ORIGINAL PAPER

Evolution of food protection levels and food vulnerability in Europe since 1950 estimated with vine‑copula models

Dominik Paprotny¹[®] • Cornelis Marcel Pieter 't Hart² • Oswaldo Morales-Nápoles²

Received: 3 April 2024 / Accepted: 12 November 2024 © The Author(s) 2024

Abstract

The magnitude of food impacts is regulated not only by hydrometeorological hazard and exposure, but also food protection levels (primarily from structural food defenses) and vulnerability (relative loss at given intensity of hazard). Here, we infer the variation of protection levels and vulnerability from data on historical riverine, coastal, and compound foods and associated impacts obtained from the HANZE database, in 42 European countries over the period 1950–2020. We contrast actual damaging foods, which imply food protection was locally inadequate, with modelled potential foods, i.e. events that were hydrologically extreme but did not lead to signifcant impacts, which imply that food protection was sufficient to prevent losses. Further, we compare the reported magnitude of impacts (fatalities, population afected, and economic losses) with potential impacts computed with depth-damage functions. We fnally derive the spatial and temporal drivers of both food protection and vulnerability through a multivariate statistical analysis. We apply vine-copulas to derive the best predictors out of a set of candidate variables, including hydrological parameters of foods, exposure to foods, socioeconomic development, and governance indicators. Our results show that riverine food protection levels are much lower than assumed in previous pan-European studies. North-western Europe is shown to have better riverine protection than the south and east, while the divide is not so clear for coastal protection. By contrast, many parts of western Europe have relatively high vulnerability, with lowest value observed in central and northern Europe. Still, a strong decline in food vulnerability over time is also observed for all three indicators of relative losses, suggesting improved food adaptation. Flood protection levels have also improved since 1950, particularly for coastal floods.

Keywords Vine copulas · Flood defences · Vulnerability · Risk · Machine learning

 \boxtimes Dominik Paprotny dominik.paprotny@pik-potsdam.de

¹ Potsdam Institute for Climate Impact Research (PIK), Member of the Leibniz Association, Potsdam, Germany

² Department of Hydraulic Engineering, Delft University of Technology, Delft, The Netherlands

1 Introduction

Floods are a major source of losses in Europe and the threat they pose constantly evolves (Bloeschl et al. [2019;](#page-27-0) Tarasova et al. [2023\)](#page-29-0). Multiple drivers are responsible for changes in risk (Merz et al. [2021](#page-28-0); Kreibich et al. [2022](#page-28-1)), not all of which are well quantifed. Two factors are especially uncertain in food risk assessments over large geographical areas. One is structural food protection, particularly in the form of dikes, and the other is vulnerability, i.e. the conditions determined by physical, social, economic and environmental factors which determine the susceptibility of an individual, a community, assets or system to floods (United Nations Office for Disaster Risk Reduction [2016\)](#page-29-1). Many continental or global food studies have excluded food defences due to lack of data, but this has implications on estimating both present and future risk from riverine and coastal foods (Ward et al. [2017;](#page-29-2) Paprotny et al. [2017](#page-28-2), [2019\)](#page-28-3). Vousdoukas et al. ([2018](#page-29-3)) has found that assumptions on food protection levels are the largest individual source of uncertainty in assessing coastal food risk in Europe, also under climate change conditions. At the same time, vulnerability models are diverse and lead to very diferent results, as showcased by studies comparing various economic damage functions in the same study area (Carisi et al. [2018](#page-27-1); Figueiredo et al. [2018;](#page-27-2) Paprotny et al. [2020](#page-29-4), [2021\)](#page-29-5). Mortality functions are similarly highly uncertain (Jonkman et al. [2008](#page-28-4); Brussee et al. [2021](#page-27-3)). Additionally, evidence shows that vulnerability changes over time, mostly with a downwards trajectory (Jongman et al. [2015;](#page-28-5) Tanoue et al. [2016](#page-29-6); Bouwer and Jonkman [2018](#page-27-4); Formetta and Feyen [2019](#page-27-5); Sauer et al. [2021\)](#page-29-7). Implications for continental-scale assessments, particularly in Europe, which employs extensive food adaptation measures (Vousdoukas et al. [2017](#page-29-8); Steinhausen et al. [2022;](#page-29-9) Dottori et al. [2023](#page-27-6)), are profound. Additionally, impact attribution of past and future losses in context of global change requires even more precise data to quantify the climatic and human drivers of those losses (Kreibich et al. [2019;](#page-28-6) Mengel et al. [2021;](#page-28-7) Scussolini et al. [2024](#page-29-10)).

In practice, pan-European datasets on food protection and vulnerability are very limited. On the structural defences, a frequently used resource is FLOPROS by Scussolini et al. [\(2016\)](#page-29-11). It combines information on nominal (design, or policy-defned) food protection standards with estimates based on the level of economic development. Several European studies have used an alternative literature-based dataset of nominal protection levels from the PESETA IV study (Dottori et al. [2023](#page-27-6)). However, that study also used empirical food impact data to improve the estimates of protection levels. A similar approach was employed by Jongman et al. [\(2014\)](#page-28-8), who diferentiated food protection assumptions by utilizing modelled food impacts and reported food losses. More detailed information is available for some countries. The national food risk assessment for the Netherlands (Vergouwe [2015](#page-29-12)) indicates that the actual reliability of food defences can be far below nominal standards prescribed by law. Data on food protection levels along rivers and coasts of England (Environment Agency [2023\)](#page-27-7) shows far more spatial variation, and usually much lower standards, than assumed in pan-European datasets.

At the same time, dozens of food vulnerability models are available (Gerl et al. [2016](#page-28-9)). The major issue is that they are typically based on local data, or even no empirical data at all, and often not transferable across case studies. Even models based on large microscale damage datasets like HOWAS21 for Germany (Kellermann et al. [2020](#page-28-10)) do not necessarily perform well in other settings (Jongman et al. [2012;](#page-28-11) Wagenaar et al. [2018](#page-29-13); Paprotny et al. [2020](#page-29-4)). Consequently, pan-European studies either have to rely on vulnerability models developed for a specifc environment (Steinhausen et al. [2022](#page-29-9)) or utilize more generic depth-damage functions such as those developed by Huizinga et al. ([2017\)](#page-28-12).

In this study, we present a novel approach to estimate present and past variation in flood protection levels and vulnerability (fatalities, population affected, and economic loss relative to exposed population or assets) in Europe based on constrasting modelled and reported impacts of foods. We build upon recent advances in pan-European riverine and coastal food modelling (Paprotny et al. [2024b;](#page-29-14) Tilloy et al. [2024](#page-29-15)), historical exposure estimation (Paprotny and Mengel [2023\)](#page-28-13), and collection of impact data from documentary sources (Paprotny et al. [2024a](#page-29-16)). We utilize data spanning from 1950 to 2020 to create a multivariate model that is able to infer the spatial and temporal variation in food occurrence and their impacts.

We apply vine copulas to model the complex dependency between the predictors (socioeconomic drivers, food risk and experience) and changes in the level of food protection and food vulnerability. Vine copulas are graphical models that allow the construction of a complex multivariate probability distribution function through bivariate pieces (bivariate copula functions). Because of their fexibility in representing asymmetries in the joint distribution they have found wide application in diferent felds. For example, they have been applied in tunnel engineering ('t Hart et al. [2024](#page-29-17)), reliability analysis of food defenses (Torres-Alves and Morales-Napoles [2020](#page-29-18); Pouliasis et al. [2021](#page-29-19)), in ocean engineering (Jäger and Morales-Napoles [2017](#page-28-14); Mares-Nasarre et al. [2024\)](#page-28-15), hydrology (Tao et al. [2021](#page-29-20)) among many other felds. Additionally, theoretical developments around vine copulas are still being proposed, as highlighted by Pfeifer and Kovács [\(2024](#page-29-21)).

The paper is structured as follows. Section [2](#page-2-0) provides details on the input data [2.1](#page-2-1) and the procedure of creating vine-copula models [2.2.](#page-6-0) In section [3,](#page-10-0) the fnal vine-copula models [3.1](#page-10-1) are validated and compared with other datasets [3.2.](#page-14-0) This enables generating pan-European maps of food protection levels and vulnerability from 1950 to 2020 [3.3](#page-17-0). As historical impact data is very incomplete, we also estimate the magnitude of unreported losses in Europe [3.4.](#page-21-0) Limitations and uncertainties are discussed in section [4](#page-23-0) before concluding in section [5.](#page-26-0)

2 Materials and methods

2.1 Data

2.1.1 Flood event data

Flood protection levels and food vulnerability is analysed in this study using historical information on foods and their impacts. The necessary information was obtained from two flood catalogues. The first, HANZE v2.1 (Paprotny et al. [2024a](#page-29-16)) contains information on date, location and impacts of 2521 riverine, fash, coastal, and compound foods between 1870 and 2020. In this study, we consider a subset of 2037 events that have occurred since 1950. Reported impact data were extracted for three indicators: fatalities (including missing presumed dead), population afected (whose homes were fooded or who have been evacuated), and economic loss (damage or destruction of tangible assets). HANZE also includes information on the area inundated (in $km²$), but was not included in the analysis due to much lower availability of data compared to the other indicators.

The other catalogue from Paprotny et al. ([2024b](#page-29-14)) is a model reconstruction of almost 15,000 potential riverine, coastal, and compound foods between 1950 to 2020. Each event of this model catalogue is "potential" as it does not predetermine whether it led to actual impacts to society or economy. However, the potential events were analysed using available historical information to determine which ones caused impacts (linking them to the HANZE database) and which did not. The modelled food catalogue also estimates potential food impacts, without food protection, for the same four indicators as HANZE: area inundated, fatalities, population afected (all population within the food zone), and economic loss (only considering direct food damage to tangible assets). Both catalogues cover the majority of the European continent (42 countries, see Fig. [14](#page-18-0)). In both catalogues, due to large variations in availability of historical records between countries, transnational floods have individual entries for each country affected.

Flood protection level in this study is defned as the return period above which a flood will occur. A flood in this context is an inundation causing significant socioeconomic impacts. The thresholds for considering a food event "signifcant" are several and described in detail in Paprotny et al. [\(2024a\)](#page-29-16), but generally require some level of loss to population and/or assets, rather than mere inundation of land. The analysis is carried out at the resolution of subnational regions (hereafter, "regions"). Regions in this context are administrative or statistical divisions of countries in the study area that are, except for minor exceptions, consistent with the European Union's Nomenclature of Territorial Units for Statistics, level 3, version 2010 (Eurostat [2020\)](#page-27-8). There are 1422 regions defned in the study area, described in detail in Paprotny and Mengel [\(2023](#page-28-13)). Both historical and modelled food impact zones are considered at this resolution, therefore the inferred protection levels indicate the chance of occurrence of a signifcant food anywhere within the region, rather than in a particular stretch of the coastline or river. It should be stressed that this approach is very diferent from the typical characterization of food protection levels, as the chance of overtopping a dike at a particular location. Our approach is rather similar to the Dutch "dike ring" concept, where the protection level is currently defned as the probability of impacts within a dike ring as a consequence of food protection failure anywhere around the protected area (Vergouwe [2015\)](#page-29-12). In this approach, the protection level strongly depends on the weakest link, as long as its failure would result in inundation and impacts. Also, we are interested in the actual reliability of flood defences, rather than the nominal, official, or otherwise desired level of protection, which has been the focus of previous datasets such as FLOPROS.

The modelled food catalogue of Paprotny et al. [\(2024b](#page-29-14)) contains 5067 potential foods for which there is good historical information enabling confdently dividing the events into those that caused signifcant socioeconomic impacts (1444, or 29 %) and those that did not (3623, or 71 %). However, even impactful foods typically do not afect all regions where they reached hydrologically extreme levels. Non-occurrence of impacts in particular regions during an otherwise impactful event is also of interest. Consequently, our modelling approach includes two levels of impact occurrence: the event level and region level. At frst, it is determined whether the food will have impacts anywhere within the country, and if so, the chance of impacts per potentially-afected region is analysed (Table [1\)](#page-4-0). Out of 45,578 regions within the 5067 potential food events, 5257 (12 %) were determined to have been actually impacted based on documentary sources. The data include riverine, coastal, and compound events, where the latter denotes events during which both riverine and coastal drivers contributed to fooding. As such compound impact cannot be determined for non-impactful foods, only separate riverine and coastal non-impactful foods were included in the data.

Table 1 Target impact variables modelled in this study

Table 1 Target impact variables modelled in this study

As our target variables of food protection failures are binary (impact or no impact), they had to be converted into a continuous variable to be used in vine-copulas. Alternative approaches were considered, but ultimately not applied, as explained in more detail in section [4.2.](#page-24-0) At the event level, potential impact was scaled according to potential impacts relative to maximum loss in the country from any event between 1950 and 2020. The impact was calculated for four indicators and then averaged: area inundated, fatalities, population affected, and economic loss. The index $(0-100\%)$ was converted into 0 to 0.5 range if no impact occurred, and to 0.5 to 1 range if impact did occur. In this way, e.g. 0.4 will represent a hydrologically large food that had no impacts, but 0.6 will represent a hydrologically small flood that nonetheless caused significant impacts. The impact indicator *I* for event *e* is follows:

$$
I_e = \frac{1}{4} \left(\frac{D_{A,e}}{\max D_A} + \frac{D_{F,e}}{\max D_F} + \frac{D_{P,e}}{\max D_P} + \frac{D_{E,e}}{\max D_E} \right) \times 0.5 + M_e
$$
 (1)

where D_e is the damage indicator value (*A* for area inundated, *F* for fatalities, *P* for population affected, and *E* for economic loss) during event *e*, and M_e has a value of 0.5 if impact occurred during event *e*, and 0 otherwise.

Impact occurrence at regional level was scaled according to the relative contribution (%) of potential losses in the region to the whole event (average of the four impact indicators). This share of losses was converted into 0–0.5 range if no impact occurred, and to 0.5 to 1 range if impact did occur. The impact indicator *I* in region *r* for event *e* is follows:

$$
I_{r,e} = \frac{1}{4} \left(\frac{D_{A,r}}{D_{A,e}} + \frac{D_{F,r}}{D_{F,e}} + \frac{D_{P,r}}{D_{P,e}} + \frac{D_{E,r}}{D_{E,e}} \right) \times 0.5 + M_{r,e}
$$
 (2)

where $M_{r, e}$ has a value of 0.5 if impact occurred in region *r* for event *e*, and 0 otherwise.

Flood vulnerability was considered for three variables: fatalities, population afected, and economic loss. Availability of data varies between variables (Table [1\)](#page-4-0). Absolute impact of each was converted into relative impact using the potential modelled loss from the other catalogue. Only historically afected regions were considered when computing potential losses. Modelled loss is based on static depth-damage functions in case of mortality and relative economic loss, while modelled population afected is simply the total exposed population within fooded grid cells. The function for mortality is a S-shaped function shown in Jonkman et al. ([2008\)](#page-28-4), and the economic loss functions for fve fxed asset types (dwellings, agriculture, industry, services, infrastructure) are from (Huizinga et al. [2017](#page-28-12)). In total, it was possible to compute potential loss for 1504 (74%) of historical foods in the HANZE database, which represent 81 % of known fatalities in Europe since 1950, and 96 % of both population afected and economic loss. For more information on potential damage modelling, as well as an analysis of the accuracy and completeness of the food reanalysis we refer to Paprotny et al. ([2024b](#page-29-14)).

Mortality (fatalities relative to exposed population) was considered as two variables, the chance of any fatalities occurring, and then the number of fatalities if more than zero was indicated. This was done because in 43 % of historical floods no deaths were reported, resulting in a highly skewed distribution of mortality. Such a distribution would severely degrade the performance of any statistical model. To compute the chance of fatalities, the same approach as for food protection was used. The chance of fatalities was scaled according to potential impacts relative to maximum loss in the country from any event between 1950 and 2020. The impact was calculated for four indicators and then averaged: area inundated, fatalities, population affected, and economic loss. The index $(0-100\%)$ was converted into 0 to 0.5 range if no fatalities occurred, and to 0.5 to 1 range if at least one fatality occurred. Equation [1](#page-5-0) is applicable here with substituting "impact" for "fatalities".

Finally, relative impact indicators for fatalities, population afected and economic loss were capped at 100 % to avoid a minor share of cases where food footprint or exposure was strongly underestimated by the modelled flood catalogue.

2.1.2 Predictors

To infer the distribution and changes in food protection levels and vulnerability, multiple candidate variables were considered (Table [2](#page-7-0)). Most of them were considered at the level of regions, aggregated to event footprint, actual or potential, where appropriate for a particular target variable. Governance and demographic indicators are at country level, while potential food impacts and hydrological intensity are based on modelled food footprints at 100-meter resolution. Each variable refers to socioeconomic situation at the time of the event, unless noted otherwise in Table [2](#page-7-0). In addition, several categories of variables ("Economic development" and all further down the table) were additionally considered in a 'lagged' version, i.e. average of conditions in the 30 years preceding the food event. The data was collected from several sources, which are indicated in Table [2](#page-7-0).

2.2 Methods

2.2.1 Vine copulas

Vine copulas are graphical models that allow the specifcation of a multivariate probability distribution through bivariate pieces. More specifcally a *vine* is a sequence of trees (undirected acyclic graphs) $\{T_1, \ldots, T_{d-1}\}$ where T_1 is a tree on *d* nodes and the edges of each tree become the nodes of next tree for T_2, \ldots, T_{d-1} . In particular vine copulas are constructed with *regular vines*. A regular vine is, roughly, a vine where two edges of T_i are joined as nodes in T_{i+1} if they share a common node in T_i for $i \geq 1$. An example of a regular vine on 5 nodes is presented in Fig. [1](#page-8-0).

Vine copulas use the graphical structure of a regular vine to construct a multivariate probability distribution. Each node in the frst tree of the regular vine is associated with a random variable with an invertible cumulative distribution function, while the probabilistic dependence between variables is approximated through bivariate (conditional) copulas (Nelsen [2006\)](#page-28-16). A bivariate copula *C* is a bivariate distribution function with uniform [0, 1] margins. If *H* is the joint distribution of *X* and *Y* then $H(x, y) = C(F(x), G(y))$ where *F* and *G* are the marginal distributions of *X* and *Y* respectively.

In Fig. [1](#page-8-0) for example, the pair of numbers in the frst tree of the vine indicate the copula joining the corresponding variables. The edges in the lower trees indicate the bivariate copula between the variables to the left of the horizontal line conditional on the variables to the right of the line. For a comprehensive treatment of vine copulas see Czado [\(2019](#page-27-9)).

The number of vine copulas that may be ftted to a dataset is extremely large (?). For "lower" dimensional datasets (up to 7 variables), ftting all possible vine copulas may be feasible. However for 8 variables or larger this task becomes computationally challenging. Recently a dataset containing all regular vine structures on up to 8 nodes (more than 660 million) has been presented in Morales-Napoles et al. [\(2023](#page-28-17)). For larger dimensions the

algorithm proposed in Dißmann et al. [\(2013](#page-27-11)) and variations of it are mostly used. In this research because of the large number of possible models under investigation we opted for Dißmann's algorithm.

2.2.2 Creating vine copula‑based models

As highlighted in Tables [1](#page-4-0) and [2](#page-7-0) in section [2.1,](#page-2-1) there are six target variables analysed in the study and numerous candidate predictors. Obtaining the optimal model confguration was a procedure with several steps. Firstly, six connected vine copula models were defned, representing one target variable each (Fig. [2\)](#page-8-1). Flood protection level is represented by two vines, the frst one estimating the impact probability at event level, and the second one at regional level. The event level is used to establish the chances of a signifcant impact of a potential food event, while the regional level is used to establish a more precise location of those impacts. Then, the vulnerability models can be applied to the afected area. Relative population afected and economic loss are modelled with one model each, while mortality (fatalities relative to population) is split into two models, one indicating the chance of at least one fatality occurring, and the second estimating the magnitude of fatalities.

Fig. 2 Structure of the modelling framework. Each target variable is modelled independently with a vine

Each vine model was tested frst with 74 potential predictors, except for the impact at regional level, for which not all predictors were appropriate, namely variables aggregated at event (potential food impact) or country level (governance and demographic indicators). Two aspects of each vine model needed to be optimised: the selection of predictor variables and the dependency structure within the vine. A diferent approach had to be applied for the two tasks. As the Chimera atlas is capable of working with vines with four to eight nodes, we built each model stepwise starting with four nodes. Within those four nodes, one was the target variable and one was the predictor variable with the highest unconditional rank correlation with the target variable, except for food impact at event level, which was purposely chosen as return period. This is because this predictor is most relevant to computing the protection level, and was also the highest correlated variable for impact at regional level. The other two variables changed at each iteration, until all possible combinations of remaining predictors were exhausted. Within each iteration, the optimal dependency structure came from the algorithm of Dißmann et al. ([2013\)](#page-27-11). Here, the best vine structure, meaning not only the confguration of arcs but also the type of copulas representing each arc, was selected according to the Akaike Information Criterion (AIC). It should be noted that we limited the search of optimal copulas within the vine models only to one- and two-parameter copulas.

Every possible variable combination was then compared according to the validation metric. In case of food protection models and the chance of fatality occurring, the % success rate in inferring the "yes" or "no" state of the target variable was calculated. For continuous quantities (relative losses), several metrics were analysed, but primarily the coefficient of determination and the Kling-Gupta Efficiency (KGE) score. The latter metric integrates correlation, bias and error and has seen wide use in hydrology (Knoben et al. [2019\)](#page-28-18). The vulnerability models were further applied to the potential losses from the food catalogue in order to compare the trends in absolute observed and inferred losses. One variable, inclusion of which improved the model's performance the most, was selected as predictor. Then, the test was repeated with fve nodes: the target variable, the original predictor from the previous test, the newly selected predictor, and two predictors changing at each iteration. Again, all combinations of the latter two variables were tested. The experiment continued with more nodes added to the vine until it reached eight nodes or addition of further nodes did not improve the model's performance. The same process was done separately for each of the six target variables. After a series of preliminary tests, due to the large computational burden of the algorithm for an adequate amount of samples, the number of candidate predictors was reduced. We excluded predictors that were strongly similar to each other and were almost never better predictors than others: food experience lagged by 5 years, food risk in constant exposure, country-level demographic indicators, and 30-year averages of the socioeconomic variables. This has reduced the potential predictors to 35 variables (23 for flood impact per region).

Each run of the vine copula models with a given set of predictor variables was analysed through a 10-fold cross-validation. In case of food protection models, the training dataset consisted of a random 10% sample of the full dataset, and the remaining 90 % of the data was used for validation. In each of the 10 runs, the training sample had no overlaps with any other run. For vulnerability models, the amount of available data is much smaller, therefore a higher share of data was used for training and therefore could overlap partially with samples in other runs. For modelling the chance of fatalities, 1/3 of data were used for training, while for models inferring the relative loss, 1/2 of the data was used. Average outof-sample validation results were considered when choosing predictor variables. Once the fnal selection of predictor was made, the vine copula models for further application were run once more using the algorithm, to defne the best dependency structure using the entire sample, except for impact at region level, where due to large sample size, 1/2 of the data was used. Such structured and quantifed models were used in section [3](#page-10-0), except validation results in Table [4](#page-13-0), which are the average in-sample and out-of-sample results.

In section [3.3.1](#page-17-1), the implied return period of signifcant damages from foods is shown. The two vine models of food protection levels provide the probability of fooding given the predictors, among which is the average return period among afected river sections or coastal segments. This return period was calculated using the peak-over-threshold approach with a Generalized Pareto distribution applied over detrended 1950-2020 6-hourly river discharge and hourly sea level. Therefore, by integrating the hydrological return period with dike failure probability, we could derive the recurrence interval of signifcant fooding. We approximated the analytical solution to this problem with a numerical method used to compute expected annual damage from food maps at given return periods based on distribution-free median plotting position (Olsen et al. [2015](#page-28-19)). This involves generating a series of 1000 random events defned by return period *T*:

$$
T = \frac{n + 0.4}{m - 0.3}
$$
 (3)

where *n* is the length of the series of events and *m* is the rank of the event.

3 Results

3.1 Vine copula models

The fnal vine copula models for inferring food protection levels and vulnerability have from five to seven nodes, i.e. they have four to six predictors of the target variables (Table [3\)](#page-11-0). In total, 18 diferent predictors are used at least once among the models. All categories of predictors are present, except country-level demographic indicators. Variables related to economic development occurred most frequently (8 out of 28 cases), with GDP per capita used in four of the models. In each case, it was negatively correlated with flood occurrence or relative losses. Though GDP per capita was not relevant for the number of fatalities, GDP per sector (either share of agriculture or industry in the economy) was important and positively correlated with the chance and magnitude of fatalities. GDP per sector was also presented in the models on food protection (event level) and population afected.

Hydrological indicators were used six times, always positively correlated with impacts, except for average water depth in the potential impact zone, which was negatively correlated with the magnitude of fatalities. Similarly, food experience of the past 20 or 30 years was negatively correlated with magnitude of fatalities and economic loss. However, the number of past foods positively correlated with chance of fooding, indicating the persistence of food protection defciencies. Relative losses strongly scale to the magnitude of damage potential, as widespread foods tend to have lower individual impacts than smaller fash foods. Similarly, food occurrence is positively correlated with potential food risk over 1950-2020, but negatively correlated with relative population afected.

Other variables are less present in the models: degree of urbanization was important only for the chance of fatalities (positive correlation). Structure of fxed assets in the afected area was included in two models. The share of residential assets was negatively correlated with

Fig. 3 Vine copula model for food impact at event level

Fig. 7 Vine copula model for relative population afected

 $3, 4|1, 2, 5, 6$

Fig. 8 Vine copula model for relative economic loss

Table 4 Validation results, average for 10-fold cross-validation. Percentages in brackets are the theoretical success rates of a random model. $KGE = Kling-Gupta$ Efficiency (Knoben et al. [2019\)](#page-28-18)

Variable	Metric	In-sample	Out-of-sample
Flood impact (event level)	Impact - % correct	64% [28%*]	63% [29%*]
	No impact - % correct	86% [72%*]	86% [71%*]
Flood impact (region level)	Impact - % correct	31% [11%*]	31% [12\%\)
	No impact - % correct	91% [89%*]	91% [88%*]
Mortality (chance of fatalities)	Fatalities - % correct	67% [58%*]	64% [57%*]
	No fatalities - % correct	53% [42%*]	54% [43%*]
Mortality (magnitude)	R^2	0.67	0.67
	KGE	0.58	0.58
Relative population affected	R^2	0.32	0.29
	KGE	0.37	0.37
Relative economic loss	R^2	0.44	0.44
	KGE	0.44	0.45

relative population afected, while the share of agricultural assets was positively correlated with relative economic loss. Land use variables appear only once, with negative correlation between the share of land under artifcial surfaces other than urban fabric (mainly industrial/ commercial zones and transport-related infrastructure) and the chance of fatalities. Finally, a governance indicator is used only in the vine copula for relative economic loss, with a negative correlation.

More detailed information on the vine copula models is provided in the Supplementary Information: joint distributions of raw, ranked, and sampled data in addition to vine tree sequences (S1), correlation matrices (S2) as well as types and parameters of copulas used (S3).

3.2 Model validation

3.2.1 Sample validation

Table [4](#page-13-0) presents validation results averaged from 10-fold cross-validation, both in-sample and out-of-sample. Three out of six models were evaluated by the success rate in correctly classifying the events as impactful or non-impactful. In all cases, the models achieved higher performance than a theoretical random model, which would have reproduced the frequency of impacts (or fatalities). Flood protection at event level best considering the ratio of impacts and non-impacts, while the mortality model was the least successful.

The remaining three models reproduced relative losses and were analysed with several metrics (Table [4](#page-13-0)). The best performance was achieved by the mortality model, though some negative bias was observed (Fig. [9](#page-15-0)a). Most of the fatality rates are very low - average ratio of reported fatalities to modelled potential fatalities with a static depth-damage function is only 0.5 %. By contrast, the ratio of average reported population afected to potential population exposure is 27 %, and for economic loss $-$ 22 % of potential loss. The latter models have lower performance overall, while relative population (Fig. [9](#page-15-0)b) tends to be overestimated and relative economic loss underestimated (Fig. [9c](#page-15-0)). In addition to computing relative loss, the models were analysed in terms of how well they reproduced trends in absolute losses in Europe overall (section [3.2.3\)](#page-15-1)

3.2.2 Comparison with other food protection level datasets

Probability of impact from the two food protection models were converted into the implied return period of an signifcant impactful event in each region (section [2.2.2](#page-8-2)). In Fig. [10](#page-16-0), our results for riverine foods in year 2020 (section [3.3.1\)](#page-17-1) are compared with nominal riverine food protection standards from FLOPROS (Scussolini et al. [2016\)](#page-29-11) and PESETA IV project (Dottori et al. [2023](#page-27-6)). In both cases, the nominal standards are far higher than our calculations. Weighted by food hazard in each region, the average protection level from our data was only 22 years, compared to 186 years in FLOPROS (Fig. [10](#page-16-0)a) and 209 years in PESETA IV (Fig. [10b](#page-16-0)). 54 % and 58%, respectively, of regions in the study area have 100-year protection standards in those datasets, compared to only 5 % in our results. However, FLOPROS includes not only direct information on local food protection policies, but also gap-flls information with a regression with GDP per capita, mostly showing less than 30-year protection for eastern and southern Europe, and 40–60 years for western and northern parts of the continent. PESETA IV, bar one exception, assumes a protection level of at least 50 years across Europe. The spatial comparison between the datasets is shown in Supplement S4.

Despite the large diference between our results and other commonly used datasets, more detailed local data show much better alignment with our results. For instance, the average food protection standard weighted by dike length for riverine foods in England (Environment Agency [2023\)](#page-27-7) is only 30 years, close to our weighted average of 26 years. Meanwhile, the other pan-European uniformly assume a 100-year protection for all of England, except 1000 years for London. The Environment Agency data indicate such design standards only for 10 % of dikes in England. Weighted average coastal protection in England according to the same source is 159 years, only marginally lower than our average of 167 years. The other datasets do not provide estimates for coastal food protection levels.

(a) fatalities (only events with at least one fatality)

3.2.3 Absolute impacts and trends

Absolute modelled food impacts for 1504 historical foods between 1950 and 2020 are compared with reported losses for the same events, where available (as indicated in Table [1\)](#page-4-0). This enables comparing the performance of the model in reproducing historical time series of losses, even though this depends not only on the accuracy of the vulnerability models of this study, but also the hazard and exposure reconstructions of Paprotny and Mengel [\(2023](#page-28-13)); Paprotny et al. ([2024b](#page-29-14)).

Overall number of fatalities is underestimated by 10 %, primarily for major events (Fig. [11](#page-16-1)). 43 % of historical fatalities occurred in only five events, all clustered in only three years (1953, 1962 and 1973). In the remaining 851 events, there is a positive bias of 49 %, with both observations and models showing slight upward trends in absolute

Fig. 10 Comparison between food protection levels for riverine foods from this study for year 2020 and from two other public datasets, as return period in subnational regions of the study area

Fig. 11 Annual reported fatalities in Europe and model estimate

fatalities (without correcting for population growth or other drivers). The modelled trend is stronger than in reported yearly data.

Population afected and economic loss is less concentrated: the top fve events were responsible for about 20 % of reported losses. However, data is only available for about half of the events. Population affected is underestimated by 23 % and economic loss by 22 % (Figs. [12](#page-17-2) and [13\)](#page-17-3). Excluding the top 10 events, for which the model strongly under-performs, the bias is 0.2% and 6 %, respectively, for population affected and economic loss. Whereas upwards trend in the model closely matches the reported one, the trend in economic loss is overestimated. In both cases, the upward trend should not be interpreted standalone, as it was not corrected for exposure growth or other drivers.

Fig. 12 Annual reported population afected in Europe and

3.3 Flood protection levels and vulnerability in Europe, 1950‑2020

3.3.1 Flood protection levels

Implied food protection levels, i.e. the return period of food with signifcant impacts, was computed for subnational regions in the study area between 1950 and 2020. We show separate estimates for coastal and riverine events, which cover 405 and 1364 regions, respectively. Separating the two types of foods was possible through three variables: economic risk, duration, and 20-year food experience, which have diferent values depending on the type of event. For this analysis, we assume that duration at event or regional level equals duration with the same return period as assumed for discharge or sea level. To estimate the average protection levels across Europe in Fig. [14](#page-18-0), we used the synthetic events with diferent return periods (section [2.2.2](#page-8-2)) to create a random timeseries of events that would have overwhelmed food defences in some regions in a given year. The total number of impacts in Europe relative to the number of applicable regions (coastal or riverine) then indicated the average return period of region-level impacts.

Overall, there is better protection from coastal foods than riverine, and the former has been improved to a slightly larger degree than the latter (Fig. [14\)](#page-18-0). The average

Fig. 14 Changes in protection levels weighted by food hazard area, by food type, 1950-2020

European coastal protection level improved from 61 years in 1950 to 135 years in 2004, before declining to 116 years in 2020. By contrast, riverine protection only increased from 15 years in 1950 to 22 years in 1979, when it started declining to only 18 years in 2020. While this is much less than nominal protection standards analysed in section [3.2.2,](#page-14-1) it is consistent with actual food occurrence. Riverine fooding, according to data collected in the HANZE dataset (Paprotny et al. [2024a](#page-29-16), [2024b](#page-29-14)), afected an average region in the study area 4.05 times from 1950 to 2020, which implies a return period of 17.5 years. Using a series of random events of diferent return periods sampled over estimated regional protection, the inferred return period of foods is 19.2 years (95 % confdence interval: 18.7−19.7 years). In case of coastal foods (including compound events), 405 regions where such foods are possible were afected with a return period of 66 years. Average protection from coastal foods over 1950-2020 is estimated here as 99 years (95 % confdence interval: 88-111 years). However, many regions afected by compound events are outside the coastal food zone; the observed return period for coastal foods only is 120 years. It should be noted that the clear diference between riverine and coastal food occurrence was captured by the model approach despite all types of foods being assembled together in the input data.

Spatially, riverine food protection levels (Fig. [15](#page-19-0)a) follow the somewhat expected pattern of better developed countries exhibiting higher protection levels, but not so much for coastal foods (Fig. [15](#page-19-0)b). As food experience is an important predictor of subsequent inundation, and we define flood protection levels in terms of occurrence of signifcant impacts, locations of major past events are often visible, e.g. North and Adriatic seas for coastal and compound fooding, and mountainous regions for riverine fooding. Many Balkan countries as well as Czechia, Italy, Spain, and Sweden are indicated as having no improvement or deterioration in protection from riverine foods, in contrast to e.g. Germany, the Netherlands, Poland and Portugal, where the opposite was found. The model estimates large improvements in coastal protection everywhere, particularly southern and eastern Europe, with the sole exception of Denmark.

(a) coastal floods

(b) riverine floods

Fig. 15 Flood protection levels (return period in years), by region and type of food, 1950 and 2020

3.3.2 Flood vulnerability

Similarly to food protection levels, inferred vulnerability at regional level was computed for all regions. The indicators presented in this section refer to inferred impacts relative to modelled potential impacts of an average event that have (potentially) afected a given region between 1950 and 2020 in constant 2020 exposure level, combined for riverine and coastal foods. The average return period along river sections or coastal segments for computing the chance of fatalities is also an average of region-specifc events. Flood experience used in two models was computed based on all types of events.

All three impact indicators - mortality, population afected and economic loss - have shown a considerable decline between 1950 and 2020 (Fig. [16\)](#page-20-0). Relative population afected, weighted by food hazard area of each region, was estimated at 34 % for the hypothetical average food event in 1950, but only 17 % in 2020, though higher than around 2010 when it was estimated at 15 %. Economic loss started declining later, in the 1970 s, but still declined substantially from 34 $\%$ to 20 $\%$. Relative fatalities had the fastest decline, from 9.0 % in 1950 to 3.8 % in 2015, before increasing narrowly to 4.0 % in 2020.

At regional level, the patterns should be interpreted with the consideration that they are relative to potential damage represented by static depth-damage functions for fatalities and economic losses. Consequently, the distribution of mortality rates (Fig. [17\)](#page-21-1) is not expected to be similar to historical fatality distribution (i.e. strongly skewed towards southern Europe relative to the northern part). Decline in mortality is estimated to be the highest in south-eastern Europe, compared to limited decline or increase in the Alpine region, though it was already relatively low in 1950. By contrast, the model pre-dicts a decline in relative population affected in all countries (Fig. [18](#page-21-2)), particularly in northern Europe and Germany, and a much smaller decline in the south-eastern part of the continent. On the other hand, the model also predicts that economic vulnerability declined in the southern countries (particularly Portugal and Spain), with less pronounced declines in the northern part of the continent (Fig. [19](#page-22-0)). For both population afected and economic loss regions impacted by slow-onset, large riverine foods (eastern and south-eastern Europe in particular) have higher vulnerability than those where rapid but spatially limited fash foods are dominant, especially in the Alpine zone.

Fig. 16 Inferred impacts of an average food relative to modelled potential impact, weighted by food hazard area, by impact type, 1950-2020

Fig. 17 Inferred fatalities of an average food relative to modelled potential impact, by region, 1950 and 2020

Fig. 18 Inferred population afected by an average food relative to modelled potential impact, by region, 1950 and 2020

3.4 Estimating unreported food losses in Europe

Published historical food impact datasets, such as HANZE (Paprotny et al. [2024a\)](#page-29-16), are not complete. Whereas fatality data was available for 99 % of events in HANZE that have occurred back to 1870, population afected was provided only for 43 % of events, and economic losses for 40 %. Therefore, the full magnitude of food losses in Europe is unknown, and has implications e.g. for validating pan-European food risk assessments. Combining the modelled flood catalogue of Paprotny et al. $(2024b)$ $(2024b)$ with our vulnerability models, it is possible to estimate the missing impact data.

Fig. 19 Inferred economic loss of an average food relative to modelled potential impact, by region, 1950 and 2020

HANZE contains 2037 events since 1950 where at least one impact statistic (out of four, including inundated area) is known ("A" list). Of these, 1504 events were captured by the model, of which 8 had unknown number of fatalities and almost half were missing the number of population afected and economic loss (Fig. [20](#page-22-1)). Consequently, the estimated number of unreported fatalities is below 100, adding less than 1 % to the known total. However, missing impacts for the other two categories add 35-38 % to

□ Unreported impact (list B) included in the model

- Unreported impact not included in the model
- Reported impact not included in the model
- □ Unreported impact included in the model
- Reported impact included in the model

the total, even if an average event without known impact is estimated to have had 60 % smaller magnitude than an event with recorded impacts.

The remaining 533 events not included in the modelled food catalogue consist mostly of fash foods. Impact reporting rate for population afected and economic loss is smaller, covering about one-third of events, and where available, they were about six times smaller on average compared to events included in this study. The missing impacts cannot be directly reproduced with our approach, as they are not in the catalogue, but we can extrapolate from the previous categories of events. We assume that the ratio between an average event with unknown impacts and known impacts for the events not included in the model is the same ratio as for those events that were captured in the catalogue. This adds little additional damage to what was reported, namely 0.1 % more fatalities and about 3 % more population afected and economic loss.

Finally, the food catalogue identifed additional 237 foods ("B" list) for which there is no complete impact data, but descriptive sources or partial data indicate that they nonetheless caused signifcant socioeconomic impacts. Estimated impacts of those events are, on average, similar to those with unknown impacts in the "A" list, and slightly lower than the estimated average for all 2037 events on that list. 12 foods on the "B" lists are related to "A" events such as they represent an earlier or later phase of the "A" event, and their impacts were reported in list "A". The remaining 225 "B" events increase the known impacts by further 7-11 %, depending on the category of impacts. Overall, the 2274 known food events resulted in an estimated 11,000 fatalities, afected 14 million people and caused 435 billion euro losses between 1950 and 2020 (Table [5](#page-23-1)). We estimate that the modelled food catalogue reproduces 83 % of fatalities and 95-96 % of other impacts. On the other hand, the reported impacts cover 92 % of estimated total fatalities, and only 66-67 % of population afected and economic loss. This indicates that while fatality data are reliable enough for use in pan-European food studies, reported population afected and economic losses should be taken with a degree of caution due to their incompleteness.

4 Discussion

4.1 Uncertainties in input data

The study is based on extensive data collection and modelling efort carried out primarily in several preceding studies (Paprotny and Mengel [2023](#page-28-13); Paprotny et al. [2024a](#page-29-16), [2024b;](#page-29-14) Tilloy et al. [2024\)](#page-29-15). Inevitably, there are numerous sources of uncertainty involved in the

Category	Estimated total	% of estimated total captured by model	Known impacts as % of estimated total
Events	2262	76.4	X
Fatalities (thousands)	11.2 [10.8-11.9]	82.7 [$82.1 - 83.6$]	91.6 [85.6-94.9]
Population affected (millions)	13.8 [13.0-14.7]	95.8 [95.8-95.8]	66.2 [62.1-70.5]
Economic loss (billions, 2020 euros)	435 [403-472]	95.3 [95.2-95.3]	67.0 [61.7-72.3]

Table 5 Estimated impacts of foods in Europe by impact category. Numbers in brackets are the 95 % confdence interval

input data. This relates to both the reported impacts and their reconstruction in the modelled potential food catalogue. Observed impacts were compiled from many sources with a varying level of reliability and completeness. The availability of information also strongly varies between countries, or even within countries, with a noticeable temporal bias, analysed in detail in Paprotny et al. ([2018\)](#page-28-20). For example, there are extensive food databases and catalogues available for France, Italy, Norway, Portugal, Spain, Switzerland, and certain Balkan countries, but large scattering of information makes it difficult to collect data for other countries, such as Austria, Germany, and the United Kingdom. There are also large gaps in data resulting from Communist-era suppression of public food impact reporting, even if they were clandestinely recorded in minute detail and sometimes made available in more modern times. Apart from missing data, major errors in food impact databases occur. They were corrected in HANZE only to the extent that alternative sources of information allowed. In addition, the data on non-impacts is incomplete and not even across the domain, due to limitations in sources and difficulty of ruling out the possibility of impacts for many events. For more detailed discussion we refer to Paprotny et al. ([2024a,](#page-29-16) [2024b](#page-29-14)).

The modelled food catalogue (section [2.1.1\)](#page-2-2), which enabled converting absolute losses into relative losses, also has limitations. Spatial and temporal resolution of the hydrological model, despite being the highest ever applied to model European riverine foods, is still not enough to capture many smaller fash foods. Also, the food footprint was reconstructed using flood hazard maps made for rivers with a minimum catchment area of 100 km^2 . Consequently, only 55 % of fash foods in HANZE were included in the model catalogue, and only 84 % of afected regions within that subset were included in the footprint. For slowonset riverine foods, the statistics are better, as the model captured 91 % of both events and their footprints. This nonetheless can lead to underestimation of exposure during the event, overestimating relative losses, and underestimate the gap-filled losses (section [3.4](#page-21-0)). Coastal foods, though modelled using a storm surge model of Paprotny et al. [\(2016](#page-28-21)) that is not as precise as some newer studies like Muis et al. ([2020\)](#page-28-22), are most complete, with 90% of events reproduced in the model together with 98 % of their region-level footprints. It should be noted that the models performed well throughout the time period in question, with only some degraded performance in the 1950 s. For a more detailed analysis of the accuracy of the modelled data we refer to Paprotny et al. [\(2024b](#page-29-14)); Tilloy et al. ([2024\)](#page-29-15).

4.2 Uncertainties in the analysis and results

Limitations in the data notwithstanding, there are many choices possible to analyse it statistically. Here, we opted for the vine-copula method, which required converting some of the target variables to continuous by combining it with another variable (section [2.2.2](#page-8-2)). We tested a diferent approach to discrete variables (impact or fatality occurrence), such as discrete Bayesian Networks and Random Forests, but they did not lead to better results than the method used herein. We also tested whether adding food event type in a Random Forest would improve results, but it was indicated as one of the least useful predictors. As for the vine-copulas, we tested the optimal model structures within a given set of variables using the algorithmic approach of Dißmann et al. [\(2013](#page-27-11)). A brute force approach (testing all possible vine confgurations) would not be computationally feasible when testing concurrently all possible variable combinations, though is an option with a fxed variable list. Applying the brute force approach to the fnal variable composition would have slightly improved the AIC score of the fnal models. The validation metrics would improve only slightly for all models, by less than 1 percentage point for food protection and chance of mortality (section [3.2.1\)](#page-14-2). Virtually no diference for magnitude of fatalities and economic loss was recorded, only for population afected the brute force model would improve both $R²$ and KGE by up to 0.02. Due to the small impact on validation and to keep the approach consistent, we utilized the algorithm throughout.

Estimating food vulnerability from a large set of very diverse food events spanning over 70 years was shown as feasible in this study, even if results are far from a perfect match with observations (section $3.2.1$). Low correlations between potential predictors and relative losses are a persistent problem even in microscale, using building-level data (Merz et al. [2013](#page-28-23); Wagenaar et al. [2018;](#page-29-13) Paprotny et al. [2020\)](#page-29-4). At European scale, validation of modelled absolute losses has rarely been published. Recent studies still show a large diference between modelled and reported economic losses. In Steinhausen et al. [\(2022](#page-29-9)) modelled residential riverine food losses for 1981-2010 were 39 % below total estimated on HANZE in its previous, much less complete iteration of (Paprotny et al. [2018](#page-28-24)). For the same period, Dottori et al. ([2023\)](#page-27-6) estimated total riverine economic losses in the same period as 75 % above those reported by the reinsurer Munich Re, and even more above EM-DAT or the 2018 version of HANZE. Our inference for 1981-2010 for all types of foods is 5 % below reported. In general, lower overall bias in estimating absolute losses was achievable, but at the detriment of reproducing the historical temporal trend. However, both total losses and trends in the past 70 years are heavily infuenced by a very small number of high-magnitude events, particularly in case of fatalities (section [3.2.3\)](#page-15-1). Excluding less than 1 % of the top events gives greater alignment in the trends and lower bias, except for bias in fatalities. Graphs of absolute losses excluding 10 largest events are shown in Supplement S5.

Flood protection levels follow a specifc defnition, not comparable to nominal protection levels used in other studies (section [3.2.2](#page-14-1)). The most comparable data to our approach is available for the Netherlands, where probability of fooding is evaluated holistically for entire systems of primary food defences ("dike rings") protecting a large area. The VNK project data (Vergouwe [2015\)](#page-29-12) indicate that dike rings have been mostly below nominal protection levels. Almost a quarter of Dutch population within dike rings lives in those with an actual protection level below 100 years. This includes the North Holland dike ring (population 1 million), which is routinely indicated as having 1 in 10,000 years protection. However, due to degraded reliability of some hydraulic structures in the area, actual protection level was shown to be below 100 years. Similarly low protection levels were found in many areas at risk of riverine foods, though no precise value was given. Our weighted average protection for the Netherlands is 59 (riverine) or 144 years (coastal), though with large variations between regions.

Still, they are consistent with empirical food occurrences in Europe. This approach has drawbacks, however, for vulnerability estimation. We calculate relative losses as the actual impact divided by potential impact, the latter of which assumes no defences in afected regions. In reality, food protection will not fail everywhere, even in the scale of a subnational region. Further, we assume a static depth-damage relation for mortality and economic loss, only using total exposure for the population afected. The relative loss magnitudes need to be interpreted in this particular model setting. On the other hand, this is comparable with the approach of previous pan-European studies, where either there is no fooding or whole regions or countries are fooded everywhere (Jongman et al. [2014;](#page-28-8) Dottori et al. [2023\)](#page-27-6). Our method introduces more spatial variation in protection levels and vulnerability, and reduces bias in estimating impacts of foods. Still, it has to be highlighted that the return period was computed as geometric averages for afected river segments or coastal sections. The actual, impactful fooding could have occurred in locations where the return period was particularly high. Using the return period of the peak would be an alternative, but increase the uncertainty and likely overestimate the protection level, which would be rather lower than the peak. Again, as pan-European food hazard maps are created for a uniform return period throughout the domain, the hydrological intensity over a broader area should be considered.

The input return periods were computed with a specifc method and temporal resolution of the data (section [2.2.2](#page-8-2)). The protection levels from this study could mismatch water levels computed with a diferent approach to extreme value analysis. However, the same is true for any published food protection levels dataset, and ultimately, our results were shown to be consistent with the empirical recurrence interval of foods in European regions for both coastal and riverine foods (section [3.3.1](#page-17-1)). Further, the maps of food protection levels and vulnerability (section [3.3\)](#page-17-0) were computed using assumptions for the values of certain variables, representing a more "average" food event. By contrast, unreported food losses (section [5](#page-23-1)) were calculated using information for specific, historical events. The models can be applied in diferent settings depending on the purpose, and trends shown in this study (Figs. [14](#page-18-0) and [16\)](#page-20-0) could be diferent in such cases.

5 Conclusions

In this study, we constructed the frst pan-European maps of food protection levels and food vulnerability at subnational level covering a period of 70 years using advanced probabilistic models such as vine-copulas. Informed by reported food impacts combined with model reconstruction of past foods, our models and data have the potential to redefne how those aspects are parameterized in pan-European food risk assessments. This work does not aim to be a substitute for local knowledge, such as microscale damage models trained on case studies, or detailed food protection data, from high-resolution elevation models or dike reliability assessments (such as the VNK study for the Netherlands). Rather, it enables applying a consistent approach to continental-scale studies and flls gaps where more detailed data or methods are not available.

An important application of the study would be impact attribution. Counterfactual food protection levels and food vulnerability would enable testing the sensitivity of impacts from diferent historical foods to changes in food management and adaptation in Europe. It could enable detecting, for instance, events and regions in the potential food impact catalogue that would have caused losses without improved protection. Conversely, it could identify foods that would not have happened without an increased return period of fooding induced by climate change or human alterations in the catchment-level water cycle. Economic analysis of adaptation options (structural protection versus reducing vulnerability) could also be improved with our results. Finally, it could help create more realistic baseline projections of future protection levels and vulnerability under changing (and uncertain) socioeconomic conditions.

Supplementary Information The online version contains supplementary material available at [https://doi.](https://doi.org/10.1007/s11069-024-07039-5) [org/10.1007/s11069-024-07039-5.](https://doi.org/10.1007/s11069-024-07039-5)

Acknowledgements The authors would like to thank Patricia Mares-Nasarre for technical discussions.

Author Contributions Conceptualization, data collection and funding acquisition: DP. Method development and writing the original draft: DP and OMN. Implementation of the methods: DP and CMPH. Visualization, reviewing and editing: all authors.

Funding This research has been supported by the German Research Foundation (DFG) through project "Decomposition of food losses by environmental and economic drivers" (FloodDrivers), grant no. 449175973.

Data Availability The input food catalogue is available on Zenodo [\(https://doi.org/10.5281/zenodo.10629](https://doi.org/10.5281/zenodo.10629443) [443\)](https://doi.org/10.5281/zenodo.10629443). The output food vulnerability and protection estimates for Europe are also available on Zenodo [\(https://doi.org/10.5281/zenodo.10911302\)](https://doi.org/10.5281/zenodo.10911302). The data can also be viewed online on [https://naturalhazards.](https://naturalhazards.eu/) [eu/.](https://naturalhazards.eu/)

Code Availability All code used to generate the results of this study are available on Zenodo [\(https://doi.org/](https://doi.org/10.5281/zenodo.10911786) [10.5281/zenodo.10911786](https://doi.org/10.5281/zenodo.10911786)). Running the code requires input data also available on Zenodo [\(https://doi.org/](https://doi.org/10.5281/zenodo.10911515) [10.5281/zenodo.10911515](https://doi.org/10.5281/zenodo.10911515)).

Declarations

Confict of interest The authors have no relevant fnancial or non-fnancial interests to disclose.

Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit [http://creativecommons.org/licenses/by/4.0/.](http://creativecommons.org/licenses/by/4.0/)

References

- Bloeschl G, Hall J, Viglione A et al (2019) Changing climate both increases and decreases european river foods. Nature 573(7772):108+. <https://doi.org/10.1038/s41586-019-1495-6>
- Bouwer LM, Jonkman SN (2018) Global mortality from storm surges is decreasing. Environ Res Lett 13(1):014008.<https://doi.org/10.1088/1748-9326/aa98a3>
- Brussee AR, Bricker JD, De Bruijn KM et al (2021) Impact of hydraulic model resolution and loss of life model modifcation on food fatality risk estimation: Case study of the bommelerwaard, the netherlands. J Flood Risk Manag.<https://doi.org/10.1111/jfr3.12713>
- Carisi F, Schröter K, Domeneghetti A et al (2018) Development and assessment of uni- and multivariable food loss models for emilia-romagna (italy). Nat Hazard 18(7):2057–2079. [https://doi.org/10.5194/](https://doi.org/10.5194/nhess-18-2057-2018) [nhess-18-2057-2018](https://doi.org/10.5194/nhess-18-2057-2018)
- Coppedge M, Gerring J, Knutsen CH, et al (2023) V-dem [country-year/country-date] dataset v13. doi: <https://doi.org/10.23696/vdemds23>
- Czado C (2019) Analyzing dependent data with vine copulas: A practical guide with r. Lecture Notes Statist 222:1–242. https://doi.org/10.1007/978-3-030-13785-4_1
- Dißmann J, Brechmann E, Czado C et al (2013) Selecting and estimating regular vine copulae and application to fnancial returns. Computational Statistics & Data Analysis 59:52–69. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.csda.2012.08.010) [csda.2012.08.010](https://doi.org/10.1016/j.csda.2012.08.010)
- Dottori F, Mentaschi L, Bianchi A et al (2023) Cost-efective adaptation strategies to rising river food risk in europe. Nat Clim Chang.<https://doi.org/10.1038/s41558-022-01540-0>
- Environment Agency (2023) Aims spatial food defences (inc. standardised attributes). https://www.data.gov.uk/dataset/cc76738e-fc17-49f9-a216-977c61858dda/ aims-spatial-food-defences-inc-standardised-attributes
- Eurostat, (2020) Statistical regions in the European Union and partner countries - NUTS and statistical regions 2021. Publications Office of the European Union, Luxembourg,. [https://doi.org/10.2785/](https://doi.org/10.2785/850262) [850262](https://doi.org/10.2785/850262)
- Figueiredo R, Schröter K, Weiss-Motz A et al (2018) Multi-model ensembles for assessment of food losses and associated uncertainty. Nat Hazard 18(5):1297–1314. <https://doi.org/10.5194/nhess-18-1297-2018>
- Formetta G, Feyen L (2019) Empirical evidence of declining global vulnerability to climate-related hazards. Glob Environ Chang.<https://doi.org/10.1016/j.gloenvcha.2019.05.004>
- Gerl T, Kreibich H, Franco G et al (2016) A review of food loss models as basis for harmonization and benchmarking. PLoS ONE 11(7):1–22.<https://doi.org/10.1371/journal.pone.0159791>
- Huizinga J, de Moel H, Szewczyk W (2017) Global flood depth-damage functions. Publications Office of the European Union, Luxembourg, Methodology and the database with guidelines. [https://doi.org/10.](https://doi.org/10.2760/16510) [2760/16510](https://doi.org/10.2760/16510)
- Jongman B, Kreibich H, Apel H et al (2012) Comparative food damage model assessment: towards a european approach. Nat Hazard 12(12):3733–3752.<https://doi.org/10.5194/nhess-12-3733-2012>
- Jongman B, Hochrainer-Stigler S, Feyen L et al (2014) Increasing stress on disaster-risk fnance due to large foods. Nat Clim Chang 4(4):264–268.<https://doi.org/10.1038/NCLIMATE2124>
- Jongman B, Winsemius HC, Aerts JCJH et al (2015) Declining vulnerability to river foods and the global benefts of adaptation. Proc Natl Acad Sci 112(18):E2271–E2280. [https://doi.org/10.1073/](https://doi.org/10.1073/pnas.1414439112) [pnas.1414439112](https://doi.org/10.1073/pnas.1414439112)
- Jonkman SN, Vrijling JK, Vrouwenvelder ACWM (2008) Methods for the estimation of loss of life due to foods: a literature review and a proposal for a new method. Nat Hazards 46(3):353–389. [https://](https://doi.org/10.1007/s11069-008-9227-5) doi.org/10.1007/s11069-008-9227-5
- Jäger WS, Morales-Napoles O (2017) A vine-copula model for time series of signifcant wave heights and mean zero-crossing periods in the north sea. ASCE-ASME Journal of Risk and Uncertainty in Engineering Systems, Part A: Civil Engineering 3(4). doi: <https://doi.org/10.1061/ajrua6.0000917>
- Kellermann P, Schröter K, Thieken AH et al (2020) The object-specifc food damage database howas 21. Nat Hazard 20(9):2503–2519.<https://doi.org/10.5194/nhess-20-2503-2020>
- Knoben WJM, Freer JE, Woods RA (2019) Technical note: Inherent benchmark or not? comparing Nash-Sutcliffe and Kling-Gupta efficiency scores. Hydrol Earth Syst Sci 23(10):4323-4331. [https://](https://doi.org/10.5194/hess-23-4323-2019) doi.org/10.5194/hess-23-4323-2019
- Kreibich H, Blauhut V, Aerts JC et al (2019) How to improve attribution of changes in drought and food impacts. Hydrol Sci J 64(1):1–18. <https://doi.org/10.1080/02626667.2018.1558367>
- Kreibich H, Van Loon AF, Schroeter K et al (2022) The challenge of unprecedented foods and droughts in risk management. Nature 608(7921):80+.<https://doi.org/10.1038/s41586-022-04917-5>
- Mares-Nasarre P, van Gent MR, Morales-Nápoles O (2024) A copula-based model to describe the uncertainty of overtopping variables on mound breakwaters. Coast Eng 189:104483. [https://doi.org/10.](https://doi.org/10.1016/j.coastaleng.2024.104483) [1016/j.coastaleng.2024.104483](https://doi.org/10.1016/j.coastaleng.2024.104483)
- Mengel M, Treu S, Lange S et al (2021) Attrici v1.1 - counterfactual climate for impact attribution. Geoscientifc Model Development 14(8):5269–5284. <https://doi.org/10.5194/gmd-14-5269-2021>
- Merz B, Kreibich H, Lall U (2013) Multi-variate food damage assessment: a tree-based data-mining approach. Nat Hazard 13(1):53–64. <https://doi.org/10.5194/nhess-13-53-2013>
- Merz B, Bloesch G, Vorogushyn S et al (2021) Causes, impacts and patterns of disastrous river foods. Nature Reviews Earth & Environment 2(9):592–609. <https://doi.org/10.1038/s43017-021-00195-3>
- Morales-Napoles O, Rajabi-Bahaabadi M, Torres-Alves GA et al (2023) Chimera: An atlas of regular vines on up to 8 nodes. Scientifc Data.<https://doi.org/10.1038/s41597-023-02252-6>
- Muis S, Apecechea MI, Dullaart J et al (2020) A high-resolution global dataset of extreme sea levels, tides, and storm surges, including future projections. Front Mar Sci. [https://doi.org/10.3389/fmars.](https://doi.org/10.3389/fmars.2020.00263) [2020.00263](https://doi.org/10.3389/fmars.2020.00263)
- Nelsen RB (2006) An Introduction to Copulas, 2nd edn. Springer, New York, NY, USA
- Olsen AS, Zhou Q, Linde JJ et al (2015) Comparing methods of calculating expected annual damage in urban pluvial food risk assessments. Water 7(1):255–270.<https://doi.org/10.3390/w7010255>
- Paprotny D, Mengel M (2023) Population, land use and economic exposure estimates for Europe at 100 m resolution from 1870 to 2020. Scientifc Data.<https://doi.org/10.1038/s41597-023-02282-0>
- Paprotny D, Morales Nápoles O, Nikulin G (2016) Extreme sea levels under present and future climate: a pan-European database. E3S Web of Conferences 7:02001. [https://doi.org/10.1051/e3sconf/20160](https://doi.org/10.1051/e3sconf/20160702001) [702001](https://doi.org/10.1051/e3sconf/20160702001)
- Paprotny D, Morales-Nápoles O, Jonkman SN (2017) Efficient pan-european river flood hazard modelling through a combination of statistical and physical models. Nat Hazard 17(7):1267–1283. [https://](https://doi.org/10.5194/nhess-17-1267-2017) doi.org/10.5194/nhess-17-1267-2017
- Paprotny D, Morales-Nápoles O, Jonkman SN (2018) Hanze: a pan-european database of exposure to natural hazards and damaging historical foods since 1870. Earth Syst Sci Data 10(1):565–581. <https://doi.org/10.5194/essd-10-565-2018>
- Paprotny D, Sebastian A, Morales Nápoles O et al (2018) Trends in food losses in Europe over the past 150 years. Nat Commun 9:1985.<https://doi.org/10.1038/s41467-018-04253-1>
- Paprotny D, Morales-Nápoles O, Vousdoukas MI et al (2019) Accuracy of pan-european coastal food mapping. Journal of Flood Risk Management 12(2):e12459.<https://doi.org/10.1111/jfr3.12459>
- Paprotny D, Kreibich H, Morales-Nápoles O et al (2020) Exposure and vulnerability estimation for modelling food losses to commercial assets in europe. Sci Total Environ 737:140011. [https://doi.org/](https://doi.org/10.1016/j.scitotenv.2020.140011) [10.1016/j.scitotenv.2020.140011](https://doi.org/10.1016/j.scitotenv.2020.140011)
- Paprotny D, Kreibich H, Morales-Napoles O et al (2021) A probabilistic approach to estimating residential losses from diferent food types. Nat Hazards 105(3):2569–2601. [https://doi.org/10.1007/](https://doi.org/10.1007/s11069-020-04413-x) [s11069-020-04413-x](https://doi.org/10.1007/s11069-020-04413-x)
- Paprotny D, Terefenko P, Śledziowski J (2024a) HANZE v2.1: an improved database of food impacts in Europe from 1870 to 2020. Earth System Science Data, 16:5145–5170. [https://doi.org/10.5194/](https://doi.org/10.5194/essd-16-5145-2024) [essd-16-5145-2024](https://doi.org/10.5194/essd-16-5145-2024)
- Paprotny D, Rhein B, Vousdoukas MI et al (2024b) Merging modelled and reported food impacts in europe in a combined food event catalogue for 1950–2020. Hydrol Earth Syst Sci 28(17):3983–4010. [https://](https://doi.org/10.5194/hess-28-3983-2024) doi.org/10.5194/hess-28-3983-2024
- Pfeifer D, Kovács EA (2024) Vine copula structure representations using graphs and matrices. Inf Sci 662:120151.<https://doi.org/10.1016/j.ins.2024.120151>
- Pouliasis G, Torres-Alves GA, Morales-Napoles O (2021) Stochastic modeling of hydroclimatic processes using vine copulas. Water.<https://doi.org/10.3390/w13162156>
- Sauer IJ, Reese R, Otto C et al (2021) Climate signals in river food damages emerge under sound regional disaggregation. Nat Commun.<https://doi.org/10.1038/s41467-021-22153-9>
- Scussolini P, Aerts JCJH, Jongman B et al (2016) Flopros: an evolving global database of food protection standards. Nat Hazard 16(5):1049–1061.<https://doi.org/10.5194/nhess-16-1049-2016>
- Scussolini P, Luu LN, Philip S et al (2024) Challenges in the attribution of river food events. WIREs Clim Change 15(3):e874.<https://doi.org/10.1002/wcc.874>
- Steinhausen M, Paprotny D, Dottori F et al (2022) Drivers of future fuvial food risk change for residential buildings in europe. Glob Environ Chang 76:102559.<https://doi.org/10.1016/j.gloenvcha.2022.102559>
- Tanoue M, Hirabayashi Y, Ikeuchi H (2016) Global-scale river food vulnerability in the last 50 years. Sci Rep. <https://doi.org/10.1038/srep36021>
- Tao Y, Wang Y, Wang D et al (2021) A c-vine copula framework to predict daily water temperature in the yangtze river. J Hydrol 598:126430.<https://doi.org/10.1016/j.jhydrol.2021.126430>
- Tarasova L, Lun D, Merz R et al (2023) Shifts in food generation processes exacerbate regional food anomalies in europe. Communications Earth & Environment.<https://doi.org/10.1038/s43247-023-00714-8>
- Tilloy A, Paprotny D, Grimaldi S et al (2024) Hera: a high-resolution pan-european hydrological reanalysis (1950–2020). Earth System Science Data Discussions 2024:1–38. [https://doi.org/10.5194/](https://doi.org/10.5194/essd-2024-41) [essd-2024-41](https://doi.org/10.5194/essd-2024-41)
- Torres-Alves GA, Morales-Napoles O (2020) Reliability analysis of food defenses: The case of the nezahualcoyotl dike in the aztec city of tenochtitlan. Reliability Engineering & System Safety 203:107057. <https://doi.org/10.1016/j.ress.2020.107057>
- United Nations (2022) The 2022 revision of world population prospects. https://population.un.org/wpp/
- United Nations Office for Disaster Risk Reduction (2016) Disaster Risk Reduction Terminology. https:// www.undrr.org/drr-glossary/terminology/
- Vergouwe R (2015) The national food risk analysis for the netherlands. Technical report, Rijkswaterstaat VNK Project Office, https://www.helpdeskwater.nl/publish/pages/131663/vnk-rapport-eng-lr.pdf
- Vousdoukas MI, Mentaschi L, Voukouvalas E et al (2017) Extreme sea levels on the rise along europe's coasts. Earth's Future 5(3):304–323.<https://doi.org/10.1002/2016EF000505>
- Vousdoukas MI, Bouziotas D, Giardino A et al (2018) Understanding epistemic uncertainty in large-scale coastal food risk assessment for present and future climates. Nat Hazard 18(8):2127–2142. [https://doi.](https://doi.org/10.5194/nhess-18-2127-2018) [org/10.5194/nhess-18-2127-2018](https://doi.org/10.5194/nhess-18-2127-2018)
- Wagenaar D, Lüdtke S, Schröter K et al (2018) Regional and temporal transferability of multivariable food damage models. Water Resour Res 54(5):3688–3703. <https://doi.org/10.1029/2017WR022233>
- Ward PJ, Jongman B, Aerts JCJH et al (2017) A global framework for future costs and benefts of river-food protection in urban areas. Nat Clim Chang 7(9):642+.<https://doi.org/10.1038/NCLIMATE3350>
- Pietert't Hart CM, Morales-Nápoles O, Jonkman B (2024) The infuence of spatial variation on the design of foundations of immersed tunnels: Advanced probabilistic analysis. Tunn Undergr Space Technol 147:105624.<https://doi.org/10.1016/j.tust.2024.105624>

Publisher's Note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.