with overnight stays) over the past three months.
- 247 • Attitudes towards transport modes.

248 While each respondent in the base sample is asked about daily trips and longer
249 journeys, detailed information related to attitudes is only available for a sub-sample.
250 Dropping individuals who did not fill out the attitude module, and cleaning the data
251 set (i.e. removing individuals who did not answer some of the questions), leaves us
252 with detailed mobility information for 11,713 individuals (from approximately 24k
253 initial observations). As the attitude module was not handed out at random, but
254 influenced by local contracting authorities, deleting observations on the basis of this
255 module runs the risk of biasing the sample. To address this concern, we perform a
256 Kolmogorov-Smirnov test to show that the base sample's characteristics do not differ
257 from our subsample (See Appendix D).

258 As with most mobility survey data, a concern is to define high- and low-emitters
259 based on a short reporting period. For daily trips only one specific sample day is
260 used, whereas the long-distance trips refer to a three-month period. While we assume
261 that over- and underrepresentation of emissions should even out over the full sample
262 size, there is a risk of confounding day-to-day variation of an individual with variation
263 between respondents. We control for this to the extent possible by using day- and
264 month-fixed effects.

265 3.2 Emissions calculation

266 **Emission factors.** We rely on emission factors from the Umweltbundesamt (UBA),
267 the German Agency for the Environment³ to convert mobility data into emissions. As
268 the most recent available emission factor for air travel from the UBA is solely based on
269 within-country flights, we use earlier estimates to compute emissions from air travel
270 in a more detailed manner (Mottschall and Bergmann, 2013). These emission factors
271 depend both on whether the flight is within or outside Germany, as well as the distance
272 traveled. Note that these estimates include non-CO2 effects by weighing emissions for
273 altitudes higher than 9000m by a specific factor. We acknowledge, however, that the
274 inclusion of non-CO2 effects of air travel is subject to large uncertainty (Lee et al.,
275 2010).

276 Since no estimates are available for motorcycles, trucks and on-demand bus, we
277 use the 2015 estimates from the Handbook Emission Factors for Road Transport
278 (HBEFA)⁴ instead. All emission factors used are reported in Table 1.

³Accessible on their website: https://www.umweltbundesamt.de/themen/verkehr-laerm/emissionsdaten#verkehrsmittelvergleich_personenverkehr

⁴Accessible on their website: <https://www.hbefa.net>

Transportation mode	Emission factors	Unit	Source
Car	220	gCO ₂ e/Vkm	UBA
Train (long-distance)	29	gCO ₂ e/Pkm	UBA
Train (short-distance)	55	gCO ₂ e/Pkm	UBA
Bus (long-distance)	29	gCO ₂ e/Pkm	UBA
Bus (short-distance)	80	gCO ₂ e/Pkm	UBA
Underground/tramway	55	gCO ₂ e/Pkm	UBA
Motorcycles	109	gCO ₂ e/Vkm	HBEFA
On-demand bus	228	gCO ₂ e/Vkm	HBEFA
Truck	815	gCO ₂ e/Vkm	HBEFA
Plane			
Within Germany (<500km)	246.7	gCO ₂ e/Pkm	UBA
Within Germany (>500km)	233.6	gCO ₂ e/Pkm	UBA
Outside Germany (<500km)	308.7	gCO ₂ e/Pkm	UBA
Outside Germany (500-1000km)	302.3	gCO ₂ e/Pkm	UBA
Outside Germany (1000-2000km)	241.3	gCO ₂ e/Pkm	UBA
Outside Germany (2000-5000km)	201.6	gCO ₂ e/Pkm	UBA
Outside Germany (5000-10000km)	229.0	gCO ₂ e/Pkm	UBA
Outside Germany (>10000km)	243.0	gCO ₂ e/Pkm	UBA

Table 1: Emission factors by transportation mode.
gCO₂e: grams of CO₂ equivalent. *Vkm*: vehicle kilometre. *Pkm*: person kilometre.

Relying on a single emission factor for road transportation has several limitations. First, we are not able to account for variation of emission intensity across private vehicles, which could be related to socio-economic characteristics of households. Second, it does not factor in that vehicle fuel consumption differs between road types, such as within-city streets and highways. While we acknowledge these limitations, relying on single emission factor is a common assumption, and can be considered acceptable as long as the heterogeneity between modes is greater than within modes.

Computing respondents' emissions. We restrict our analysis to adults, i.e. individuals above 18 years old. To convert the information about mobility into individual emissions we account for the number of travelers in the vehicle whenever private motorised travel modes are used.

For daily mobility, the interviewees are asked about the number of trips on the given day, but can only report details for a maximum of 8 to 12 trips. Similarly, for long-distance travels, only three journeys are reported comprehensively over the past three months, although the total number of journeys over that period may be higher. We account for these unreported trips by increasing emissions proportionately for each type of travel, assuming that the footprint of unreported travels would be the same as that of reported ones. One conceptual problem when distinguishing long- and short-distance travel with the MiD survey is that some trips without overnight stays can be longer than some of the journeys, which include overnight stays. We follow the conventional cut-off of 100km in the literature (Reichert et al., 2016) and exclude day trips that are further, and overnight stays that are less than 100km away. The

304 former reduces the sample size of daily trips from 960 to 880 thousand observations.
 305 The latter lowers the observations of long-distance trips from 39 to 35 thousand.

306 Finally, of all the work-related travels, we choose to keep commuting in the individ-
 307 uals' emissions, but exclude any other form of mobility for work on both a daily basis
 308 or long-distance trips in the main regression models. We do so, because first, such
 309 trips can be seen as more constrained choices compared to mobility for commuting or
 310 leisure purposes, especially when driving is part of the job. Second, business travel
 311 may create spill-over effects into leisure travel behaviour (see e.g. Cohen et al., 2018),
 312 which would cause endogeneity bias. Third, it is debatable whether work-related
 313 travel emissions should be accounted to consumption- or production-based emissions.
 314 For comparison, we present the results of our analysis when business travel is included
 315 as a sensitivity analysis.

316 Hence, for $k \in$ (daily trips, long-distance journeys), we can compute the two
 317 sources of annual emissions:

$$E_k = p_k m_k \sum_{j \in \text{trips}} e_j l_j \quad (1)$$

318 Where e_j is the emission factor of the vehicle used for trip j , l_j is a load factor (the
 319 inverse of the number of participants for private mobility, 1 otherwise), m_k is a scale
 320 factor accounting for unreported travels (equal to the number of actual trips divided
 321 by the number of trips reporting details on distance, travel mode, etc.), and p_k is a
 322 scale factor to get the annual emissions, accounting for the period over which trips
 323 are reported (equal to 365 for daily mobility, and to four for long-distance journeys).
 324

Hence, individuals' total annual emissions are

$$E_{total} = E_{daily} + E_{journeys} \quad (2)$$

325 3.3 Quantile regression

326 To understand how different individual characteristics contribute to explaining differ-
 327 ences in mobility emissions, we regress E_{total} on a number of predictors. To highlight
 328 differences between daily and long-distance mobility, we also perform a regression
 329 analysis with annual emissions from daily E_{daily} and long-distance travel $E_{journeys}$
 330 separately.

331 Given that outliers are common with mobility data and that the distribution of the
 332 target variable, mobility emissions, is likely different for low or high emitters, we chose
 333 a quantile regression (QR) strategy. QR is an extension of the linear regression model
 334 (OLS) whereby coefficients are allowed to vary along the distribution of the dependent
 335 variable (Koenker, 2017). While OLS minimizes the sum of squared residuals, QR
 336 minimizes a sum of absolute residuals Koenker and Hallock (2001). QR performs a
 337 local fitting of the data, segmenting the sample according to the outcome variable.
 338 The coefficients of the model are point estimates that can be interpreted as a one-unit
 339 change in y at quantile τ .

340 It is particularly interesting in the context of mobility emissions given their un-
 341 equal distribution, as it relaxes the assumption of a linear relationship between the

342 dependent variable and the predictors. Instead, it uncovers how this relationship
 343 varies along the distribution of the outcome variable.

344 This method differs from the one- (Büchs and Schnepf, 2013; Brand et al., 2013)
 345 or two-stage OLS regressions (Reichert et al., 2016) typically used in earlier stud-
 346 ies. With quantile regression, we can analyse how individuals' characteristics play a
 347 different role for moderate or high emitters, as in Bel and Rosell (2017).

348 We apply both OLS and quantile regression to highlight the difference in results.
 349 Rather than testing a specific theory, the study explores a large range of potential
 350 drivers of mobility emissions and quantifies their correlations with mobility emissions.
 351 The independent variables we consider in our model are household income, household
 352 size, age, employment, gender, location, migration background, ownership of a car,
 353 having a second home, and a car-sharing membership. In addition, we include atti-
 354 tudes of individuals towards different transportation modes. All regressions include
 355 time fixed-effects for the sample days and month of the interview. To avoid inflating
 356 the number of covariates, we distinguish between weekdays (Monday through Friday),
 357 Saturdays and Sundays, and group the interview months according to three distinct
 358 travel seasons. We categorise months as low, medium or high travel season based on
 359 the actual emissions from long-distance travel observed.

360 Similar to Bel and Rosell (2017) we define the quantile function as:

$$Q_{Y_i}(\tau) = \beta(\tau)X_i + \theta_i \quad (3)$$

361 where X_i is the regression matrix containing the above-mentioned covariates (e.g.
 362 age, income group, gender, etc.). This function estimates the coefficients at any
 363 quantile point τ (in our case the median, the 75th and 90th percentile). In principle,
 364 any number of quantiles can be chosen for a QR model. As our data set contains a
 365 large number of zeros, however, the lowest quantile to return an interpretable result
 366 is the median. This holds particularly for long-distance emissions. All QR and OLS
 367 models are estimated in R (Version Version 1.3.959) using the *quantreg* packages, as
 368 well as in STATA (Version 14 using functions *qreg* and *qreg2*). The code can be found
 369 at this link on Github.

370 We compare all QR results with a standard linear OLS regression model. A Vari-
 371 ance Inflation Factor (VIF) test is performed to test the OLS model of total emissions
 372 for potential multicollinearity. The resulting low values indicate no multicollinearity
 373 among the independent variables (See Appendix C for detailed results). Plotting the
 374 emission data reveals that residuals are non-normally distributed and heteroskedastic
 375 (see Appendix C). Hence we report heteroskedasticity-robust bootstrapped standard
 376 errors for all models (Koenker and Hallock, 2001). The patterns observed in Figure
 377 4 for a selected number of variables also suggest that the variance of the dependent
 378 variable increases with the level of the outcome variable.

379 3.4 Descriptive statistics

380 Analyzing daily and long-distance travel data for our sample of more than 11k in-
 381 dividuals, we find that overall, private vehicles contribute to the vast majority of
 382 total emissions, as they represent a large share of daily mobility. However, air travel
 383 dominates emissions from long-distance travels (see Figure 1).

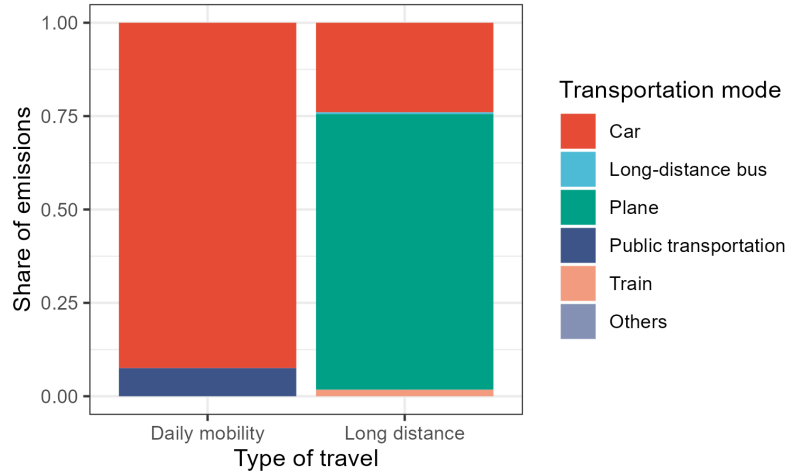


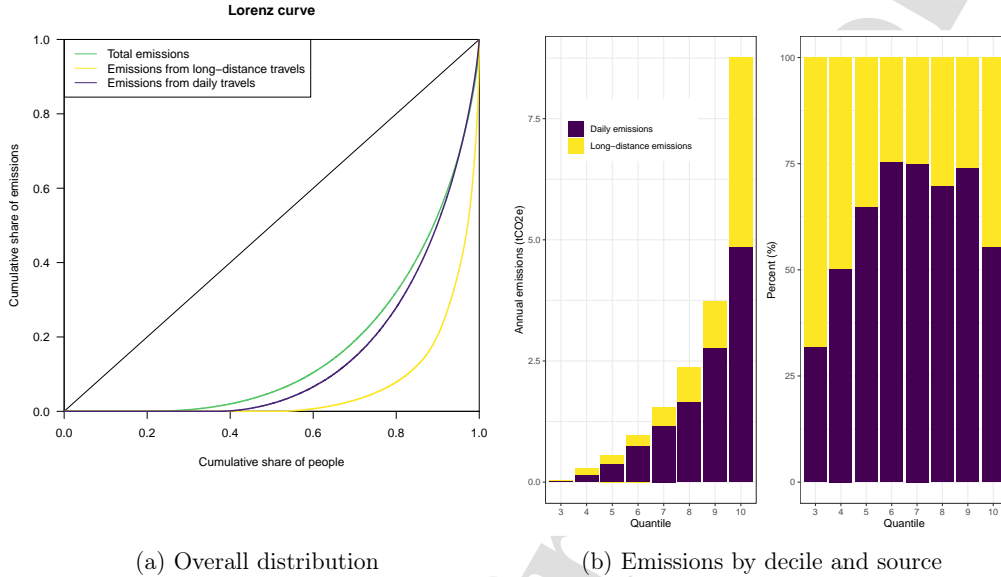
Figure 1: Emission contribution by transport mode and type of travels.

384 In line with previous studies, we find that emissions are unequally distributed
 385 among individuals. Computing the Lorenz curves for emissions from total, daily and
 386 long-distance mobility (see Figure 2a), we find the top 10% emitters contribute to
 387 51% of total emissions, and as much as 80% of emissions from long-distance travel,
 388 respectively. The Gini coefficients are respectively 0.67 for total emissions, 0.71 for
 389 emissions from daily mobility and 0.88 for emissions from long-distance travels⁵. This
 390 indicates that individuals emitting through long-distance trips are not the same ones
 391 who emit most in their day-to-day mobility, so both types of mobility contribute
 392 to making individuals belong to the high emitters group. This further justifies the
 393 decision to look into drivers of long-distance versus daily mobility.

394 It should be highlighted that there is a very high number of zero mobility emis-
 395 sions. Given that information on daily trips is based on a single sample day, the
 396 regression sample contains two types of zeros: those who traveled only by foot or
 397 bike, and those who did not travel at all on the sample day. Based on this distribu-
 398 tion of emissions, we focus on estimating the 50th, 75th and 90th percentiles. Their
 399 coefficients should be interpreted as the marginal effect at the respective quantile of
 400 emissions of the whole representative sample, rather than as quantiles of emissions
 401 from all travellers (as there are non-travellers in the sample).

402 Figure 2 presents the distribution of absolute emissions across quantiles on the left
 403 and the share of mobility emissions on the right, both by type of travel: daily versus
 404 long-distance trips. It shows that individuals belonging to the highest emission decile
 405 emit on average 11 tCO₂e per year, which is about ten times more than those at the

⁵It should be noted that the single sample day for daily travel is less reliable than the three-month period considered for long-distance travel. If sampling inflates the share of zeros in daily travel, this will inflate the inequality measure compared to long-distance travel.



(a) Overall distribution

(b) Emissions by decile and source

Figure 2: Distribution of mobility emissions

406 median.

407 Appendix D presents the summary statistics of the main covariates used in our
 408 regression analysis.

409 4 Results

410 Given that we are interested in the impact of attitudes in particular, we first per-
 411 form the regression model without any attitudinal variables and then add attitudes
 412 as additional control variables to investigate their effect on the other coefficients. An
 413 overview of the of the OLS and quantile regression results including attitudinal vari-
 414 ables is shown in Figure 3. The detailed tables can be found in Appendix E (results of
 415 the same models without attitudes in Appendix F.3). We will first discuss the role of
 416 various socio-economic and demographic factors in Section 4.1. We then present the
 417 correlation of attitudes towards different transport modes (Section 4.4) with mobility
 418 emissions. All variable labels and reference categories are detailed in Appendix A.



Figure 3: Regression coefficients for different types of mobility emissions

4.1 Socio-economic factors

Income. In line with previous research, we find that higher household income is strongly and positively associated with mobility emissions, both for total and long-distance trips. Being within the highest income group increases total annual mobility emissions by 1.6 tons of CO₂e (tCO₂e) at the 75th percentile, and by as much as 2.9 tCO₂e at the 90th percentile of emitters. However the contribution of income to emissions seems to saturate, especially for long-distance emissions, as the coefficients for the highest income group is not much higher than the one for the second highest. This result stands in contrast to findings by Büchs and Schnepf (2013) that annual transport emissions increase strongest above a household income of approximately GBP 8000. One possible explanation for this difference could be the emissions calculation. Since Büchs and Schnepf (2013) rely on fuel expenditure their emissions capture distances travelled as well as emission-intensities of different vehicles. Combining this with our results would indicate that the highest income group does not necessarily travel longer distances, but that they are travelling more fuel intensively. When we look at day-to-day travel specifically, income only contributes to emissions at the 75th percentile of emitters. Emissions increase disproportionately strong with income at the upper end of the distribution, meaning that it is a more important contributor among high daily emitters.

When attitudes are not controlled for, some income coefficients are underestimated and less significant. This concerns the effect of being in the medium household income

group (2000-4000e/m) on daily emissions and the impact of being in the moderate income group (900-2000e/m) on long-distance emissions. Being in these income groups appear to intersect with attitudes that correlate with lower mobility emissions.

Education. Similar to income, education - and especially tertiary education - is associated with significantly higher mobility emissions. Again, the effect is significant across all quantiles, with increasing effect size for high emitters. The coefficient for tertiary education is more than six times as high at Q90 as at Q50 in the case of total emissions. This pattern is driven by long-distance emissions, while daily emissions are not significantly different across education groups.

Employment. Consistent with Ko et al. (2011), we find that being in employment is a significant contributor to total mobility emissions, with low variation across quantiles. This result translates to daily emissions. While Brand and Preston (2010) find that the effect of employment is not significant for long-distance mobility emissions, we can only confirm this through the OLS model. The quantile regression results suggest that being employed increases long-distance emissions significantly. Note that this is despite the exclusion of business travel here.

Second home In the main regression model living in a second home part of the time significantly increases long-distance emissions for all emitter quantiles, while the OLS regression coefficient is insignificant at the pre-specified levels. When we include business travel (see sensitivity analysis and Appendix 16), the effect is larger and the OLS coefficient also becomes significant. It should be noted that access to a second home can have numerous different explanations. People may have a second home where they spend holidays or weekends, but also a partner or work place in a location relatively far from home. Such multi-locality, a lifestyle where a person lives in more than one place with regular or irregular presences and absences, can arise from mobility needs for the purpose of work or personal relationships and can impact mobility behaviour in a sustained way. There is a growing body of research on such multi-modal behaviours in Germany (Danielzyk et al., 2020; Greinke and Lange, 2022).

4.2 Demographic factors

Household size. Larger household size is associated with lower per capita mobility emissions, probably due to scale effects, such as ride sharing. Besides, different household sizes can reflect different lifestyles. Singles may spend more time out of their home and with friends, whereas couples, especially with children, might centre more around their home, having to transport furniture, or they travel together. Büchs and Schnepf (2013), for instance, find that in the UK transport emissions of two-adult households are three times higher than for single adult households on average. At the same time, it should be noted that dividing travel emissions by passenger numbers attributes part of the mobility emissions to children. Under the assumption that adults are making the travel choices, this can lead to an underestimation of per capita emissions from a responsibility perspective.

Age. Total mobility emissions decrease with age. Both daily and long-distance emissions drop significantly for the age groups above 60 years. This result is in line

483 with Brand and Preston (2010), but can only partly be attributed to the absence of
 484 everyday commuting, because the commuting effect of many participants is already
 485 captured in the positive and significant coefficient of being employed. Yet, some peo-
 486 ple commute, of course, even though they are not formally employed (e.g. students).
 487 Probably older people also move around less for errands and leisure activities.

488 Total mobility emissions are also lower in the age group 40-45 years, driven by
 489 differences at the median emitters. The negative effect of age on long-distance emis-
 490 sions only holds for the OLS regression model. The quantile model finds no evidence
 491 that older people create fewer long-distance emissions. When we exclude attitudes
 492 from the regression model the coefficient for being in the age group 45-59 on daily
 493 mobility emissions turns significant.

494 **Gender.** Long-distance emissions differ by gender only at the lower end of the
 495 distribution. In this group, being female is associated with higher emissions. On the
 496 other hand, women produce significantly less daily (and overall) mobility emissions.
 497 This gender gap is particularly pronounced within the top 25% of emitters. The
 498 effect of being female on total emissions is overestimated in size and significance
 499 when attitudes are not controlled for.

500 **Migration background.** Having a migration background is negatively corre-
 501 lated with daily mobility emissions, but it has no significant effect on long-distance
 502 emissions. The resulting total mobility emissions do not show a significant difference
 503 for those with a migration background.

504 **Car sharing.** Unsurprisingly, car sharing is especially relevant for day-to-day
 505 emissions. It significantly reduces this type of mobility emissions across all estimated
 506 models. For long-distance emissions car sharing has a significant positive effect at the
 507 lower emitter percentiles, but not among the highest emitters. Car sharing remains a
 508 marginal phenomenon in Germany, with only 4.8% of respondents in our regression
 509 sample reporting a car sharing membership. It is also restricted mostly to young,
 510 male, urban dwellers Nobis and Kuhnimhof (2018). This mode is typically used for
 511 shopping or errands, often on the weekends and either for very short trips (<2km)
 512 or for trips between 15 and 30km. Since about half of the car sharing members do
 513 not own a car, and even those who own a car travel disproportionately much by bike
 514 and public transport, we suspect that any single trip done with a shared car adds a
 515 comparatively large amount of emissions to their carbon footprint.

516 However, the positive link with long-distance emissions may not arise directly
 517 from the use of car sharing, but rather reflect lifestyle characteristics that are not
 518 captured by the other variables in our model. Recent research on German cities
 519 could provide an explanation for the link between certain social milieus and sharing
 520 behaviours. Findings by Groth et al. (2023) suggest that the supply of new mobility
 521 services, such as car- or bike sharing, is concentrated on socio-spatially exclusive areas
 522 in Global Cities, providing disproportionate access for economic and cultural elites.

523 4.3 Spatial factors and modal access

524 **Location.** In line with the existing literature, the OLS results suggest that on average
 525 urbanites emit more for long-distance trips and less for daily and overall mobility.

526 However, the significance levels vary strongly between groups. For total emissions,
 527 only the OLS result is significant. When the outcome variable is daily emissions,
 528 any sort of agglomeration shows a significant negative effect on emissions, compared
 529 to the baseline (rural area). Long-distance emissions are higher for dwellers of all
 530 agglomerations, with the exception of middle-sized cities. An interesting result is
 531 that living in a smaller urban area increases long-distance emissions in the high-
 532 emitter group, whereas the effect of living in a metropolis is more relevant for the
 533 lower emitter quantiles.

534 Comparing location coefficients with and without controlling for attitudes reveals
 535 the correlation between where people live and their psychological characteristics. The
 536 negative effect of living in a big city on total emissions, for instance, is overestimated
 537 when attitudes are not controlled for. Similarly, part of the lower daily emissions
 538 of city- and metropolis-dwellers seems to be explained by attitudes, rather than the
 539 built environment alone. This possibly supports the self-selection hypothesis.

540 **Car ownership.** Finally, we find that car ownership has a strong and positive
 541 effect on all emission outcomes we model: total, daily and long-distance emissions.
 542 This appears to be in contrast to results by Ottelin et al. (2014), who find that
 543 reduced driving in dense urban areas rebounds as higher flight emissions. However,
 544 their study has a more narrow focus on middle-income households in the Helsinki
 545 metropolitan region, while we use data from a representative national sample, which
 546 includes various urban forms and income levels. Furthermore, there might be cultural
 547 and geographical differences at play. The car has a dominant role in German mobility
 548 behaviour, including in long-distance travel. Almost one quarter of all long-distance
 549 emissions in our sample is caused by car travel. However, the effect of car ownership
 550 in terms of coefficient size is larger for daily (and overall) mobility emissions. The
 551 latter is also more affected by attitudes towards different transport modes. Without
 552 accounting for attitudes, the model overestimates the effect size of car ownership on
 553 total and daily mobility emissions.

554 4.4 Attitudes towards transportation modes

555 Our results show that attitudes towards various transportation modes (enjoyment of
 556 biking, driving a car and taking public transport) can be relevant to understand the
 557 heterogeneity in personal mobility footprints.

558 **Enjoying to drive.** Individuals who enjoy driving a car tend to have higher
 559 levels of daily mobility emissions. This is in line with findings from Hunecke et al.
 560 (2007). Even moderately disliking to drive (compared to fully rejecting the statement
 561 that one enjoys to drive) is associated with significantly higher daily emissions. It is
 562 worth noting that the size of these effects is large: enjoying very much to drive a car
 563 is associated on average with an additional 0.68 tCO₂ of total mobility emissions on
 564 average (0.79 tCO₂ for daily emissions), which is greater than the median emissions,
 565 and almost on par with the effect of tertiary education.

566 While the effect of enjoying to drive is strong and significant for daily emissions,
 567 total mobility emissions are only significantly higher for respondents who fully enjoy
 568 driving. This effect holds across all quantiles. The association with long-distance

569 emissions is less clear. This coincides with insights from other studies about a weaker
570 association of attitudes towards specific transport means with holiday travel, com-
571 pared to day-to-day mobility (Prillwitz and Barr, 2011b; Böhler et al., 2006). It stands
572 in contrast to other types of attitudes or values, especially lifestyle-related attitudes,
573 such as cosmopolitanism, which may be more closely related to long-distance travel
574 (see e.g. Czepkiewicz et al., 2020; Kim and Mokhtarian, 2021).

575 **Enjoying public transportation.** As far as public transportation is concerned,
576 the OLS regression suggests that agreeing with the statement "I enjoy public trans-
577 portation" is associated with lower total and daily emissions, compared to partici-
578 pants who fully disagree with this statement. The QR results reveal that this effect
579 is driven by the high-emitters alone. Strongly agreeing with this statement only has
580 a significant effect on daily mobility emissions.

581 **Enjoying to bike.** Attitudes about biking also affect emissions. Moderate atti-
582 tudes towards biking are associated with higher daily and total mobility emissions.
583 Especially for the moderately positive attitude, this result is counter-intuitive. One
584 would expect that people who enjoy biking are more likely to use this low-carbon
585 means of transportation for daily mobility. However, it seems that if this should be
586 the case, attitudes towards biking are associated with other unobserved characteris-
587 tics, which foster mobility emissions. The results tend to be more relevant (effect size
588 and significance) towards the upper end of the distribution.

589 Long distance emissions are higher for participants who enjoy biking. This effect
590 is not straightforward. As mentioned above, opinions about biking potentially over-
591 lap with other characteristics, which are themselves linked to higher emissions. An
592 interesting result is that for long-distance travels, strongly positive attitudes towards
593 biking are associated with higher emissions for the bottom 75% of emitters. As already
594 discussed in Section 2, there may be different explanations for this. First, enjoying to
595 bike could correlate with enjoying to travel because both are indicators for an active
596 life. The finding may at least in part represent a group of 'multimodal' young people,
597 who tend to belong to a 'cosmopolitan' social milieu (Groth et al., 2021), a lifestyle
598 typically connected with above average air travel (Czepkiewicz et al., 2020). A second
599 explanation could be some form of compensation behaviour, for instance in the form of
600 'moral licensing', i.e. a justification of emission-intensive long-distance trips through
601 pro-environmental behaviour (biking) at home. Indeed, the effect for a positive bike
602 attitude vanishes when we control for the frequency of bike and public transportation
603 use (see Appendix 9). Still, among the lower 50% of emitters, using a bike regularly
604 is associated with higher long-distance emissions. Last but not least, assuming that
605 people who like to bike often take their bikes with them on long-distance trips, they
606 may may find it more convenient to use a car.

607 To summarise, the most important contributing characteristics (coefficients of
608 more than 1 tCO_{2e}) of high day-to-day emitters are enjoying to drive and a high
609 household income (>6000€/month). Living in a three-person household (compared
610 to a single person household) and being older in the age group 60-64 years show
611 the largest significant negative coefficients showing that certain aspects of culture
612 and lifestyle, as well as mobility needs, are just as relevant in determining mobility
613 emissions as are economic factors. It should be noted, that the weekday, and Sunday

614 in particular, has a large significant impact on short-distance trips as well.

615 Regarding long-distance emissions, the most important drivers among the top 10%
616 emitters are by far second home access, high household income (>4000€/month) and
617 tertiary education. Also significant, but with an effect size almost a magnitude smaller
618 are car ownership, being employed and living in an urban environment. Interestingly,
619 attitudes towards biking and driving are more relevant for long-distance emissions
620 of the lower emitter groups. Again, the timing of the survey is highly relevant with
621 significantly higher long-distance emissions during typical travel months.

622 Comparing the OLS model with the QR model, the direction of the effects observed
623 are similar. However, the QR model allows for a better understanding which
624 categories of emitters are driving the results. It also enables us to highlight the
625 association of the independent variables with emissions for high emitters in particular.
626 Figure 4 shows the OLS (red) and QR (black) estimates for selected variables.
627 The dashed red lines represent the OLS confidence intervals, whereas the grey-shaded
628 area is the equivalent for the QR. Except for the age groups 25-40 and 60-64, some
629 of the quantile regression coefficients always fall outside of the OLS confidence inter-
630 vals. This suggests that the effects of these covariates may not be constant across the
631 conditional distribution of total mobility emissions.

632 For OLS models we compute R squared and for quantile regressions pseudo R
633 squared. This shows that the OLS regressions explain approximately 10% of variation
634 in total emissions, long-distance travels and daily mobility, but only 3% for emissions
635 from flights. Relatively low squared R are common when studying transportation
636 behaviors, because of the diversity of omitted variables potentially contributing to
637 mobility patterns and thus emissions. Low values are also due to a high number of
638 both zero-value and extreme observations owing to limited time coverage of mobility
639 (Stewart, 2018). Finally, we report the Machado-Santos-Silva test for heteroscedas-
640 ticity.

641 5 Sensitivity analysis and additional models

642 The choice of covariates and outcome variables in the main regression model is se-
643 lective. This section tests whether and how other control variables and a different
644 specification of the outcome variable affect the results. In particular, we test the sen-
645 sitivity of our results with respect to satisfaction with and frequency of use of various
646 transport modes as additional controls. We also test the model for emissions from air
647 travel specifically to identify drivers of high air travel emitters compared to overall
648 long-distance emitters. Lastly, we test how results change when mobility emissions
649 include business travel. Regression results including other spatial variables can be
650 found in the supplementary materials.

651 5.1 Additional control variables

652 **Satisfaction with travel modes.** Besides enjoyment of different transportation
653 modes, the MiD survey asks participants about their satisfaction with different trans-
654 portation modes. In particular, participants are asked to state how they rate the

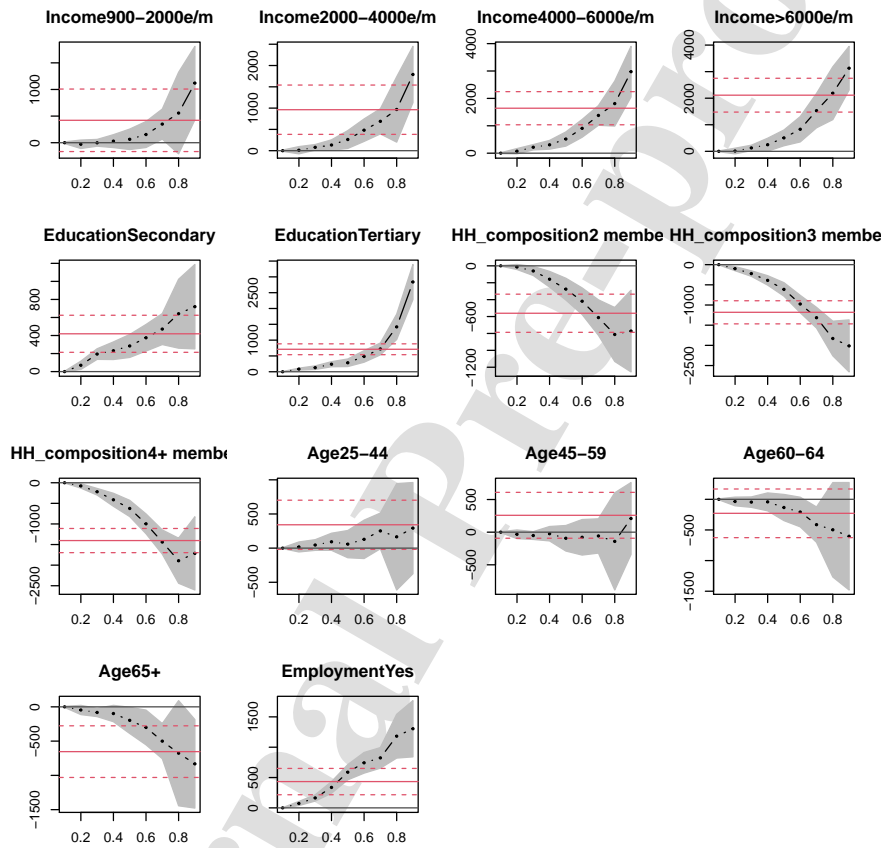


Figure 4: Quantile versus OLS regression

Note: Values of QR regression coefficients along the distribution of the dependent variable (dotted black line) and associated confidence interval for a selected number of independent variables. The red lines represent the point estimate and confidence interval for the OLS model.

655 general traffic situation in their area for the respective transportation means. Such
 656 satisfaction with different modes could be argued to be a precondition for enjoyment,
 657 so we expect some overlap in the effect of both variables. Indeed, when we add satisfac-
 658 tion to the model for total mobility emissions, most coefficients are not significant
 659 at the specified levels (see Appendix F.2, table 11 for detailed results).

660 The only coefficients that are significant are those from satisfaction with public
 661 transport. An increase in satisfaction with public transportation tends to be asso-
 662 ciated with an increase in emissions, this being driven mostly by the high emitter
 663 category. A potential explanation for this surprising result could be that using public
 664 transportation often is linked to lower emissions, but also to a more negative percep-
 665 tion of this transportation mode.

666 **Use of sustainable transport modes.** As mentioned in the previous sections,
 667 one hypothesis is that more frequent use of sustainable transport modes in the day-
 668 to-day life of an individual may contribute to higher long-distance emissions due to
 669 various potential compensation mechanisms. To investigate this issue a little closer,
 670 we perform a regression including the frequency of bike and public transportation use
 671 (see Appendix F.1).

672 Those who cycle daily produce less daily mobility emissions over the whole dis-
 673 tribution. When it comes to spillover effects to long-distance travel, daily cycling is
 674 associated with significantly higher long-distance emissions. The effect is also notice-
 675 able for at the lower quantile for people who bike less than 3 times a month. It should
 676 be highlighted, though, that the reduction in daily emissions outweighs the increase
 677 in long-distance emissions by a factor of 2.

678 Regarding the use of public transportation, only daily use of public transporta-
 679 tion is associated with a reduction in daily emissions. Using public transportation
 680 occasionally, on the other hand, is associated with higher long-distance emissions
 681 compared to the reference category (people who never use it).

682 5.2 Emissions from air travel

683 Long-distance travel emissions are dominated by emissions from air travel and are
 684 partly driven by superflyers (Gössling and Humpe, 2020). Table 2 shows the regression
 685 results for emissions from air travel only. Since emissions from air travel are even
 686 more concentrated, i.e. there are no air travel emissions at all for the bottom half
 687 of emitters (see Appendix B), we choose different quantiles here than for the main
 688 regression model (namely 90th, 95th and 99th percentile).

689 Overall, the OLS regression shows that income, education, household composition,
 690 age, migration background and attitudes toward different transportations modes have a
 691 significant effect on emissions from flying. High income plays an even more important
 692 role for air travel than for overall long-distance travel emissions. A household size of
 693 three or more people has a significant negative effect on air travel emissions, indicating
 694 that "superflyers" tend to live in small households.

695 Perhaps the most interesting result is the influence of attitudes towards driving
 696 a car. While our results for total emissions showed that enjoying to drive increases
 697 total mobility emissions - especially in the lower emitter groups-, even a moderately

698 positive attitude towards the car is associated with a 0.5 tCO₂e reduction in air
 699 travel emissions, the effect being highly concentrated the top emitters. Conversely, a
 700 positive attitude towards biking is associated with an increase of emissions linked to
 701 air travel.

702 Dropping the control variables for attitudes in the regression model does not affect
 703 significantly the coefficients for air travel (see in Appendix table 15).

Table 2: Regression for emissions from air travel.

	<i>OLS</i>	<i>quantile regression</i>		
		90th	(95th)	(99th)
Constant	442.6** (211.8)	372.0 (1, 191.2)	1,875.6 (4, 812.0)	14,105.2* (7, 302.3)
Month2_medium	196.8*** (54.9)	348.5** (164.7)	489.2 (300.5)	1,897.1** (940.6)
Month3_high	93.2 (61.4)	330.2* (182.5)	195.5 (282.5)	132.4 (617.4)
Weekday2_Saturday	-164.5** (68.7)	-98.7 (185.5)	-416.6 (262.3)	-1,279.6** (595.4)
Weekday3_Sunday	-183.3*** (64.1)	-248.9 (188.6)	-558.3* (310.4)	-769.6 (1,076.2)
Income900-2000e/m	39.5 (150.0)	114.0 (886.2)	660.9 (1,580.6)	608.5 (1,142.5)
Income2000-4000e/m	200.8 (150.9)	454.4 (918.4)	1,172.4 (1,597.9)	2,872.9** (1,310.4)
Income4000-6000e/m	468.5*** (160.6)	996.0 (935.1)	1,698.5 (1,636.1)	4,870.8*** (1,655.0)
Income>6000e/m	390.2** (177.4)	1,067.1 (948.2)	1,751.4 (1,676.8)	4,161.2** (1,724.4)
EducationSecondary	-108.2 (66.0)	-145.8 (209.8)	-145.6 (253.4)	-371.4 (730.6)
EducationTertiary	152.8*** (55.3)	-121.0 (206.5)	911.7 (620.6)	4,565.2*** (1,200.2)
HH_composition2 members	24.7 (69.8)	527.3** (241.2)	503.0 (531.8)	69.4 (1,188.0)
HH_composition3 members	-338.5*** (84.9)	-419.7* (253.8)	-755.7 (558.8)	-2,231.2** (1,013.8)
HH_composition4+ members	-358.3*** (86.9)	-569.2** (277.4)	-973.7 (652.7)	-2,131.6* (1,249.7)
Age25-44	80.5 (93.7)	263.8 (310.9)	124.0 (655.2)	459.0 (1,106.6)
Age45-59	-125.7 (97.1)	7.1 (345.9)	-189.7 (621.4)	90.8 (1,009.0)
Age60-64	-238.6** (121.4)	-84.8 (453.4)	-572.7 (684.1)	-580.6 (1,172.9)
Age65+	-251.2** (111.4)	-145.6 (354.5)	-615.8 (725.6)	-1,474.6 (1,169.8)
EmploymentYes	-72.2 (66.4)	221.2 (215.5)	-278.1 (399.7)	-1,227.1 (947.2)
GenderWoman	68.6 (46.1)	56.2 (146.7)	-149.2 (271.4)	608.5 (748.7)
LocationUrban environment	8.4 (68.8)	166.0 (255.2)	59.0 (479.8)	3,610.1** (1,561.7)
LocationMiddle city	-126.9* (68.2)	49.9 (181.7)	60.0 (399.0)	270.4 (908.8)
LocationBig city	115.2 (80.6)	608.9** (310.3)	667.7 (761.3)	1,672.6 (1,769.5)
LocationMetropole	-42.8 (76.5)	225.6 (189.2)	161.5 (334.8)	93.4 (879.2)
Migration_backgroundYes	388.6*** (89.9)	1,090.2*** (387.2)	1,074.6 (910.6)	3,088.0 (2,477.1)
Enjoy_BikingDisagree	21.5 (82.3)	-24.1 (215.8)	349.1 (455.8)	1,279.6 (879.2)
Enjoy_BikingAgree	153.9* (79.8)	127.2 (259.4)	865.9** (414.1)	2,556.4*** (970.8)
Enjoy_BikingFully agree	34.3 (82.4)	-203.9 (195.5)	200.8 (386.8)	869.3 (924.9)
Enjoy_CarDisagree	-512.5*** (112.3)	-535.6 (763.2)	-1,228.7 (4,447.5)	-14,246.3** (7,109.8)
Enjoy_CarAgree	-459.0*** (108.1)	-510.5 (761.3)	-1,120.3 (4,463.6)	-13,368.0* (7,036.2)
Enjoy_CarFully agree	-364.4*** (112.6)	-311.3 (758.2)	-608.1 (4,447.7)	-11,458.1 (6,997.0)
Enjoy_PublicTransportDisagree	-48.0 (58.9)	-5.7 (243.5)	-372.2 (444.2)	-699.8 (1,067.8)
Enjoy_PublicTransportAgree	-81.1 (68.0)	-30.9 (233.3)	-583.6 (411.6)	-808.9 (995.5)
Enjoy_PublicTransportFully agree	-142.6 (100.3)	-203.2 (264.5)	-626.0 (540.7)	-190.3 (1,193.5)
Car_ownershipYes	283.8*** (84.5)	252.3 (177.9)	184.3 (418.0)	844.3 (855.5)
Second_HomeYes	-69.5 (110.1)	213.8 (323.6)	-627.6 (504.6)	-1,201.7 (1,310.0)
Car_sharingYes	82.5 (104.7)	59.9 (253.2)	188.7 (1,603.9)	72.9 (1,338.2)
Observations	6,292	6,292	6,292	6,292
(pseudo) R ²	0.03	0.04	0.03	0.12
Machado-Santos-Silva test		41.1	61.5	35.9

Note:

*p<0.1; **p<0.05; ***p<0.01

5.3 Emissions from business travel

We excluded business travel from the main regression model, based on three arguments. First, the decision about trip frequency or means of transportation may be more constrained, as compared to private trips. Second, it is unclear whether such emissions should be attributed to the individual traveling, or added to the emissions for the production of the good or service for which the trip is made, and hence attributed to the consumer household. Third, business travel may create spill-over effects into leisure travel behaviour (see e.g. Cohen et al., 2018), which could cause endogeneity bias. Yet, checking whether the main contributing factors differ when we include business trips may be an interesting base for discussion. The detailed results across quantiles can be found in Appendices 18, 16 and 19.

Total emissions. Most associations are similar for overall mobility emissions, independent of whether business travel is included. The size effect of having a high income tends to be greater when including business travels. Older age and household composition tend to be similar, whether business travels are accounted for or not. However, there are two variables which stand out: employment status and gender. While the effect of being employed is about 50% higher when we include business trips, the negative coefficient for being a woman increases more than threefold.

Long-distance emissions. The pattern looks rather similar for long-distance emissions. Income tends to have a greater effect on emissions once we include business emissions, and we observe the same differences for gender and unemployment, gender playing a greater role in the regression with business emissions while unemployment becomes significant. Having a second home also becomes significant, and has a strong effect once we include business trips. The causality for the latter likely runs in the opposite direction, meaning that those who often travel to a specific place on business are more likely to maintain a second home there. Conversely, car ownership, becomes insignificant when we account for business emissions. We also observe some changes in the effect of attitude towards transportation modes.

Daily emissions. For daily mobility, most coefficients are fairly similar when including business travels.

6 Conclusions

This article studies the greenhouse gas emissions from individual mobility in Germany. Using a recent mobility survey, we document the unequal contribution of individuals to emissions. We show that the top 10% of emitters contribute respectively to 51% of total mobility-related emissions, and to 80% of long-distance travel emissions. This stresses the importance of better understanding the drivers of these high-emitters for effective emission mitigation policies in the transport sector.

We thus analyse the drivers of mobility emissions for different categories of emitters. Performing a quantile regression allows us to highlight how socio-economic, demographic and attitudinal drivers affect total mobility emissions as well as emissions from daily versus long-distance travels. Our results confirm findings from the literature that higher income and education both correlate positively with emissions

746 from all types of mobility. However, the effect of income seems to saturate at some
747 point, and the coefficients for the two highest income groups are similar in magnitude.
748 In addition, the results also reveal that the relevance of various drivers differs across
749 emitter groups and depends on the type of travel performed.

750 The most important drivers for day-to-day mobility emissions are a high household
751 income, enjoying to drive and car ownership, as well as tertiary education. Each of
752 these characteristics is associated with additional emissions of more than 1 tCO₂e.
753 Long-distance travel emissions are dominated by high incomes and tertiary education.
754 Living in a city is found to lead to fewer emissions overall, an effect that is driven
755 by the decrease in daily emissions, and outweighs the countervailing increase in long-
756 distance trips.

757 Besides socio-economic and demographic variables, we also analyse the role of at-
758 titudes towards different transportation modes. In terms of coefficient size, a positive
759 attitude regarding the car is particularly relevant for total emissions. This effect is
760 mainly driven by the clear association between attitude towards the car and daily
761 emissions, suggesting an important role of preferences on the modal choice and dis-
762 tance traveled for commuting and daily leisure. At the same time, we find that
763 enjoying to drive a car is associated with lower air travel emissions among 'super-
764 flyers' (top 1%), suggesting some substitution of even more polluting long-distance
765 travel by car in our sample.

766 Attitudes towards more sustainable transportation modes are also significantly
767 associated with differences in mobility emissions. A positive attitude towards public
768 transport, for instance, is associated with lower daily emissions, particularly in higher
769 emitter groups. Enjoying to bike is associated with higher daily emissions among high
770 emitters and with higher long-distance emissions among lower emitter groups.

771 While we establish robust links between mobility emissions and different socio-
772 economic and attitudinal characteristics of individuals, more research is needed to
773 uncover the causal mechanisms behind these associations. An important limitation
774 with respect to the data set is the short reporting period for daily emissions (one
775 sample day). Regarding the role of psychological factors in mobility decisions, we are
776 limited by data availability and thus solely focus on attitudes to different transport
777 modes. Hence our study is by no means exhaustive on the psychological side. While
778 attitudes have been shown to be important - and we found significant effects - one
779 might argue that broader values or worldviews should be considered as well. A serious
780 criticism of attitudes specifically has recently been raised by (van Wee et al., 2019) and
781 (Moody and Zhao, 2020). They stress the importance of the built environment and
782 travel behaviours on the formation of attitudes (Mattauch et al., 2016), which raises
783 concerns regarding reverse causality (in our case from travel emissions to attitudes).
784 This is a topic that needs further investigation before causal inferences can be drawn.

785 All in all, our results point to strong heterogeneity in the effect of socio-economic
786 and attitudinal drivers on emissions, which need to be considered to design climate
787 policies. For instance, knowing that enjoyment of distinct transport modes is relevant
788 at different points of the emitter distribution can help better target information poli-
789 cies or infrastructure development. Focusing on high-emitters could significantly in-
790 crease abatement potential, given the share of emissions they represent. Our findings

791 can further be used to better understand the distributional effects of transportation
792 policies. The importance of drivers beyond income suggest that these policies could
793 have strong effects on horizontal equity.

794 **Code availability**

795 We performed the analysis and built figures using R and Stata. The code is available
796 on Github at the following link: https://github.com/ntaconet/Mobility_Germany/.
797 The German Household Mobility survey is not publicly accessible but can be obtained
798 from the Bundesministeriums für Verkehr und digitale Infrastruktur (BMVI).

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- The study examines the distribution of emissions from mobility in Germany and identifies the drivers of these emissions, particularly among high emitters.
- The top 10% of emitters in Germany are found to be responsible for 51% of total emissions from mobility, highlighting the unequal distribution of emissions.
- Emissions from long-distance travel are particularly concentrated among high emitters, with the top 10% accounting for 80% of emissions in this category.
- Socio-economic factors, such as income and location, significantly influence emissions from mobility across different levels of emitters.
- Attitudes towards different transportation modes are shown to strongly correlate with variations in mobility behaviors and emissions, underscoring the importance of behavioral factors.